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**The respiratory health hazard of tephra from the 2010 Centennial eruption of Merapi with implications for occupational mining of deposits**

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## UNIVERSITÀ DEGLI STUDI DI TORINO

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1 **The respiratory health hazard of tephra from the 2010 Centennial eruption of Merapi with**  
2 **implications for occupational mining of deposits**

3

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20 **Abstract**

21

22 Ashfall into heavily populated areas during the October-November 2010 eruption of Merapi volcano,  
23 Indonesia created anxiety regarding the growing impacts to health as the eruption escalated and the  
24 hazard zone widened. We made a preliminary assessment of the respiratory hazards to human health  
25 of the tephra deposits (ashfall, lahar, and PDC surge) from the eruption using a laboratory protocol  
26 specifically developed to study the toxic potential of volcanic ash particles. Twenty samples collected  
27 from a range of locations were analysed for health-pertinent mineralogical parameters (grain size,  
28 crystalline silica content, morphology, surface area, bulk chemistry, and leachable elements) and bio-  
29 reactivity (hydroxyl radical generation, haemolytic potential, oxidative capacity, pro-inflammatory  
30 response). The grain size pertinent to respiratory health was variable, ranging from 1.4–15.6 vol. %  
31 sub-4  $\mu\text{m}$  and 3.0–28.9 vol. % sub-10  $\mu\text{m}$  diameter material. No fibre-like particles were observed.  
32 Cristobalite was present in all samples, ranging from 1.9-9.5 wt. %, but surface reactivity and *in vitro*  
33 toxicity assays showed low reactivity for all samples tested. The risk of direct exposure to ash from  
34 fallout was in any case low due to seasonal rains limiting its re-suspension and the immediate and  
35 effective clean-up of communities by local people who supplied the ash to the Indonesian construction  
36 industry for use as aggregate. However, mining of the lahar and thick PDC deposits in the valleys  
37 draining the volcano is performed on a vast, industrial scale which could result in high occupational  
38 exposure to thousands of sand miners at Merapi during the dry seasons. Further study of the health  
39 hazard of the mined Merapi deposits is warranted.

40 **Abbreviations**

41 PDC – Pyroclastic density current

42 BAF – Block-and-ash flow

43 NGO – Non-governmental organisation

44 IVHHN – International Volcanic Health Hazards Network

45 CVGHM - Centre for Volcanology and Geological Hazards Mitigation

## 46 **1. Introduction**

47 Merapi volcano, located in Central Java, Indonesia, is one of the most persistently active volcanoes in  
48 the world, and has a history of deadly eruptions in the last century occurring every 3-5 years. Today,  
49 there are 1.1 million people living on its slopes. Merapi has displayed both explosive and effusive  
50 activity throughout its eruptive history; however, activity over the last 225 years has been dominated  
51 by the viscous extrusion of basaltic-andesite lava domes and subsequent small gravitational, or  
52 explosive, dome collapse (Camus et al., 2000), producing pyroclastic density currents (PDCs) with  
53 associated plumes. Lava dome eruptions have been interspersed with explosive events every 26-58  
54 years (Camus et al., 2000; Thouret et al., 2000). The most common, potentially lethal hazards are  
55 lahars, which result in damage far beyond the more immediate area affected by dome collapse  
56 (Thouret et al., 2000).

57

58 The October-November 2010 explosive eruption of Merapi was its largest since 1872, when there  
59 were reports of up to 30 cm of ashfall, and caused its worst disaster since 1930, when 36 villages  
60 were destroyed and 1369 people killed in PDCs. Major explosions on 26 October, 30-31 October, and  
61 4-5 November dispersed large amounts of ash to the west and south, affecting major urban areas of  
62 Magelang and Yogyakarta. The total erupted tephra volume was estimated to be  $50-100 \times 10^6 \text{ m}^3$ ,  
63 approximately 10 times greater than deposits from typical eruptions of the past few decades (Lavigne  
64 et al., 2011; Surono et al., this volume). Varying contributions of juvenile dome and older 2006 dome  
65 material were incorporated into the discrete eruptive events (Surono et al., this volume). PDCs and  
66 associated detached surges extended beyond the previously delineated hazard zones (Thouret et al.,  
67 2000), resulting in the destruction of some villages and 367 official fatalities (Gertisser, 2011; Surono  
68 et al., this volume). Nearly a third of a million people were displaced due to the risk of PDC impact to  
69 their villages, and many of these individuals would be exposed to re-suspended ash throughout the  
70 clean-up and rebuilding phase.

71

72 Quarrying of the extensive lahar and PDC deposits for construction began soon after the 2010  
73 eruption and continued on an industrial scale involving thousands of people from local villages and  
74 further afield. In response to the immediate and future hazard posed by the erupted material, we  
75 embarked on a laboratory assessment of the respiratory health hazard of Merapi ash found in

76 different tephra deposits (ash fall, a detached surge, and a lahar) following a protocol previously  
77 developed at Rabaul, Papua New Guinea, Chaitén, Chile and Eyjafjallajökull, Iceland (Horwell et al.,  
78 In Prep-b; Horwell et al., 2010a; Le Blond et al., 2010) (see Figure 1). Health-pertinent mineralogical  
79 data of grain size, bulk composition, particle morphology, and crystalline silica content are combined  
80 with toxicological assays (surface hydroxyl radical generation, particle oxidative capacity, haemolysis,  
81 pro-inflammatory response) to inform the particle hazard.

82

### 83 **1.1 Potential diseases related to the inhalation of volcanic ash**

84 Horwell & Baxter (2006) give a comprehensive review of respiratory diseases associated with inhaling  
85 volcanic ash particles <10 µm diameter. Following immediate exposure, susceptible individuals may  
86 develop asthma and bronchitis (Baxter et al., 1983); heavy and prolonged exposure to ash containing  
87 a high concentration of crystalline silica in the 'respirable' fraction (<4 µm diameter and able to  
88 penetrate to the lung alveoli) may lead to the fibrotic lung disease silicosis and possibly lung cancer  
89 (International Agency for Research on Cancer, 1997). In addition, pulmonary tuberculosis is common  
90 in Indonesia and its incidence and severity can be promoted by heavy exposure to dusts containing  
91 crystalline silica (Hnizdo and Murray, 1998). It is well understood that substantial quantities of  
92 crystalline silica can crystallise in volcanic domes as cristobalite through vapour-phase deposition and  
93 devitrification of groundmass, e.g. up to 20 wt. % in ash from both the Soufrière Hills volcano on  
94 Montserrat, West Indies (Baxter et al., 1999; Horwell et al., In Prep-a) and Chaitén, Chile (Horwell et  
95 al., 2010a). Eruptions at dome forming volcanoes can generate extremely fine-grained ash (Horwell,  
96 2007), and can further contain cristobalite through incorporation of altered edifice material in an  
97 explosive event (Baxter et al., 1999; Horwell et al., 2010a).

98

### 99 **1.2 Evidence of crystalline silica generation at Merapi**

100 Cristobalite has previously been identified in volcanic ash near Merapi, with a particular emphasis on  
101 its presence in volcanic soils (Hardjosoestastro, 1956). However, no studies have quantified the  
102 concentration of cristobalite or addressed it as a respiratory hazard. Archived samples were therefore  
103 informative at the onset of the 2010 eruption. Analysis of an ash sample from a 1998 dome collapse  
104 event (named MER\_arc in this study, see Table 1) raised concerns about the amount of respirable  
105 material produced by Merapi, with this sample containing 13 vol. % sub-4 µm particles (Horwell, 2007)

106 but only 3.8 wt. % cristobalite (Table1). Before this study, we also analysed the cristobalite content of  
107 dome rock samples from 1996 and 1998 block-and-ash flows (BAFs) and confirmed the presence of  
108 cristobalite (3-5 wt. %). Cristobalite content in dome rock is expected to be substantially lower than  
109 that in co-PDC plume deposits due to enrichment of cristobalite in the plume by fractionation (Horwell  
110 et al., 2001).

111

### 112 **1.3 Initial response to the eruption**

113 When international news agencies began reporting health-related problems attributed to exposure to  
114 volcanic ash (27 October) it became clear that rapid dissemination of information on preparedness for  
115 ashfall and the health hazards of ash was warranted. The International Volcanic Health Hazard  
116 Network ([www.ivhhn.org](http://www.ivhhn.org)) is a recognized source of such advice, and IVHHN fact sheets were rapidly  
117 translated for us into Bahasa Indonesia by a native speaker. High-resolution copies were sent for  
118 local printing by NGOs, such as Save the Children, for widespread distribution at the evacuation  
119 camps. We sought ash samples for analysis from these contacts and others at Universitas Gadjah  
120 Mada, Jogjakarta, with the first batch of samples arriving at Durham University on 18 November 2010,  
121 weeks before Indonesia's Centre for Volcanology and Geological Hazards Mitigation (CVGHM)  
122 reduced the activity warning from its highest level, level 4. Local scientists and NGOs collected fresh  
123 ash fall samples at 10 discrete locations, but it was not possible to co-ordinate a systematic sampling  
124 strategy in the emergency period. We undertook a field mission to collect further samples and to  
125 monitor air quality in populated areas.



126 **2. Materials and methods**

127

128 **2.1 Sample collection and selection**

129 Samples from three different types of tephra deposits were collected, namely ash fall, a detached  
130 surge and a lahar, to provide insight into the hazard posed by variably-sourced ash at Merapi. Eleven  
131 ash fall samples were collected by local scientists and volunteers between 30 October and 13  
132 November 2010. A further 9 samples were collected during our field mission from 29 November to 11  
133 December 2010 from *in situ* village and field deposits. The village investigated (Bronggang) had been  
134 impacted by two detached PDC surges and samples were taken both outside and within affected  
135 houses. Although efforts were made to collect samples as fresh as possible, many samples were  
136 rained on prior to collection. The eruption took place during the wet season and normal rainfall may  
137 have been augmented by the 2010 La Niña event. During this time heavy rain fell on the region for  
138 more than 2 hours per day. A complete listing of samples is presented in Table 1, and sample  
139 locations are displayed in Figure 2. Nine samples collected after the 5 November eruption are likely  
140 composite samples from the different major eruptive events as the plume pattern for the 5 November  
141 eruption overlaid portions of the deposits from both the 25 October and 30 October eruptions.  
142 Samples MER\_10\_01, MER\_10\_03 and MER\_10\_04 were collected prior to being rained upon.  
143 Samples MER\_10\_13 through \_15 are detached surge samples, where MER\_10\_13 and \_14 were  
144 collected from elevated surfaces (1 m) and correspond with floor-collected samples MER\_10\_12 and  
145 \_15, respectively.

146

147 Mineralogical analyses were performed on samples in accordance with the described protocol (Figure  
148 1). A set of four samples (MER\_10\_02, MER\_10\_03, MER\_10\_04, and MER\_10\_12) was selected as  
149 a representative cross section for additional surface reactivity and toxicology analyses based on  
150 eruption date, exposure to the environment following deposition, and potential hazard defined by grain  
151 size and crystalline silica data. Imaging was performed on the four sub-samples as well as  
152 MER\_10\_01, a freshly collected ash sample, and MER\_10\_13, the fine surge component associated  
153 with MER\_10\_12. An archived sample (MER\_arc), collected pristine 200m from a PDC on the slopes  
154 of Merapi during a dome collapse eruption in 1998 (Horwell, 2007; Horwell et al., 2007), was included  
155 for comparison.

156

157 All samples were oven dried in glass dishes at 90 °C for 24 hours and then sieved through 2 and 1  
158 mm mesh sieves to remove any material that is too coarse for laser diffraction grain size analysis.  
159 Unless otherwise stated, all samples used for analysis are the sub-1 mm fraction in order to keep  
160 results directly comparable to previous studies.

161

## 162 **2.2 Mineralogical and compositional analyses**

163 All of the methods in the health-hazard assessment protocol have been previously described (Le  
164 Blond et al., 2010) with the exception of inflammatory potential and are only discussed briefly here.

165

166 Grain-size distribution of volcanic ash is important as particles with <10 µm aerodynamic diameter are  
167 able to penetrate into the human lung (Quality of Urban Air Review Group, 1996). Measurements  
168 were carried out using a Malvern Mastersizer 2000 laser diffractometer with a Hydro MU attachment  
169 at the Department of Geography, University of Cambridge, UK after Horwell (2007). This technique  
170 measures particles within the 0.2-2000 µm range and data reported are the average of three  
171 measurement cycles (Table 2). Samples were measured with a refractive index of 1.63 and an  
172 absorption coefficient of 0.1 (Horwell, 2007). Results were rescaled to incorporate the 1-2 mm fraction  
173 of the ash using the fraction weights measured after sieving. All samples contained <<1.0 wt. % 1-2  
174 mm fraction except for samples MER\_10\_07 (7.21 wt. %), MER\_10\_11 (4.09 wt. %), MER\_10\_16  
175 (11.27 wt. %) and MER\_10\_20 (5.57 wt. %). Only samples MER\_10\_11 and MER\_10\_16 contained  
176 substantial >2 mm material at 6.17 wt. % and 7.33 wt. %, respectively; however, this material is  
177 outside of the ash fraction and is not further discussed in this study.

178

179 The major elemental oxide composition of samples MER\_10\_01 through \_12 was determined using  
180 X-ray fluorescence (PANalytical Axois Advanced X-ray fluorescence (XRF) spectrometer at the  
181 Department of Geology, University of Leicester, UK).

182

183 Scanning electron microscopy (SEM) was used to investigate particle morphology. Particles mounted  
184 on polycarbonate discs adhered to Al stubs and coated with 30 nm of gold/palladium were imaged  
185 using the Hitachi SU-70 FEG SEM at the Department of Physics, Durham University, UK.

186

187 The total quantity of crystalline silica as quartz, cristobalite, and tridymite was determined by X-ray  
188 diffraction (XRD) at the Natural History Museum, London following the Internal Attenuation Standard  
189 (IAS) method of Le Blond et al. (2009) using an Enraf-Nonius X-ray diffractometer with an INEL  
190 curved position sensitive detector (PSD). This method allows for the rapid quantification of a single  
191 mineral phase (e.g., a crystalline silica polymorph) in a mixed dust without prior knowledge of the  
192 mineralogical composition of the bulk sample. The method has a < 3 wt. % error.

193

194 Leachate analyses were carried out on a sample split that was not oven-dried as it is possible that  
195 drying can alter surface volatile species. The concentrations of readily-soluble elements in the ash  
196 were determined by gently shaking 0.25 g of ash with 25 ml of deionized water for 1 hour. The pH of  
197 the ash leachate was measured prior to filtration on 0.2 µm membrane filter. Anions (F<sup>-</sup>, Cl<sup>-</sup>, SO<sub>4</sub><sup>2-</sup>) and  
198 major elements (Si, Al, Fe, Mg, Ca, Na, K) were measured in the filtrated ash leachate by ion  
199 chromatography (IC) and inductively-coupled plasma-optical emission spectroscopy (ICP-OES),  
200 respectively. Trace metals (As, Cd, Co, Cr, Cu, Ni, Pb and Zn) were determined by inductively  
201 coupled plasma-mass spectroscopy (ICP-MS).

202

### 203 **2.3 Surface reactivity and *in vitro* toxicity testing**

204 The specific surface area (SSA) of a particle is an indicator of the total available surface for reactions  
205 to occur in the lung. Surface area for the selected subset of 4 samples was measured using the  
206 Brunauer-Emmett-Teller (BET) method of nitrogen adsorption on a Micromeritics TriStar 3000 Surface  
207 Area and Porosimetry Analyser in the Department of Chemistry, Durham University, UK. Prior to  
208 analysis samples were degassed under N<sub>2</sub> at 150° C for at least 2 hours.

209

210 Particle surface reactivity can be estimated by the ability of a particle to generate reactive oxygen  
211 species (Fubini et al., 1995). Horwell et al. (2007; 2003a) suggest that all volcanic ash is capable of  
212 generating the hydroxyl radical via the iron-catalysed Fenton reaction, with more mafic samples being  
213 able to elicit a more pronounced response. Electron Paramagnetic Resonance (EPR) spectroscopy in  
214 association with the 'spin trap' technique was used to quantify the generation of hydroxyl radicals by  
215 the samples in solution through replication of Fenton chemistry (Fubini et al., 1995; Horwell et al.,

216 2007; Horwell et al., 2003a). Measurements were carried out on a Miniscope 100 ESR spectrometer,  
217 Magnettech, at the Dipartimento di Chimica, Università degli Studi di Torino, Italy following a standard  
218 methodology (Horwell et al., 2007). Measurements are collected at 10, 30, and 60 minutes, and data  
219 are averages of three individual experiments expressed per unit surface area.

220

221 The amount of removable ferrous ( $\text{Fe}^{2+}$ ) and ferric ( $\text{Fe}^{3+}$ ) iron on the sample surfaces was determined  
222 as this represents the iron available for participation in the Fenton reaction. Removable iron was  
223 measured spectrophotometrically at 562 nm using the chelator Ferrozine to remove ferrous iron from  
224 the particle surface (Horwell et al., 2007; Horwell et al., 2003a). Total iron mobilised was measured by  
225 first reducing  $\text{Fe}^{3+}$  to  $\text{Fe}^{2+}$  with ascorbic acid. Samples were analysed in a Uvikon spectrophotometer  
226 also at the Dipartimento di Chimica, Università degli Studi di Torino. Experiments were carried out for  
227 9 days with measurements taken every 24 hours (excluding the weekend). In addition to the four  
228 Merapi samples chosen, four ash samples from other volcanoes, analysed previously by Horwell et al.  
229 (2007), were re-analysed here for comparative purposes. Data are presented alongside results from  
230 Horwell et al. (2007) for Min-U-Sil quartz standard and MER\_arc.

231

232 The capacity of ash to cause cell membrane damage was determined by lysis of human erythrocytes  
233 (red blood cells) sourced from fresh venous blood. Ash samples were diluted in saline to final  
234 concentrations between 0.031 and 1.0  $\text{mg ml}^{-1}$  and sonicated for 10 minutes. Particle suspensions  
235 were incubated with isolated erythrocytes for 30 minutes at room temperature, shaking gently.  
236 Following incubation, plates were centrifuged at 400 x g for 5 minutes and the amount of released  
237 haemoglobin was determined by absorbance at  $\lambda=550$  nm. The quartz standard DQ12 was used as a  
238 positive control and rutile  $\text{TiO}_2$  as a negative control. Percentage haemolysis compared to complete  
239 haemolysis (0.1 % Triton X-100) was calculated and results are the average of three experiments,  
240 each performed in triplicate.

241

242 The potential of ash to induce oxidative stress by the depletion of antioxidants in artificial preparations  
243 of human respiratory tract lining fluid (RTLFL), a measure of bio-reactivity in the airways thought to be  
244 associated with asthma (Ayres et al., 2008), was determined at Kings College London. The RTLFL  
245 exposure assay comprises a composite solution containing a final concentration of 200  $\mu\text{M}$  ascorbic

246 acid (AA), urate (UA) and reduced glutathione (GSH), which was exposed to 50  $\mu\text{g ml}^{-1}$  ash or  
247 particulate matter controls (in-house negative (M120) and positive (NIST1648a and roadside PM1-3))  
248 for 4 hours at 37 °C. The post-exposure concentrations of AA, UA and GSH were indicators of the  
249 oxidative potential of the ash samples.

250

251 The pro-inflammatory potential of ash was assessed against human lung epithelial type 1-like (TT1)  
252 cells (van den Bogaard et al., 2009) to establish their acute and chronic cellular reactivity. Analyses  
253 were performed at the National Heart and Lung Institute, Imperial College London. TT1 cells were  
254 grown to confluence in DCCM-1 cell culture media with 10 % New Born Calf serum (Invitrogen,  
255 Paisley, UK), 1% penicillin/streptomycin/glutamine (Invitrogen) and 0.5  $\text{mg ml}^{-1}$  gentacin (G148;  
256 Sigma Aldrich). Cells were serum starved for 24 hours prior to sample treatment, and cell monolayers  
257 were exposed to tephra samples dispersed in serum-free DCCM-1. Mediator conditioned cell culture  
258 medium including the sample doses was removed after 24 hours and centrifuged. Exposed TT1 cell  
259 cultures were washed x3 with PBS and fresh cell culture medium was added. Conditioned media  
260 removal was repeated at 24 and 72 hours post exposure. Acute (24 hour exposure) and chronic (24  
261 and 72 hours post exposure) inflammatory potential was determined by measuring the release of  
262 interleukin 6 (IL-6), interleukin 8 (IL-8) and monocyte chemotactic protein (MCP-1) from exposed TT1  
263 cells using an enzyme-linked immunosorbent assay (ELISA). A dose range of 0.5 to 50.0  $\mu\text{g ml}^{-1}$  of  
264 tephra sample was used.

265

#### 266 **2.4 Ambient air particle concentrations and personal exposure**

267 Ambient air particle concentrations were collected during the field mission using a TSI DustTrak  
268 Aerosol Monitor. The instrument was run for 30 minutes each morning and evening in Yogyakarta for  
269 9 days as well as at a selection of field sites for short-term (20 minute) ambient readings.  
270 Measurements in Yogyakarta were taken at 3 locations: by a main street with heavy traffic, near a  
271 school away from major traffic but with frequent foot traffic, and in a quiet neighbourhood with little  
272 traffic (1 car every five minutes). Concentrations were also measured in urban air at Magelang,  
273 Muntilan, and Klaten. Rural locations monitored were near Selo, Dukun, Argo Mulyo, and Prambanan  
274 temple.

275

276 We monitored the personal exposure for 2 workers while they were extracting river bed deposits and  
277 wearing a TSI Sidepak Personal Aerosol Monitor. Two TSI DustTraks were set up concurrently with  
278 the personal measurements; one near the working area to collect proximal occupational levels  
279 resulting from re-suspension of the deposit, and the other at a distance upwind to collect  
280 environmental background levels. All instruments were calibrated prior to use to ensure that directly  
281 comparative readings were obtained. We were unable to monitor exposure for the main mining  
282 activities on the PDCs because large-scale excavation of material had not yet begun during the field  
283 mission.

## 284 **3. Results**

285

### 286 **3.1 Particle size**

287 Grain size data are presented as a cumulative volume percent according to health pertinent size  
288 fractions (Horwell, 2007) (Table 2). Overall, grain size distributions for the 14 ashfall samples varied in  
289 the amount of respirable ash (<4 µm diameter) from 1.4–15.6 vol. %. Aside from a sample from the 26  
290 October eruption which had 15.6 vol. % sub-4 µm material, the rest of the samples' ranges are typical  
291 for respirable material generated during Vulcanian to Plinian (VEI 3-4) explosive eruptions (<~10 vol.  
292 % sub-4 µm) (Horwell, 2007; Horwell et al., 2010a; Horwell et al., 2010b; Le Blond et al., 2010).  
293 Inferring differences in grain size between the major explosive events of the 2010 eruption is difficult,  
294 however, due to the effects of sample location distance from the crater, plume dispersal axis and  
295 reworking of material by rain.

296

297 Detached surge deposits collected from within 2 homes in Bronggang village were, in general, much  
298 finer grained than the ash fall and parent surge deposits, with approximately 13 vol. % sub-4 µm  
299 material. The component of the detached surges which comprised the samples collected from  
300 elevated surfaces within the homes was especially enriched in fines, with 19 vol. % sub-4 µm material  
301 (2 samples). This differed greatly from the sample collected from the exterior of the house which only  
302 contained 6.5 vol. % sub-4 µm material, although sheltered and appeared pristine when collected,  
303 suggesting the coarsest material settles in the parent surge outside of the houses (Jean-Christophe  
304 Komorowski, personal communication).

305

306 The sample from the lahar deposit was collected 22 km from the crater and contained no ash particles  
307 <10 µm. The lahar sample was not investigated further as there was no respirable component.

308

### 309 **3.2 Particle morphology**

310 The general morphology of the health-pertinent fractions (sub-10 µm) of Merapi ash was poorly  
311 vesiculated, sub-angular and blocky with varying amounts of sub-micron particles adhering to the  
312 surfaces of larger particles (Figure 3), as found in samples from other volcanoes (Hillman et al., 2012;  
313 Horwell and Baxter, 2006; Horwell et al., 2010a; Horwell et al., 2010b; Le Blond et al., 2010). No

314 morphological differences were observed between the PDC samples and the fall samples. No  
315 respirable mineral fibres were observed in any of the samples, eliminating the asbestiform concern  
316 raised from health studies of other eruptions (discussed further in Le Blond et al. (2009) and Reich et  
317 al. (2009)).

318

### 319 **3.3 Bulk composition**

320 Ash samples ranged from trachy-basalt to trachy-andesite (51.90-57.42 wt. % SiO<sub>2</sub>, 1.97-2.42 wt. %  
321 K<sub>2</sub>O) and are reported on a total alkali versus silica plot (Figure 4). Bulk oxide elemental data for all  
322 samples are listed in Table 3. The majority of erupted products at Merapi are calc-alkaline, high-K  
323 basaltic-andesites that range from 52-57 wt. % SiO<sub>2</sub> and 1.80-2.84 wt. % K<sub>2</sub>O (Camus et al., 2000;  
324 Gertisser and Keller, 2003); the array observed in this study encompasses the full range previously  
325 reported, and corresponds with bulk rock data of juvenile blocks from the 5 November pyroclastic  
326 flows (Surono et al., this volume). While the bulk composition from the 26 and 30 October eruptions  
327 reflects the narrowly constrained whole rock data from the 2006-2008 eruptive period (~55-56 wt. %  
328 SiO<sub>2</sub> versus ~55-57 wt. % SiO<sub>2</sub> for the current data), ash samples collected from the 5 November  
329 eruption appear to be more primitive, with 52-55 wt. % SiO<sub>2</sub>. However, there is no evidence of  
330 chemical or petrologic changes throughout the eruption in bulk rock samples collected by other  
331 scientists (John Pallister, personal communication). Further, deconvolving the effects of  
332 environmental exposure from the eruptive phase is difficult as all but one sample from 5 November  
333 were composite samples or had been rained on. The apparent discrepancy could be the result of  
334 winnowing from wind and rain, especially since the major element composition of the dry 5 November  
335 sample (i.e., MER\_10\_03) overlaps with the 5 November PDC results of Surono et al (this volume). A  
336 rough correlation was seen between grain size (sub-4 µm) and bulk SiO<sub>2</sub> ( $R^2=0.80$ , data not shown)  
337 for wet/composite samples, where finer samples correspond with higher bulk SiO<sub>2</sub> values. This could  
338 indicate removal of a fine groundmass fraction from the deposits, which is typically high in silica (e.g.,  
339 ~70 wt. % SiO<sub>2</sub> for groundmass glass for Merapi 2010). The removal of a fine, vesicular glassy  
340 component is consistent with the observation that fall-deposits are dominantly composed of angular,  
341 lithic fragments (John Pallister, personal communication). The lower SiO<sub>2</sub> observed could also be the  
342 result of physical fractionation in the plume (Horwell et al., 2001), however no correlation was  
343 observed between bulk SiO<sub>2</sub> and collection distance.



344

### 345 **3.4 Crystalline silica content**

346 Cristobalite was detected in all samples (1.9-9.5 wt. %) and minor amounts of quartz were  
347 quantifiable in samples MER\_10\_06 (1.25 wt. %), \_16 (0.75 wt %), and \_18 (0.66 wt. %). No tridymite  
348 was identified in any sample. Abundances of cristobalite reported in Table 2 are for sieved bulk  
349 samples (<1 mm) rather than the respirable fractions as it is not feasible to separate sufficient  
350 respirable material with such small quantities of ash (often <20 g) and within the timeframe of an  
351 urgent study. The finer, PDC surge samples collected from raised surfaces in houses (MER\_10\_13  
352 and \_14) were depleted in crystalline silica compared to their ground-collected counterparts  
353 (MER\_10\_12 and \_15) (3.2 and 4.5 wt. % versus 7.9 and 8.8 wt. % respectively).

354

### 355 **3.5 Leachate analysis**

356 The Merapi ash leachates for the subset of 4 samples displayed slightly acidic pH values (5.1-6.3)  
357 (table 4). There is no observable difference in pH between wet and dry samples. Calcium was the  
358 dominant dissolved cation in all samples. Sulphate was present in higher concentration than  $\text{Cl}^-$  and  $\text{F}^-$   
359 . The molal S/Cl ratio in MER\_10\_02 and MER\_10\_12 leachates is approximately four times higher  
360 than in MER\_10\_03 and MER\_10\_04. Copper and Zn were the most abundant trace metals mobilized  
361 upon exposure to water. Nickel and Cr were also measured in significant concentrations (0.2 – 1 mg  
362  $\text{kg}^{-1}$ ) in the leachates, while concentrations of Fe, Cd, and Pb were about ten times lower. Compared  
363 to other samples, Zn was ten times greater in MER\_10\_03. This sample is also enriched in Cr and Ni.  
364 There were no discernible differences between wet and dry samples, however only one wet sample  
365 was analysed (MER\_10\_02).

366

### 367 **3.6 Particle specific surface area**

368 The specific surface area for the selected ash samples ranged from 0.51–1.03  $\text{m}^2 \text{g}^{-1}$  (Table 2), which  
369 is at the low end of the range previously observed for explosive volcanic ash (0.2-6.9  $\text{m}^2 \text{g}^{-1}$ ) (Horwell  
370 et al., 2007; Horwell et al., 2010b). The PDC surge samples MER\_10\_12 and MER\_10\_02 had the  
371 greatest surface area per mass and also contained the greatest proportion of respirable material.

372

### 373 **3.7 Hydroxyl radical generation and iron release**

374 Surface reactivity determined by hydroxyl radical generation for samples MER\_10\_02, \_03, \_04 and  
375 \_12 was low with reference to comparative ash samples analysed concurrently (Figure 5). All four  
376 samples generated fewer than  $0.1 \mu\text{mol m}^{-2}$  at 30 minutes with little change in kinetics through each  
377 individual experiment. The data follow the expected trend observed in previous studies, where mafic  
378 samples generate more hydroxyl radicals than more silicic samples (Horwell et al., 2007; Horwell et  
379 al., 2003a). Considering results from tephra fall samples only, the most mafic sample (MER\_10\_03)  
380 generated the greatest number of radicals, whereas the most silicic sample (MER\_10\_02) generated  
381 the fewest radicals. The low overall Fe-catalysed surface reactivity for all samples is unlikely  
382 explained by exposure to rain (e.g. Fubini et al. (1995), Fubini (1998), Le Blond et al. (2010)) as  
383 MER\_10\_03 was collected fresh and generated a similar number of radicals to samples exposed for 5  
384 days ( $0.094 \mu\text{mol m}^{-2}$ ). The 1998 dome collapse ash sample MER\_arc analysed by Horwell et al  
385 (2007) gave similar results to the present study. All samples produced more radicals than the Min-U-  
386 Sil quartz standard, and data are consistent with all previously analysed volcanic ash samples  
387 (Hillman et al., 2012; Horwell et al., 2007; Horwell et al., 2003a; Horwell et al., 2010b).

388

389 The amount of total iron released for the 4 Merapi samples was on the order of that released by the  
390 sample comparisons from Montserrat (andesitic) and Pinatubo (dacitic), and far lower than that  
391 released by the Etna and Cerro Negro samples (both basaltic) (Figure 5). No trend was observed for  
392 preferential mobilisation of either  $\text{Fe}^{2+}$  or  $\text{Fe}^{3+}$  for any of the samples, with MER\_10\_02 releasing  
393 more  $\text{Fe}^{2+}$ , MER\_10\_03 releasing more  $\text{Fe}^{3+}$ , and MER\_10\_04 and \_12 releasing equivalent  
394 quantities of both (data not shown for brevity).

395

396 A comparison of hydroxyl radical generation and total iron released (Figure 5) shows very little  
397 correlation. As has been seen with other data, there seems to be a lack of trend between iron in either  
398 oxidation state (data not shown) or total iron and hydroxyl radical generation when iron release is low  
399 (Hillman et al., 2012; Horwell et al., 2007). Production of hydroxyl radicals via Fenton chemistry is a  
400 catalytic reaction related to the coordination of iron atoms, and, as such, even trace amounts of  $\text{Fe}^{2+}$   
401 may be sufficient to trigger hydroxyl radical generation.

402

403 **3.8 *In vitro* toxicology**

404 None of the samples showed haemolytic activity at the doses used when compared to a positive  
405 quartz control. This is in accord with previous reports for the haemolytic potential of other explosive  
406 ash samples (Horwell et al., In Prep-b; Le Blond et al., 2010). No significant difference was observed  
407 between tephra fall samples and the PDC surge sample analysed.

408

409 The ash samples did not exhibit an oxidative potential within the RTLf exposure model. The in-house  
410 negative and positive controls performed as expected. Again, this is in line with work carried out on  
411 volcanic ash from other volcanoes (Horwell et al., In Prep-b; Le Blond et al., 2010).

412

413 Merapi ash was found to be largely non-reactive against the TT1 cell model, failing to incite an acute  
414 inflammatory response or impart a lasting (chronic) inflammatory effect following exposure. Three  
415 tephra fall samples used in the experiment (MER\_10\_02, \_03, and \_04) did not produce a significant  
416 increase in cytokine or chemokine release over the unexposed controls. The PDC surge deposit  
417 sample (MER\_10\_12) did produce a significant 1.4-fold increase in MCP-1 at the 24 hour acute  
418 exposure, and a significant 1.2-fold increase in IL-8 at 24 hours post acute exposure, however this  
419 response was only associated with the highest dose ( $50 \mu\text{g ml}^{-1}$ ) and resolved by Day 5.

420

### 421 **3.9 Ambient air monitoring**

422 Particle concentrations measured in the ambient air at Magelang and Yogyakarta were low and less  
423 than anticipated from the resuspension of ash by traffic due to the rapid removal and transport of ash  
424 by locals to sell as aggregate, in addition to seasonal rains washing the ash away. In Magelang (26  
425 km WNW), where up to 5 cm of ash was reported during the eruption, nearly all of the ash had been  
426 removed or incorporated into the environment prior to the field mission.

427

428 No background information on daily concentrations of particulate matter  $<10 \mu\text{m}$  ( $\text{PM}_{10}$ ) levels was  
429 available for comparison with the data collected during our field mission. The measurements we took  
430 may not be representative of all districts surrounding Merapi, but we had no volcano-specific  
431 respiratory health concerns as ambient  $\text{PM}_{10}$  concentrations outside of urban areas were at or below  
432  $0.05 \text{ mg m}^{-3}$ . Twice daily measurements of ambient concentrations in Yogyakarta and three discrete  
433 measurements in Magelang did not rise above  $0.10 \text{ mg m}^{-3}$  ( $0.07\text{--}0.10 \text{ mg m}^{-3}$  in the morning and

434 0.04-0.05 mg m<sup>-3</sup> in the afternoon) except for a couple of weekdays (7-9 Dec) in Yogyakarta when  
435 construction work was taking place nearby (approximately 1 km away), and concentrations reached  
436 upwards of 0.80 mg m<sup>-3</sup>. Air quality limits are further discussed in section 4.1.

437

438 Personal exposure levels for two workers manually shovelling wet ash that had been dredged out of a  
439 river using hand tools were 0.055 mg m<sup>-3</sup> and 0.029 mg m<sup>-3</sup> (PM<sub>10</sub>, 1 hour average). Maximum  
440 concentrations as one minute averages within those readings were 0.224 mg m<sup>-3</sup> and 0.175 mg m<sup>-3</sup>,  
441 respectively. Ambient concentrations in the working area were 0.042 mg m<sup>-3</sup> for a continuous 2 hour  
442 reading (PM<sub>10</sub>), and background monitoring levels at a distance of 100 m (no wind) were 0.022 mg m<sup>-3</sup>.  
443 <sup>3</sup>. These results were not representative of exposure conditions for the mining of lahar or PDC  
444 deposits in the dry season, but were typical for working with wet ash.

445 **4. Discussion**

446

447 **4.1 Tephra fall and corresponding particle hazard**

448 Work by Surono et al. (this volume) suggests that the 26 October eruption had a phreatomagmatic  
449 component. The directed explosion also destroyed part of the existing dome from 2006 (as evidenced  
450 by the non-juvenile lithic material present in the associated PDC deposit; Surono et al., this volume).  
451 The combination of a phreatomagmatic explosion and disruption of an existing dome likely resulted in  
452 the unusually fine grain size distribution of samples from 26 October despite the eruption being  
453 regarded as less energetic than the 5 November event. Incorporation of dome material has been  
454 shown to produce finer ash than plinian or vulcanian explosive magmatic eruptions where no dome is  
455 present, with dome collapse eruptions expected to generate ash with 10-18 vol. % sub-4  $\mu\text{m}$  (Horwell,  
456 2007; Horwell et al., 2010a). Correspondingly, the finest grained samples are from the first eruption  
457 on 26 October as well as later composite samples which include ash from the 26 October eruption,  
458 i.e. samples MER\_10\_17 and \_18. Sample MER\_10\_02 from 26 October is exceptionally fine grained  
459 (28.9 vol. % sub-10  $\mu\text{m}$ , 15.6 % sub-4  $\mu\text{m}$ ); compared to our data base of results for ash from other  
460 volcanoes, it is the finest of 250 samples after two samples from the final phreatomagmatic stages of  
461 plinian and sub-plinian eruptions at Mt Vesuvius (Horwell, 2007; Horwell et al., 2010b) and a dome-  
462 collapse sample from Chaitén volcano in 2008 (Horwell et al., 2010a).

463

464 The 5 November eruption also destroyed a dome which had been growing between 31 October and 4  
465 November, yet evidence of this is not distinguishable in the grain size or cristobalite data while  
466 accounting for distance of sample collection. This absence of observable indicators of dome  
467 incorporation could have been due to integration of tephra from other eruptive events or by the effects  
468 of the seasonal climate as most samples associated with that eruption had been exposed to the  
469 environment prior to collection.

470

471 Previous analysis of a 1998 archive sample before the eruption alerted us to the potential respiratory  
472 hazard of the exceptionally fine-grained ash in Merapi's dome eruptions. However, the rapidity with  
473 which cities and the surrounding environment recovered from the extensive amount of tephra fall was  
474 unforeseen and indeed unprecedented in our experience. This was largely due to the rapid and

475 effective removal of deposited material by local people to sell as aggregate for construction and the  
476 daily downpours of rain, which mobilised massive amounts of material from roads and around houses  
477 and limited re-suspension shortly after each eruptive event. There was therefore little risk of long-term  
478 exposure to the ash for the general population.

479

480 Health guidelines on airborne particulate levels for volcanic ash were developed for Montserrat based  
481 on the UK standard for ambient PM<sub>10</sub> (Expert Panel on Air Quality Standards, 1995; Expert Panel on  
482 Air Quality Standards, 2001) and the US occupational exposure limit for crystalline silica (National  
483 Institute of Occupational Health and Safety (NIOSH), 2002). Searl et al. (2002) adjusted for 24 hour  
484 exposure in the general population of Montserrat as a guide to limiting exposure, recommending no  
485 action at concentrations below 0.05 mg m<sup>-3</sup> and raised alertness for individuals who have experienced  
486 adverse health effects during past dusty episodes for concentrations between 0.05-0.10 mg m<sup>-3</sup> (1  
487 hour averages for PM<sub>10</sub>). The guidelines developed for Montserrat were based on a cristobalite  
488 concentration of up to 20 wt. % in the ash and are therefore not directly applicable; however, rural  
489 PM<sub>10</sub> concentrations near Merapi were never measured above 0.05 mg m<sup>-3</sup> (30 minute readings) at  
490 the locations studied, reducing concerns of elevated volcanically-associated PM. Urban PM<sub>10</sub>  
491 concentrations were between 0.04-0.10 mg m<sup>-3</sup> (30 minute readings, excluding days when  
492 construction was taking place in the area), however these above background levels were most likely  
493 due to vehicle emissions.

494

#### 495 **4.2 PDC surges and hazard to displaced populations**

496 The only detached surge deposits investigated were from the 5 November eruption in Bronggang  
497 village. The substantial grain size differences between the surge deposits collected outside of houses  
498 compared with samples collected indoors, however, emphasizes the nature of the hazard. Ash  
499 entered houses through ventilation gaps in the walls and below the roofs, preferentially excluding the  
500 coarser, basal layer of the surge. One-quarter (23.3 and 25.5 vol. %) of the sample volume for 2  
501 samples collected from indoor floor deposits were in the sub-10 µm fraction. Additionally, more than  
502 one-third (35.3 and 35.7 vol. %) of the 2 samples collected from 1 metre surfaces inside homes where  
503 some people died was sub-10 µm. Only 2 people were inside a house at the time the surge struck  
504 and survived their thermal injuries; both were diagnosed as having possible inhalation burns to the

505 respiratory tract. Twenty-five bodies were retrieved by rescuers from inside houses and eighteen from  
506 outside; 11 others died in, or on the way, to hospital, but their location at the time of injury was not  
507 known. Inhalation of fine ash in hot surges (>300 °C) into the small airway combined with severe  
508 burns to the body has a bad prognosis (Kobayashi et al., 1993).

509

510 Fresh surge-related deposits in protected indoor environments are substantially finer than the outdoor  
511 surge deposit and airfall tephra measured at Merapi or any other volcanoes (Horwell, 2007). Results  
512 from previous studies on ash particles from Montserrat comparing grain-size and crystalline silica in  
513 BAFs and corresponding co-PDC plume deposits (Horwell et al., 2001; Horwell et al., 2003b) suggest  
514 that the finer surge deposits from raised surfaces, which stayed lofted for longer, would be enriched in  
515 cristobalite compared to their corresponding bulk floor samples; however, the opposite was observed  
516 in the Merapi samples. As many of the affected houses remained standing, if damaged and ash  
517 covered, these indoor deposits are a post-eruption source of exposure during the clean up and  
518 rebuilding process. Floors, walls, and belongings were covered with a high proportion of respirable  
519 ash with up to 8.8 wt. % cristobalite, which was isolated from the seasonal rains. We recommend that  
520 hazard managers and NGOs involved in re-habitation efforts provide respiratory protection in the form  
521 of disposable light weight, high efficiency masks and disseminate information on how to reduce  
522 exposure, e.g., by first wetting deposits. The kitchens of most houses are commonly situated outside  
523 the main living area, so indoor air pollution from cooking fires is much reduced.

524

#### 525 **4.3 Respiratory hazard to occupational groups**

526 The number of workers and the size of the sand-mining operations indicated the importance of this  
527 occupation as the main source of regular and high exposure to volcanic ash. A variable proportion of  
528 adults in villages work full or part time in sand mining, with up to 20 % in a few villages. Whereas  
529 deposits at other volcanoes are often quarried with heavy machinery and explosives, e.g., on  
530 Montserrat, deposits at Merapi are almost entirely excavated by hand. The heavy manual labour  
531 involves excavating the ash using shovels and loading it onto trucks.

532

533 Material from lahars in the area we visited was being taken directly from the river and river banks, and  
534 was wet when extracted. Exposure levels for workers on site were therefore low. The lahar sample

535 investigated contained no particles smaller than 10  $\mu\text{m}$ . At the sample collection distance of  
536 approximately 20 km from the crater, the finer fraction had been removed in the streamflow and only  
537 the coarser material was left as sediment. By the time lahars at Merapi reach approximately 20 km  
538 from the summit they have deposited so much of their sediment load that they become dilute muddy  
539 streamflows (Newhall et al., 2000). Further field work is required to determine whether the fines will  
540 make up a more significant fraction in areas of active quarrying once the deposits begin to dry. The  
541 material is mined and transported to other districts while it is still wet, but further processing of the  
542 material off-site before use in construction could considerably alter the grain size and surface of the  
543 material, i.e., by generating freshly-fractured crystalline silica surfaces which has been shown to  
544 increase reactivity (Fubini, 1998).

545

546 The range of samples provides insight into the hazard posed by variably-sourced ash at Merapi;  
547 however, the limited suite does not allow for a comprehensive definition of the occupational hazard for  
548 individuals involved in mining of volcanic deposits. Crucially, we were unable to fully investigate the  
549 PDC deposits as a mining hazard, such as the 18 km flow down the Kali Gendol valley, due to the  
550 temperature and cohesiveness of the recently deposited unit. These pyroclastic flow deposits are  
551 mined to over 10 metres depth and eventually removed from the entire channel. Much of this is fine  
552 material and in dry weather there could be high exposure to the dust from directly working in the  
553 deposit. To get an indication of the hazard from mining PDC deposits, we collected 16 dome rock  
554 samples from the hot deposit in Kali Gendol valley deposited on 5 November. These rocks contained  
555 up to 8 wt. % cristobalite, which is comparable to that seen in dome rock at Soufrière Hills volcano  
556 (Horwell et al., In review). Determining which samples originated from the dome that existed prior to  
557 the 26 October eruption as opposed to the new dome which grew in the lead up to the 5 November  
558 eruption was difficult in the field; however, high levels of cristobalite in new dome material would  
559 indicate rapid cristobalite formation in the 5 days of dome growth, in line with previous constraints by  
560 Williamson et al. (2010) and Horwell et al. (In Prep-a). These results give an early indication that  
561 crystalline silica may be a factor when mining PDC deposits during the dry season.

562

563 The fertile soils surrounding Merapi are intensely farmed, with nearly half of the population of the  
564 Yogyakarta province dependent on agriculture for their livelihood (Wilson et al., 2007). Two of the



565 tephra fall samples were specifically collected to gain insight into the effects of volcanic ash on the  
566 respiratory hazard posed to farmers (MER\_10\_17 and \_18). MER\_10\_17 was taken from a flat,  
567 terraced crop field, and MER\_10\_18 directly from vegetation. Aside from MER\_10\_12, these two  
568 samples contained the largest proportion of fine material of all tephra fall samples (approximately 10  
569 vol. % sub-4  $\mu\text{m}$  material), as well as 9.2 and 4.2 wt. % cristobalite, respectively. During the field  
570 mission, most ash had already dispersed from urban areas, but ash was not as quickly washed away  
571 or removed in agricultural areas. In the 6 months following the eruption, however, virtually all of the  
572 agricultural fields had been ploughed and reworked (John Pallister, personal communication). Ash will  
573 be removed from any un-worked agricultural land through aeolian re-mobilisation of deposits as they  
574 dry out and through incorporation into the soil horizons.

575

#### 576 **4.4 Bioreactivity and health**

577 The large volume of erupted material over the southern flank of Merapi is likely to expose sand miners  
578 to high levels of dust in the dry season. The cristobalite in the fine dust may be enough to adversely  
579 influence the course of pulmonary tuberculosis, which is common in Indonesia (it ranks number five in  
580 the world for the incidence of pulmonary tuberculosis (World Health Organisation: Regional Office for  
581 South-East Asia, 2011)). Crystalline silica is known to increase the risk of developing tuberculosis and  
582 to exacerbate its clinical course in certain groups of underground miners (Hnizdo and Murray, 1998;  
583 teWaterNaude et al., 2006). The abundance of crystalline silica in the ash and the potential exposure  
584 of the sand miners at Merapi to airborne dust warrants further investigation in the dry season. Heavy  
585 dust exposure may also contribute to the development of chronic obstructive pulmonary disease,  
586 especially in smokers (smoking is common in Indonesia).

587

588 Sustained inflammation in the lung plays a key role in the fibrotic changes in silicosis or mixed lung  
589 fibrosis which are caused by siliceous natural dusts. Our various assays, however, showed minimal  
590 bio-reactivity, indicating that the crystalline silica may be less reactive in a mixed dust or when hosted  
591 in a mineral matrix, e.g., Donaldson et al. (2001). Similar results have been observed at Mt. St.  
592 Helens (Vallyathan et al., 1984), Soufrière Hills (Cullen et al., 2002), despite much higher levels of  
593 cristobalite (up to 20 wt. %) (Horwell et al., In Prep-a), and Rabaul (Le Blond et al., 2010). These  
594 toxicity tests are not infallible guides to disease end-points and should not be taken as ruling out

595 chronic disease processes if exposure to the ash is sufficiently high. Instead, these *in vitro* tests if  
596 found positive may point to disease mechanisms requiring further research: the cellular mechanisms  
597 behind chronic fibrosis due to silica are still poorly understood.

598

599 Compared to other ash leachate composition, the Merapi ash samples show relatively high water-  
600 soluble Zn and Cu contents (Armienta et al., 2002; Christenson, 2000; Cronin et al., 1998; Hinkley  
601 and Smith, 1982). This could reflect a material which underwent prolonged exposure to high  
602 temperature volcanic gases prior to the eruption (i.e., within the dome), resulting in deposition of  
603 metallic compounds, predominantly as sulphate and chloride salts (Moune et al., 2010). The  
604 unusually high levels of soluble Ni, Cr and Zn from MER\_10\_03 are higher than have been measured  
605 in water leachates from other volcanoes (Geoff Plumlee, personal communication), and possibly  
606 reflect ash derived from edifice/conduit/dome where intense metal deposition from hot magmatic  
607 gasses occurred. The concentrations of Mn, Ni, Pb and Cd in the Merapi ash samples fall in the  
608 ranges reported for other volcanoes. From a respiratory health hazard perspective, the highly soluble  
609 Zn and Cu may be important. Zinc compounds such as  $ZnCl_2$  may be involved in reactions with cell  
610 mitochondria (e.g., Lemire et al. (2008)), while Cu may play a role in the inflammatory response of the  
611 lung tissues (e.g., Rice et al. (2001)).

612

613 For this study, attendances at clinics in evacuation centres or hospital attendance data were not  
614 adequate for the surveillance of patients with complaints of eye and nose irritation or acute respiratory  
615 ailments.

616

617 We recommend future research to evaluate the health risk to the thousands of people employed in  
618 mining and processing of the tephra-sand as aggregate for the Indonesian construction industry.  
619 Aggregate mining utilizing mainly manual labour is extremely common at Indonesia's active and  
620 recently active volcanoes, and, consequently, occupational exposure to volcanic particulate is likely a  
621 common issue throughout the country. An exposure assessment study over the wide range of  
622 activities performed by the workers on the main types of deposits should be carried out in the dry  
623 season, with tephra samples concurrently collected for analysis. At least one cohort of miners should  
624 be studied over time, incorporating lung function studies and symptom recording, alongside a control

625 group of non-miners. This could serve as a case study for a national level assessment of the risks  
626 involved in aggregate mining, and should be undertaken alongside a study of hospital and clinic  
627 routine statistics for evaluating the prevalence of TB in the Merapi area compared to non-mining  
628 areas.

629 **5. Conclusions**

630 The October-November 2010 eruption was one of the most explosive eruptions of Mt Merapi in the  
631 past two centuries, providing unusually extensive tephra deposits, and leaving a legacy of exposure to  
632 fine ash containing low levels of crystalline silica for the thousands of people who work in the  
633 aggregate industry. There is a need to study this group further to evaluate their risk from their long  
634 term occupational exposure to ash and its effect on the common, but serious, condition tuberculosis.  
635 Our protocol, comprising a suite of mineralogical and toxicological assays, serves to better define the  
636 potential of the dust to become a respiratory hazard. Health impact studies are essential for informing  
637 officials and the public on the potential health impacts of volcanic emissions and on where to focus  
638 resources for protecting public health.

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640

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653 **Table Captions**

654 **Table 1:** Summary of sample information and analytical techniques carried out on each sample.

655 Further studies include: BET, EPR, iron release, haemolysis, oxidative capacity, inflammatory  
656 potential, leachate. Sample MER\_10\_10 was excluded from the study due to a lack of eruption and  
657 collection information. State of ash samples upon collection is reported in Table 2. Samples supplied  
658 by <sup>a</sup>Agung Harijoko, Gadjah Mada University; <sup>b</sup>Marie Boichu, Cambridge University, and Noer Cholik,  
659 BPPTK; <sup>c</sup>Maharani Hardjoko, Save the Children; <sup>d</sup>Jean-Christophe Komorowski, Institut de Physique  
660 du Globe de Paris; <sup>e</sup>Jochen Berger, University of Hohenheim.

661 **Table 2:** Results from grain size analyses for 'respirable' (<4 µm) and 'thoracic' (<10 µm) size  
662 fractions, the IAS-XRD method of quantifying cristobalite (<3 wt. % error), and BET specific surface  
663 area measurements for the set of samples examined further for reactivity. Data for MER\_arc are from  
664 Horwell et al. (2007) and Horwell (2007) except for cristobalite content which was quantified for this  
665 study.

666 **Table 3:** XRF analyses for samples MER\_10\_01 – MER\_10\_12 plus an archived sample from activity  
667 in 1998 (MER\_arc). Samples MER\_10\_01, \_03, \_04, \_12 and \_arc were collected pristine and  
668 believed to be sourced from a single eruptive event. Results are organised by dry versus  
669 wet/composite as the data for wet/composite samples do not necessarily reflect magmatic  
670 composition. All data were collected for this study. Results are quoted as component oxide weight  
671 percent and recalculated to include loss on ignition (LOI) in final total.

672 **Table 4:** Results from leachate analyses for samples MER\_10\_02, \_03, \_04 and \_12.

673 **Figure Captions**

674 **Figure 1:** Protocol for rapid assessment of health hazard after Le Blond et al. (2010).

675 **Figure 2:** Location map of Merapi volcano with locations of samples analysed in this study and town  
676 and river locations for geographical reference.

677 **Figure 3:** SEM images of Merapi ash showing (a) a particle which appears to be crusted with sub-  
678 micron particles, and (b) a single inhalable particle with numerous sub-micron particles adhered to the  
679 surface.

680 **Figure 4:** Total alkali vs. silica plot for selected Merapi samples. 'Dry ash' are samples collected fresh  
681 and believed to be of a single eruptive event (MER\_10\_01, \_03, and \_04). Wet/composite are  
682 samples collected following exposure to the environment and/or are believed to be composite  
683 samples from multiple eruptive events (MER\_10\_02, \_05-09, and \_11). The indoor surge sample is  
684 MER\_10\_12. MER\_arc sample was collected 200 m from a PDC on the slopes of the volcano during  
685 a dome collapse eruption in 1998.

686 **Figure 5:** Hydroxyl radical generation after 30 minutes against total iron released at day 7 for a  
687 subset of four Merapi samples plus the 1998 ash sample (MER\_arc), Min-U-Sil quartz standard, and  
688 four ash samples from other volcanoes for comparison: Cerro Negro (1995), Etna (2002), Pinatubo  
689 (1991), Soufrière Hills (MBA 5/6/99). Min-U-Sil quartz and MER\_arc values are those published in  
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Table 1

Table 1

Sample	Eruption Date	Collection Date	Deposit Type	Collection Location	Distance from Vent (km)	Collection Coordinates	GSA	XRF	XRD	SEM	Further Studies
MER_10_01 <sup>a</sup>	30-Oct-10	30-Oct-10	tephra fall	Jogonalan Lor, Bantul (Rumah)	32.0 SSW	07° 49.6850 S 110° 21.0817 E	✓	✓	✓	✓	
MER_10_02 <sup>a</sup>	26-Oct-10	31-Oct-10	tephra fall	Kepuh Harjo	5.5 S	07° 35.4928 S 110° 27.0510 E	✓	✓	✓	✓	✓
MER_10_03 <sup>a</sup>	05-Nov-10	05-Nov-10	tephra fall	Jogonalan Lor, Bantul	32.0 SSW	07° 49.6850 S 110° 21.0817 E	✓	✓	✓	✓	✓
MER_10_04 <sup>b</sup>	31-Oct-10	31-Oct-10	tephra fall	BPPTK, Yogyakarta	28.0 SSW	07° 47.8743 S 110° 23.0713 E	✓	✓	✓	✓	✓
MER_10_05 <sup>c</sup>	05-Nov-10	13-Nov-10	tephra fall	Desa Ngasem, Gulon, Magelang	18.0 SW	07° 36.1527 S 110° 17.2003 E	✓	✓	✓		
MER_10_06 <sup>c</sup>	05-Nov-10	13-Nov-10	tephra fall	Desa Manquncari, Sawangan, Magelang	16.0 W	07° 32.2773 S 110° 18.5593 E	✓	✓	✓		
MER_10_07 <sup>c</sup>	05-Nov-10	13-Nov-10	tephra fall	Desa Krogowanan, Sawangan, Magelang	11.0 W	07° 31.7755 S 110° 24.6047 E	✓	✓	✓		
MER_10_08 <sup>c</sup>	05-Nov-10	13-Nov-10	tephra fall	Desa Mbelan, Sawangan, Magelang	15.0 W	07° 32.6385 S 110° 18.6133 E	✓	✓	✓		
MER_10_09 <sup>c</sup>	05-Nov-10	13-Nov-10	tephra fall	Desa Ngadipuro, Dukun, Magelang	18.0 W	07° 33.7400 S 110° 17.2003 E	✓	✓	✓		
MER_10_11 <sup>d</sup>	Unknown	09-Nov-10	tephra fall	Jembatan Kali Juweh	6.0 NW	07° 29.9445 S 110° 24.6047 E	✓	✓	✓		
MER_10_12	05-Nov-10	30-Nov-10	surge	Bronggang, Argo Mulyo	17.0 S	07° 39.7307 S 110° 27.7743 E	✓	✓	✓	✓	✓
MER_10_13	05-Nov-10	30-Nov-10	surge	Bronggang, Argo Mulyo	17.0 S	07° 39.7307 S 110° 27.7743 E	✓		✓	✓	
MER_10_14	05-Nov-10	30-Nov-10	surge	Bronggang, Argo Mulyo	17.0 S	07° 39.7407 S 110° 27.8217 E	✓		✓		
MER_10_15	05-Nov-10	06-Dec-10	surge	Bronggang, Argo Mulyo	17.0 S	07° 39.7407 S 110° 27.8217 E	✓		✓		
MER_10_16	Unknown	08-Dec-10	lahar	Siderejo (Sinduharjo)	22.0 S	07° 44.9555 S 110° 26.9088 E	✓		✓		
MER_10_17	Unknown	03-Dec-10	tephra fall	SW slope Merbabu, near Selo	4.5 NW	07° 34.7699 S 110° 19.4860 E	✓		✓		
MER_10_18	Unknown	03-Dec-10	tephra fall	SW slope Merbabu, near Selo	8.0 NW	07° 31.0715 S 110° 22.7871 E	✓		✓		
MER_10_19 <sup>e</sup>	05-Nov-10	08-Dec-10	surge	N of Gadingan, E of Bronggang	17.0 S	07° 39.7367 S 110° 27.9133 E	✓		✓		
MER_10_20 <sup>f</sup>	Unknown	04-Dec-10	tephra fall	S of Dukun, Magelang	12.0 W	07° 33.8713 S 110° 20.3336 E	✓		✓		
MER_arc	11-19 Jul 1998	09-Aug-98	tephra fall	Volcano flanks	0.2	Unknown	✓	✓	✓		✓

Table 2

Sample Name	Ash Type	Grain Size (cumulative vol. %)		Surface Area (m <sup>2</sup> g <sup>-1</sup> )	Cristobalite (wt. %)
		< 4 μm	< 10 μm		
MER_10_01	dry ash	2.7	7.6		1.9
MER_10_02	wet, composite	15.6	28.9	0.99	3.2
MER_10_03	dry ash	8.4	17.0	0.51	2.7
MER_10_04	dry ash	5.8	14.5	0.78	6.0
MER_10_05	wet, composite	6.7	14.9		3.4
MER_10_06	wet, composite	2.9	6.3		4.3
MER_10_07	wet, composite	1.4	3.0		5.5
MER_10_08	wet, composite	7.4	16.7		4.4
MER_10_09	wet, composite	7.3	15.5		4.6
MER_10_11	unknown	9.2	19.5		6.0
MER_10_12	indoor surge	13.0	25.5	1.03	8.8
MER_10_13	surge fines	19.0	35.7		3.2
MER_10_14	surge fines	19.1	35.3		4.5
MER_10_15	indoor surge	12.8	23.3		7.9
MER_10_16	lahar	0.0	0.0		3.2
MER_10_17	wet, composite	10.1	20.5		9.5
MER_10_18	wet, composite	9.7	20.8		4.2
MER_10_19	outdoor surge	6.5	13.1		3.4
MER_10_20	dry, composite	5.6	12.0		10.3
MER_arc	dry ash	12.7	27.2	1.83	3.8

Table 3

	Sample	SiO <sub>2</sub>	TiO <sub>2</sub>	Al <sub>2</sub> O <sub>3</sub>	Fe <sub>2</sub> O <sub>3</sub>	MnO	MgO	CaO	Na <sub>2</sub> O	K <sub>2</sub> O	P <sub>2</sub> O <sub>5</sub>	SO <sub>3</sub>	LOI	Total
<i>dry</i>	MER_10_01	55.52	0.71	18.02	7.73	0.18	1.81	6.69	3.92	2.67	0.282	0.437	1.07	99.04
	MER_10_03	54.69	0.74	19.29	7.76	0.19	2.25	8.12	3.73	2.16	0.295	0.050	0.28	99.55
	MER_10_04	56.48	0.65	18.17	6.64	0.17	1.75	6.61	4.02	2.84	0.287	0.420	1.16	99.20
	MER_10_12	55.91	0.57	19.88	5.89	0.16	1.81	7.70	3.91	2.38	0.284	0.054	0.54	99.09
	MER_arc	59.28	0.52	18.53	5.34	0.16	1.43	6.20	4.02	2.99	0.29	n/a	1.15	99.90
<i>wet, composite</i>	MER 10_02	57.42	0.51	19.13	5.25	0.15	1.51	6.52	4.11	2.84	0.279	0.12	1.38	99.24
	MER 10_05	52.98	0.79	18.59	8.41	0.21	2.89	9.11	3.64	2.10	0.300	0.03	0.10	99.15
	MER 10_06	53.58	0.83	18.64	8.61	0.21	2.91	8.87	3.51	1.97	0.287	0.00	0.20	99.62
	MER 10_07	51.90	1.02	17.41	10.72	0.26	3.77	9.47	3.31	1.80	0.311	<0.002	-0.02	99.95
	MER 10_08	54.37	0.75	19.29	7.86	0.20	2.67	8.52	3.75	2.19	0.273	0.01	0.11	99.98
	MER 10_09	54.75	0.73	19.04	7.53	0.19	2.48	8.32	3.71	2.23	0.266	0.234	0.31	99.78
	MER 10_10	53.45	0.82	18.93	8.72	0.21	2.89	8.98	3.52	2.03	0.280	<0.002	-0.17	99.67
	MER 10_11	53.77	0.79	19.15	8.35	0.20	2.56	8.75	3.60	2.08	0.284	0.096	0.04	99.67

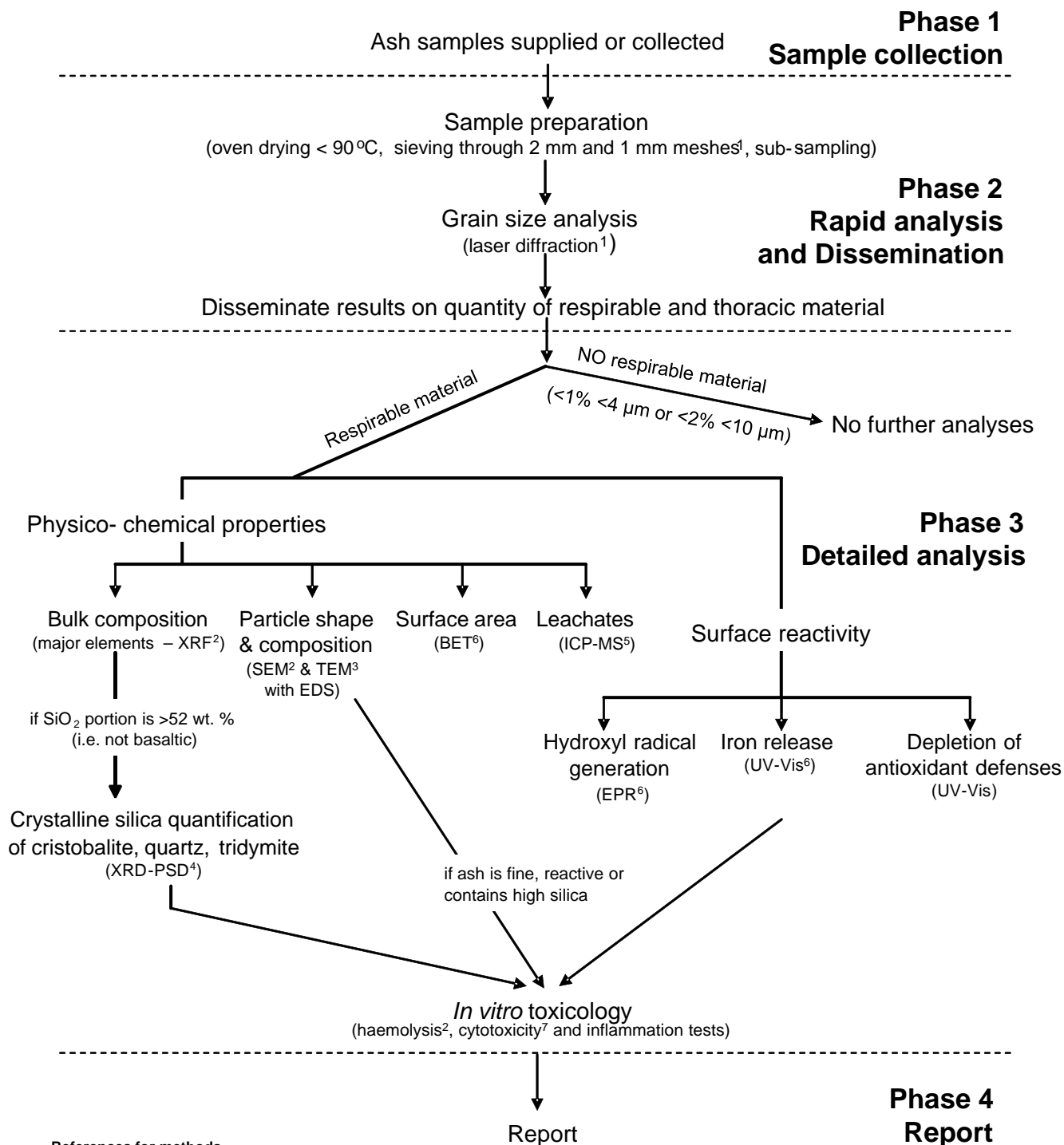
Table 4

	MER_10_02	MER_10_03	MER_10_04	MER_10_12
pH	5.8	5.9	5.1	6.3
<i>mg/kg ash</i>				
Si	<10	20.8	181.2	53.6
Al	20.0	16.3	63.7	8.0
Fe	7.3	16.3	10.9	5.8
Mg	94	26	166	36
Ca	1900	1351	4590	1013
Na	397	87	822	167
K	110	41	189	49
Mn	20	4	31	6
F	6	17	40	12
Cl	132	192	934	49
SO <sub>4</sub>	5450	2186	10574	1769
molal S/Cl	15.0	4.2	4.1	13.2
<i>µg/kg ash</i>				
As	0	0	0	0
Cd	15	12	49	3
Co	0	0	0	0
Cr	208	710	241	206
Cu	1738	1100	1188	1228
Ni	481	1415	511	453
Pb	18	42	15	27
Zn	3053	30401	4084	2753

Figure 1



# Protocol for analysis of bulk ash samples for health hazard assessment



**References for methods**

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For full references and method summaries please visit [www.ivhhn.org](http://www.ivhhn.org) or contact Dr Claire Horwell ([claire.horwell@durham.ac.uk](mailto:claire.horwell@durham.ac.uk))

Figure 2

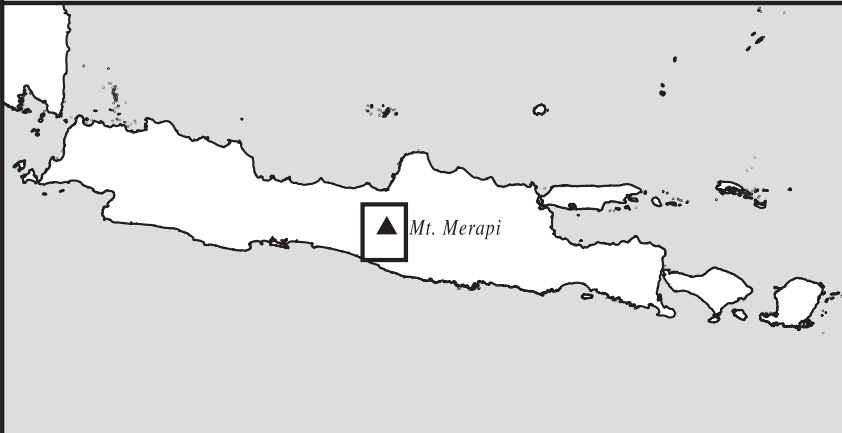
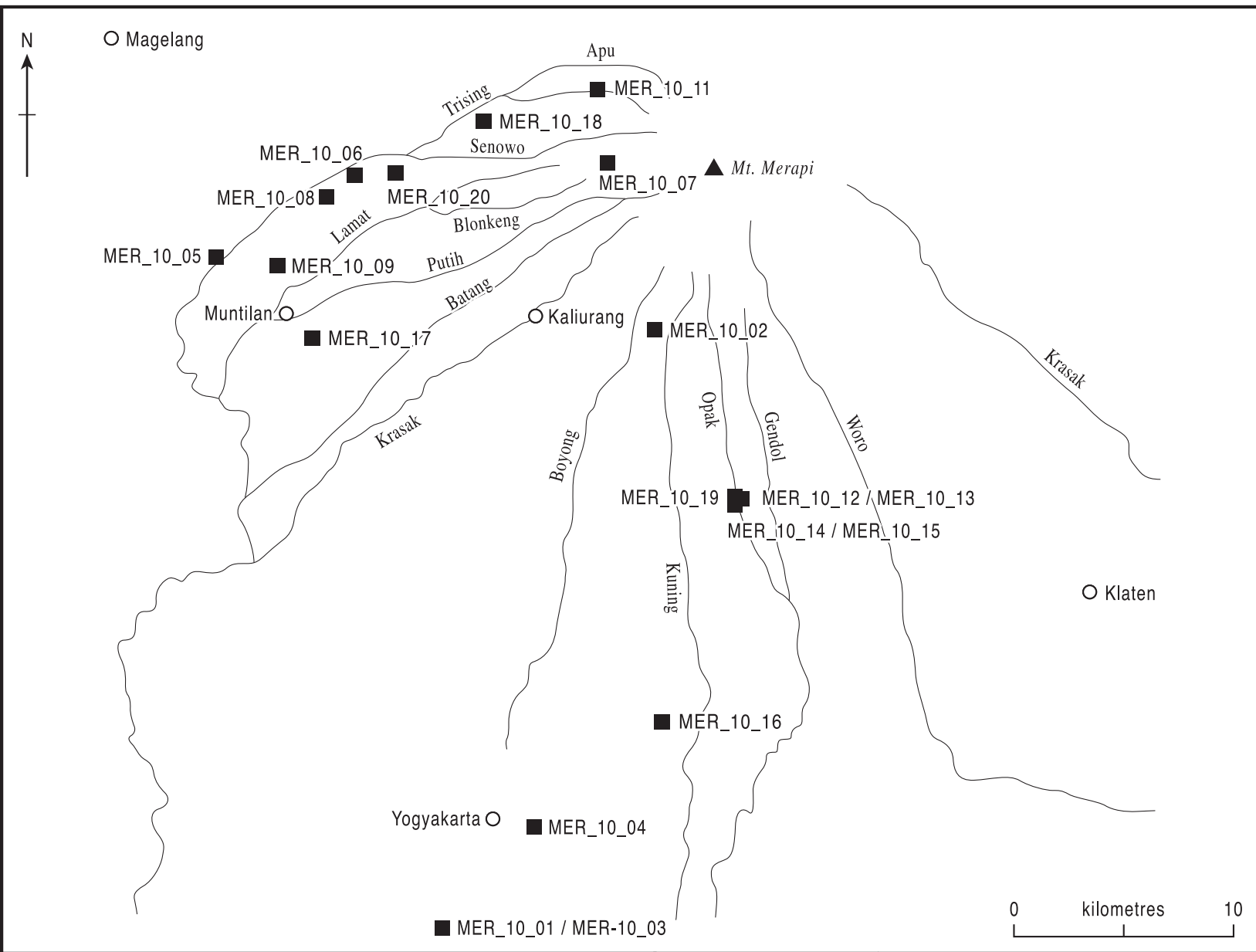




Figure 3

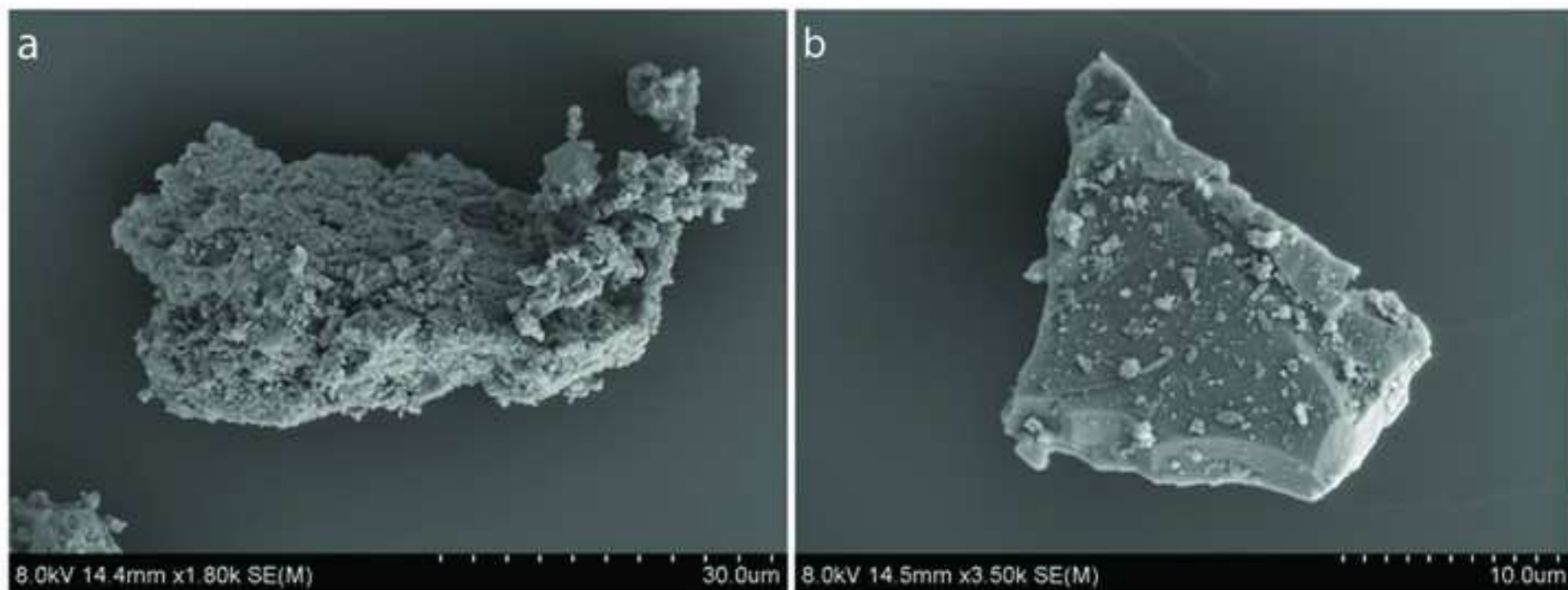


Figure 4

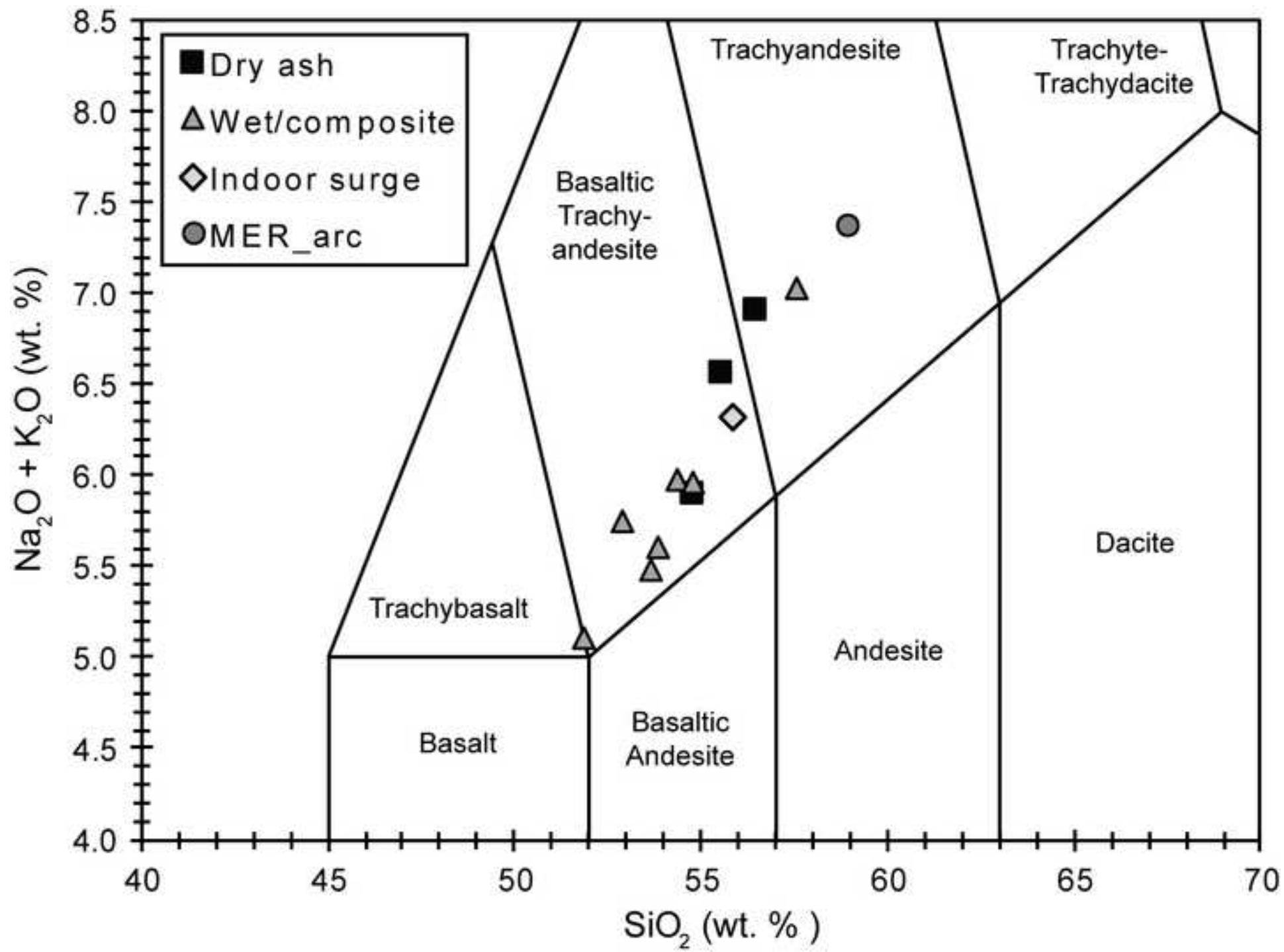


Figure 5

