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1	Host preference and growth patterns of ivy (Hedera helix L.) in a temperate alluvial forest
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## 27 ABSTRACT

29	Recent studies have highlighted the role of lianas in shaping stand dynamics both in tropical and
30	temperate forests. However, English ivy (Hedera helix L.), one of the most widespread liana in
31	Europe, has received little attention. We conducted a study in the Siro Negri alluvial forest (NW
32	Italy) to determine what factors most affected ivy distribution, and to investigate its interactions
33	with the trees in the stand. We evaluated the influence of tree size, age, species and neighborhood
34	crowding on ivy occurrence. In addition, growth ring widths were used to explore the development
35	pattern of climbing stems.
36	Fifty-two percent of trees in our study plots carried ivy, a value comparable to liana incidence found
37	in mature tropical forests. Tree characteristics and their spatial pattern significantly influenced ivy
38	distribution. Preferred hosts were large, isolated trees, while the effect of tree age and species on ivy
39	occurrence was marginal. Growth pattern analysis revealed that radial growth was positively related
40	to the available space on the tree trunk for each ivy stem. We conclude that neighborhood crowding
41	around trees and competition among climbing stems relying on the same trunk may reduce the
42	colonization rate of ivy.
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### 53 INTRODUCTION

54

Lianas are woody climbing plants that begin their life as terrestrial seedlings and later ascend to the 55 56 canopy using trees as structural supports (Putz and Mooney 1991). Over the past few years their 57 importance in forest biodiversity has been recognized (Schnitzer and Bongers 2002): they increase 58 woody species diversity and provide food and refuge for many animal species (Emmons and Gentry 59 1983; Campanello et al. 2007; Page et al 2010). Lianas play a central role in forest dynamics, as 60 they can suppress tree regeneration, and they compete with adult trees for light, water, and soil 61 nutrients. In some cases they have been found to reduce tree growth rates and fecundity (Ladwig 62 and Meiners 2010a; Ichihashi and Tateno 2011). Moreover, abundant lianas increase crown 63 susceptibility to wind and, in temperate climates, to snow damage (Allen et al. 2005). Lianas can 64 also influence forest composition, as they affect tree species differentially, modifying inter-specific 65 relationships. Furthermore, lianas can occupy available space on the ground earlier than pioneer tree 66 species, delaying gap colonization and forest succession (Laurence et al. 2001; Schnitzer and 67 Bongers 2002; Toledo-Aceves and Swaine 2008). 68 Since lianas have such an important role in forest dynamics, research on their distribution, ecology 69 and relationships with other plants is critical. Several studies have investigated liana dynamics at 70 stand level, revealing that abiotic factors, forest structure and stand dynamics influence their 71 distribution (Laurence et al. 2001; Allen et al. 2005; Londrè and Schnitzer 2006; Morrissey et al. 72 2009; Gianoli et al. 2010; Macía 2011; Nogueira et al. 2011). Few studies have taken an individual 73 tree approach to tree liana-tree relationships, and there is little information on how lianas interact 74 with their hosts, especially for temperate forests (Blick and Burns 2011; Ichihashi and Tateno 75 2011). As structural parasites, lianas need suitable supports to reach the canopy and maintain 76 themselves there. As a consequence, host tree characteristics, such as size and species, influence 77 lianas colonization patterns, growth rates, distribution and diversity (Leicht-Young et al. 2010; 78 Nogueira et al. 2011).

79 At a global scale, lianas are widespread in tropical forests, where they represent between 10 and 80 40% of the woody species (Schnitzer and Bongers 2002), but they are even common in several 81 temperate ecosystems. In Europe, there are few autochthon liana species, and white vine (Clematis 82 vitalba L.), grapevine (Vitis spp.) and English ivy (Hedera helix L.) are the most diffuse. The genus 83 *Hedera* originated in the Tertiary, experiencing a postglacial re-colonization from the 84 Mediterranean regions. As ivy belongs to a largely tropical family (Araliaceae), it benefits from 85 warm summers, and is intolerant to cold temperatures (Metcalfe 2005). It favours nutrient-rich and 86 moist soils. Unlike many of the woody species found in southern-European plain forests, ivy is an 87 evergreen species. It presents two distinct forms: the chamephyte form at the beginning of its life, 88 with prostrate, sterile stems and palmately-lobed shade leaves that can create extensive carpets on 89 the forest floor; and the phanerophyte form when suitable supports are available and ivy stems can 90 ascend toward better light conditions. The climbing stems support reproductive shoots, with 91 unlobed oval sun leaves (Nola 1997; Garfi and Ficarrotta 2003; Metcalfe 2005). Like other lianas, 92 ivy is a serious concern for forest managers. Despite its contribution to biodiversity maintenance and to nutrient cycling processes (Badre et al. 1998), ivy is often regarded as a problem, as it is 93 94 believed to reduce tree growth rates. Moreover, outside of its natural distribution areas, ivy is an 95 invasive species that poses a grave threat to native plants (Clarke et al. 2006; Biggerstaff and Beck 96 2007), and active management is required to control it.

97 Despite its wide home range, extending from southern Scandinavia to the Mediterranean area, few
98 studies have been conducted on ivy dynamics, and those mainly in the Rhine valley, France

99 (Trèmolières et al. 1988; Schnitzler 1995; Badre et al. 1998; Schnitzler and Heuzè 2006; Heuzè et
100 al. 2009). Research on other species can be only partially transposed to ivy, as morphological and
101 ecological features vary between liana species (Gianoli et al. 2010).

We conducted a study in the Siro Negri forest, one of the very few fragments in northern Italy of
alluvial forest, a preferential habitat for this species (Schnitzler and Heuzè 2006). Our general

104 purpose was to explore ivy dynamics on an individual tree basis. The first goal was to quantify ivy

abundance in the forest, as quantitative assessments of lianas in Europe are extremely scarce.
Subsequent analyses were conducted to study individual ivy-host relationships. We aimed at
identifying the characteristics of preferred host trees, such as size, age, species, and neighborhood
crowding, in order to better understand which factors affect ivy occurrence. We then explored the
relationships between ivy radial growth rates and host tree characteristics, to obtain information on
ivy development following host ascension.

- 111
- 112 METHODS
- 113

#### 114 Study site

115 The study was conducted within the Siro Negri Reserve, on the southern side of the Ticino Regional 116 Park, province of Pavia, northern Italy (45°12'N, 9°03'E, 65 m a.s.l.). Mean annual rainfall is about 117 800 mm, with most precipitation in spring and autumn, and the average annual temperature is 13.6 118 °C (Pavia weather station, 10 km from the study site). The forested area, described as a Polygonato 119 multiflori-Quercetum roboris association (Sartori 1984), covers about 9 ha, and is one of very few 120 relicts of the original alluvial forest in northern Italy. The main tree species are Acer campestre L., 121 Carpinus betulus L., Crataegus monogyna Jacq., Populus alba L., Populus nigra L., Quercus robur 122 L., Robinia pseudacacia L., Ulmus minor Mill. The forest accommodates trees of over 100 years. 123 The last important human disturbances coincide with the two world wars, when there was an 124 elevated demand for firewood (Motta et al. 2009), while the forest has been unmanaged since the establishment of the reserve in 1970. 125 126

## 127 Field sampling

Two permanent plots ( $50 \times 50$  m) were established during 2005 at the south-eastern border of the reserve (Motta et al. 2009). Two other permanent plots ( $60 \times 60$  m) were established at less than 100 m of distance in the inner part of the reserve during 2009. Trees with diameter at breast height 131  $(DBH) \ge 7.5$  cm were identified, labeled with numbered plastic tags and mapped. All trees in the 132 reserve were mapped between 2009 and 2011.

133 During 2011, tagged trees in the four permanent plots (total area equal to 1.22 ha) were re-surveyed, 134 and a census of ivy was carried out. For each tree, we measured the diameter of living ivy stems > 135 0.5 cm at 1.3 m from the base. All the stems were measured without distinguishing between ramets 136 and genets, because it was difficult to identify their origin, and their dynamics and effects on trees 137 are comparable (Gerwing et al. 2006; Schnitzer et al. 2006; Campanello et al. 2007). The degree of 138 ivy occupancy on trees was classified according to a five class system: 0, ivy absent or below 1.3 m 139 on the tree trunk; 1, ivy present on the lower half of the trunk; 2, ivy reaching the second half of the 140 trunk, but not the crown; 3, ivy on the lower part of the principal branches; 4, ivy covering the 141 whole crown.

142

### 143 **Dendrochronological analysis**

An increment core from all the trees with  $DBH \ge 10$  cm and from a subsample of trees with DBH <10 cm was taken at 50 cm from the tree base. In addition, an increment core was taken at 50 cm from the tree base from 60 randomly chosen ivy stems in the plots. The stems were cored from side to side to get a core with two radii, in order to account for possible asymmetrical growth of ivy stems. In the laboratory, all the cores were fixed to wooden supports and prepared with a razor to obtain an optimal surface resolution.

150 Cores from trees were used to estimate tree age. For cores that did not intersect the pith, its position 151 was estimated by means of a pith locator, and the number of missing rings was estimated (Motta 152 and Nola 2001). When it was not possible to estimate missing rings, the core was discarded. 153 Ring widths of ivy cores were measured to the nearest 0.01 mm, using LINTAB device and 154 CATRAS software (Aniol 1983). Cores with rotten wood or other irregularities were discarded.

155 Cores were also discarded when the two series of radii for each did not match (visual comparison).

Series from the two radii of each of the remaining 43 cores (28.3% discarded) were averaged toestimate annual stem growth rate.

158

## 159 Data analysis

160 The relationships between ivy and trees were explored on an individual tree basis. Descriptors of 161 ivy, computed for each host tree, were number of stems, diameter sum, basal area, diameter of the 162 largest stem, and the degree of ivy occupancy. Tree features were diameter, age (when available), 163 and species. Also, to analyze the effects of neighborhood crowding on ivy-tree interactions, we 164 calculated the number and the sum of diameters of trees in a circle of 10 m around the tree; and the 165 neighborhood ratio calculated as the ratio between the sum of neighbors' diameters and the subject 166 tree diameter (Daniels et al. 1986; Castagneri et al. 2008). A radius of 10 m was selected because 167 the largest trees in the stand have a crown radius > 8 m, thus they compete with trees at a distance 168 of more than 8 meters.

169 Firstly, we ran all the variables in a Redundancy analysis (RDA), to order ivy attributes in relation 170 to tree and neighborhood characteristics. RDA was done on Canoco 4.0 software (Ter Braak and 171 Smilauer 1998) and the statistical significance of the relations between variables was tested by a 172 Monte Carlo test with 10,000 permutations. Next, to refine the assessment of ivy-tree interaction, 173 we performed two types of models: (1) binary logistic regression models to evaluate the 174 characteristics that determined the presence/absence of ivy on a tree, with stepwise entry of 175 predictor variables (tree characteristics and neighborhood measures); (2) General Linear Models 176 (GLMs) to determine which factors influenced ivy abundance on host trees, again, via stepwise 177 entry of predictor variables (tree characteristics and neighborhood measures). In this case, response 178 variables were number of ivy stems on the tree, sum of stem diameters, and their basal area. When 179 continuous variables were not normally distributed, a transformation was applied to improve 180 linearity of relationships. Variable collinearity and distribution of the residual were checked for 181 each model. Finally, we analyzed ivy growth rate in relation to tree characteristics, neighborhood

- 182 crowding and ivy abundance. Individual growth rate, which was the response variable, was
- 183 calculated as mean radial increment during the last 5 and 10 years, obtained from the ivy cores. All
- 184 models were computed using SPSS 17.0 (SPSS Inc. Chicago IL).
- 185

#### 186 **RESULTS**

187

#### 188 Ivy distribution

189 Ivy density in the Bosco Negri alluvial forest was 657, 346 and 225 stems per hectare considering

190 stems of respectively  $\ge 0.5$ , 1 and 2.5 cm (three diameter thresholds commonly used in liana

191 censuses, Gerwing et al. 2006). The basal area of stems  $\ge 0.5$  cm was 5609 cm<sup>2</sup> per hectare. Of the

surveyed trees (n = 374), 52% hosted one or more ivy stems, although ivy was found only on the

- tree trunk in many cases, and just 17% of the trees had ivy at crown level (classes of occupancy 3
- and 4). Other liana species (*Clematis vitalba* L., *Vitis* spp.) occurred sporadically (0-2 individuals
- 195 per plot).

196 A few differences were observed between the four plots. Plot 3, characterized by low tree density 197 and large diameters, had the highest percentage of trees hosting ivy (Table 1), the largest number of 198 ivy stems larger than 0.5 cm (Fig. 1), and the largest ivy basal area (Fig. 2). Like Plot 3, Plot 4 was 199 located in the inner part of the reserve, but had a higher tree density and a lower mean diameter. It 200 had fewer ivy stems  $\ge 0.5$  cm, while the number of stems  $\ge 2.5$  cm was close to that in Plot 3. Plots 201 1 and 2, located at the forest border, had a much higher tree density. The percentage of trees hosting 202 ivy, the density of stems  $\geq 1$  and 2.5 cm, and ivy basal area were rather similar to those observed in 203 Plot 4, whereas the number of stems  $\geq 0.5$  was greater (Fig. 1).

204

### 205 Ivy-tree relationships

RDA highlighted the complexity of relationships between ivy and trees attributes. In the first
analysis (Fig. 3a) we included the descriptors of ivy, tree and neighborhood but not tree age for

208 each of the 374 sampled trees. In the second analysis (Fig. 3b) we included age as a variable. As 209 this information was not available for all the trees, the second RDA was performed on a sub-set of 210 207 samples. Relationships between variables were statistically significant (p < 0.001) in both the 211 RDA analyses, and the models explained 43.8% and 44.1% of variance respectively. 212 The ivy descriptors *HE\_ba*, *HE\_s\_d*, *HE\_d\_M*, *HE\_n* and *HE\_cl* expressed liana abundance on 213 trees in different ways, and they were positively associated to each other. With regards to ivy-host 214 interactions, while tree diameter  $(TR_d)$  and age  $(TR_age)$  were positively related to ivy descriptors, 215 neighborhood crowding (expressed as neighbors number  $(NE_n)$ , sum of diameters  $(NE_s_d)$ , and 216 neighborhood ratio (NE\_ra)) were negatively related to ivy abundance. In terms of species, RDA 217 showed that Q. robur trees were big and isolated, hosting large amounts of ivy whereas A. 218 campestre and C. monogyna trees were small with many neighbors, and hosted little ivy. The other 219 species were in an intermediate position. The inclusion of plot as a factor did not significantly 220 improve RDA (the variance explained by the model was 44.3%, graph not shown), suggesting that 221 the observed relationships were not influenced by site characteristics. RDA provided a general 222 pattern of ivy-host relationships. However, it was not possible to identify which tree characteristics 223 were most important in favoring ivy, as many variables correlated with each other. To discriminate 224 between these factors, several models were applied. The logistic regression model showed that the 225 presence of ivy was positively related to big tree diameters and negatively related to the number of 226 neighboring trees (Table 2), indicating that larger isolated trees were more likely to host ivy than the others. Tree species increased the Nagelkerke  $R^2$  from 0.276 to 0.303, but was not significant at 227 p = 0.05. Tree age was positively related to ivy, while the sum of diameters of neighbors and 228 229 neighborhood ratio correlated negatively, but these factors were discarded as redundant in the 230 stepwise regression. Including plot as a factor did not improve the model, indicating that the 231 observed patterns did not vary significantly between the four plots. 232 To determine which characteristics of host trees (n = 194) influenced the amount of ivy, we looked

at the number of ivy stems per tree, and their sum of diameters and basal area. These direct

234 measurements of ivy abundance correlated more closely to tree descriptors than diameter of the 235 biggest stem and tree occupancy category. GLMs indicated that abundance of ivy was influenced by 236 the same factors that determined its presence. Tree diameter was the most important feature, 237 positively influencing ivy abundance, while the number of neighbors had a negative effect (Table 238 3). Tree species emerged as a significant factor only in predicting ivy stem number, with O. robur 239 and, more evidently, C. monogyna, hosting fewer stems that the other species. The best model performed similarly in predicting the number of ivy stems on the tree and their sum of diameters, 240 241 while it explained a lower percentage of variance for ivy basal area.

242

### 243 Ivy growth patterns

The largest sampled ivy stem had a diameter of 14 cm at 50 cm from the base, and 53 counted rings, while the oldest one had 69 rings, and a diameter of 9.5 cm. Five stems had an estimated age > 50 years. The mean annual radial growth rate was 1.05 ( $\pm$  0.54) mm over the last five years, and 1.03 ( $\pm$  0.47) mm over the last ten years. Relationships between host tree characteristics and ivy growth were rather weak, and only host tree age was positively related to the rate of growth over the last ten years (Pearson's r = 0.36, p < 0.05, n = 34).

250 We also explored possible interactions between ivy stems growing on the same trunk. The basal 251 area of other stems on the same host tree was negatively related to the ivy growth rate over the last 252 five years (Pearson's r = -0.36, p < 0.05, n = 43) and over the last ten years (r = -0.31, p < 0.05, n = 253 43). To analyze this finding in greater detail, we tested whether ivy growth was related to the surface available on the tree trunk for each ivy stem, calculated as the ratio between the trunk 254 255 circumference and the sum of ivy stems' diameters. Available space was positively correlated to the growth rate both over the last five years (Pearson's r = 0.39, p < 0.01, n = 43) and over the last ten 256 257 years (r = 0.40, p < 0.01, n = 43) (Fig. 4).

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#### 260 **DISCUSSION**

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#### 262 Ivy abundance

263 An increase in the abundance and distribution of lianas has been observed both in tropical and in 264 temperate forests over recent decades, possibly due to the increase in atmospheric CO<sub>2</sub>, warmer 265 temperatures, and increased forest fragmentation caused by disturbances and land use changes 266 (Allen et al. 2005; Londré and Schnitzer 2006; Van der Heijden and Phillips 2008). It is therefore 267 important to carry out quantitative surveys, and to monitor temporal variations in liana incidence in 268 different forest ecosystems (Ladwig and Meiners 2010b; Schnitzer and Bongers 2011). It is usually 269 assumed that lianas have higher incidence in tropical than in temperate ecosystems (Schnitzer and 270 Bongers 2002; Leicht-Young et al. 2010). However, the percentage of trees carrying ivy in the Siro 271 Negri alluvial forest (45 - 79%) was similar to those observed in many mature tropical and 272 subtropical forests (Putz 1984; Campanello et al. 2007; Nesheim and Økland 2007); while the 273 number of stems and basal area per hectare was lower than that generally found in tropical forests 274 (Schnitzer and Bongers 2002; Schnitzer et al. 2006; Van der Heijden and Phillips 2008; Madeira et 275 al. 2009; Gianoli et al. 2010). Studies in two temperate alluvial forests in the Rhine valley, France 276 (Schnitzler and Heuzé 2006), found 61.9 and 29.4 English ivy stems per hectare, both being much 277 lower than the numbers observed in the Siro Negri forest.

In addition to macro-ecological factors, stand density and tree size distribution can also influence the quantity of liana (Van der Heijden and Phillips 2008; Madeira et al. 2009). Our survey was carried out in four plots very close to each other, but differing in tree density and diameter distribution. One of the plots (Plot 3) had a lower tree density compared to the others, but the number of ivy stems and basal area were similar or slightly larger than in the other plots. Within the same forest, patches with few but suitable supports can bear similar or larger amount of ivy than patches with many less suitable trees.

### 286 Host preference

287 Lianas are structural parasites that require suitable supports to access the canopy. Tree attributes, 288 such as species morphological features, size and age, may alter liana-host interactions, influencing 289 liana colonization rates (Campanello et al. 2007; Blick and Burns 2011; Nogueira et al. 2011). A 290 few recent studies, conducted both in tropical and temperate forests, indicate that the availability of 291 suitable host trees may have more influence than climate and soil properties on vine and liana 292 distribution (Nesheim and Økland 2007; Van der Heijden and Phillips 2008; Leicht-Young et al. 293 2010). In our study, we scrutinized several tree attributes, some of which have rarely been considered in previous research. The use of multiple models allowed the elimination of redundant 294 295 factors, and identification of the most important ones. Larger trees were more likely to host ivy than 296 smaller ones, and the largest host trees had the most ivy. A similar pattern has been observed in 297 several studies on lianas conducted in various ecosystems. However, different liana species exhibit 298 different host size preferences, depending on their climbing mechanism. Stem twiners and tendril 299 climbers better ascend small supports, and thus are more abundant on small diameter trees. On the 300 other hand, root climbers such as ivy adhere to trunks with adventitious roots rather than wrapping 301 around them, and prefer to ascend trees of larger diameters and rough bark (Putz 1984; Allen et al. 302 1997; Schnitzler and Heuzé 2006; Yuan et al. 2009; Madeira et al. 2009; Morrissey et al. 2009; 303 Leicht-Young et al. 2010; Blick and Burns 2011; Nogueira et al. 2011). It has been suggested that 304 large trees have more lianas because they are, generally, older than smaller ones, meaning more 305 colonization time for lianas (Campanello et al. 2007; Nesheim and Økland 2007; Ladwig and 306 Meiners 2010a). However, tree age is not usually measured directly. In our study, tree age of a sub-307 set of trees (55% of the samples) was assessed by tree-ring analysis. This attribute turned out to be 308 less important than diameter (it was not significant in the GLMs that included diameter as a 309 predictor variable), indicating that tree age was of secondary importance in ivy colonization. This 310 suggests that large trees hosted more ivy than smaller ones mainly because they provided a larger 311 surface that could accommodate many stems.

312 The majority of studies analyzing liana-tree interactions have found evident species associations, 313 and some tree species seem to have a higher probability of being colonized. Trees with rougher bark 314 host more root climbers, because they provide a good surface to ascend. Moreover, crown 315 morphology also seems to affect liana-tree relationships, and trees with high stature, small leaves 316 and a light crown host more lianas, due to better light conditions (Nesheim and Økland 2007; 317 Morrissey et al. 2009). In the Siro Negri forest, O. robur appeared to be preferred, and almost all 318 individuals of this species hosted ivy. However, thorough analysis revealed that ivy climbed Q. 319 robur trees because they were large and had few neighbors, whereas the effect of species per se was 320 negligible: when variations in size between species were excluded, different tree species were found 321 to be colonized to a similar degree by ivy. This finding, though, cannot be generalized, as the 322 species composition in the Siro Negri forest is limited to deciduous broadleaves, all characterized 323 by relatively rough bark in the adult phase: Schnitzler and Heuzè (2006) observed that ivy does in 324 fact show species preference in alluvial forests in northeastern France.

325 Using information on tree position in the stand, we analyzed the influence of neighboring trees on 326 the likelihood of a tree to carry ivy, to our knowledge an aspect never considered in previous 327 studies. Isolated trees, i.e. trees with few neighbors, were more likely to host ivy, and isolated host 328 trees had more ivy stems than host trees in crowded areas. As tree size and density are usually 329 inversely related, we checked whether the effect of neighborhood was discernible from that of tree 330 diameter. In the analyzed stand, the two parameters were slightly negatively correlated (r = -0.24). 331 However, the neighborhood ratio, expressed as the ratio between tree diameter and the sum of 332 neighbor's diameters, had a stronger (negative) relation with ivy descriptors than tree diameter 333 alone. The stepwise inclusion of neighborhood variables also enhanced the performance of all the 334 regression models. We concluded that neighborhood crowding directly influenced ivy colonization. 335 This is possibly because isolated tree trunks received more light than the others, making them more 336 attractive to light-demanding ascendant ivy stems. This finding provides an interesting contribution 337 to research on how forest structure affects lianas. Several authors report an increase in liana density

in disturbed areas and forest gaps, but the causes of this pattern have not yet received much attention (Ladwig and Meiners 2010b; Gianoli et al. 2010). We observed that smaller numbers of neighbors directly increased both tree likelihood of being colonized, and ivy abundance on host trees. It is likely that the incidence of light-demanding root climbers increases when a disturbance reduces the number of neighbors around surviving trees.

343

## **Growth patterns**

345 Very little information exists on the longevity of lianas, and on what affects their growth patterns 346 (Brandes et al. 2011). In comparison to other liana species, however, relatively numerous 347 dendrochronological studies have been conducted on ivy, possibly because tree rings are more 348 evident than in many tropical liana species. The mean annual growth rate for ivy found in the Bosco 349 Negri (mean-ring width in the last five years equal to  $1.05 \pm 0.54$  mm) and maximum age (69) 350 years), were roughly comparable to those reported by other authors. Nola (1997) measured a mean 351 annual radial increment of 0.94 mm ( $\pm$  0.34) and a maximum age of 68 years in southern Italy. 352 Garfi and Ficarrotta (2003) found a mean ring width of 0.86 mm ( $\pm$  0.25) and maximum age of 47 353 years in Sicily. Heuzè et al. (2009) reported a mean growth rate between 0.50 mm ( $\pm$  0.42) and 2.06 354 mm ( $\pm$  0.67), and maximum age of 66 years in northeastern France. 355 Growth pattern analysis can provide interesting indications on liana dynamics. Unlike ivy 356 distribution, we found that ivy growth was scarcely influenced by host tree characteristics and 357 neighborhood crowding. We did, however, find a slight but significant relationship between stem 358 growth and trunk surface available for each stem. We hypothesize that when many ivy stems rely 359 on the same tree, they compete. Positive interaction between liana stems has been reported, but this 360 was related to mechanical facilitation in the colonization of trees, as older stems provided support to 361 new ones for reaching the upper canopy (Campanello et al. 2007). We are not aware of any other

362 findings on competition between lianas, and further analyses should be conducted on other species

to support our finding.

364

## 365 Conclusions

Ivy is an important component of temperate forests but, especially outside Europe, it is often 366 367 regarded as a threat to other woody plants. Management of this species should be driven by better 368 knowledge of its ecology. As a structural parasite, its distribution depends strongly on the 369 availability of suitable host trees. Light-demanding climbing ivy stems ascend trees to reach better 370 light conditions. Indeed, we observed that preferred hosts were large, isolated trees that probably 371 receive more light than others. Tree species and tree age appeared weakly related to ivy distribution. 372 We conclude that the climbing form of ivy can be limited by an increase in neighborhood crowding 373 around potential host trees, i.e. by natural forest closure in the absence of disturbances. Competition 374 between ivy stems relying on the same trunk could also limit the vertical colonization of host trees.

375

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382

#### 383 **REFERENCES**

384

Allen BP, Pauley EF, Sharitz RR (1997) Hurricane impacts on liana populations in an old-growth
southeastern bottomland forest. J Torr Bot Soc 124:34–42

387 Allen BP, Sharitz RR, Goebel PC (2005) Twelve years post-hurricane liana dynamic in and old-

- 388 growth southeastern floodplain forest. For Ecol Manage 218:259–269
- 389 Aniol RW (1983) Tree-ring analysis using CATRAS Dendrochronologia 1:45-53

- 390 Badre B, Nobelis P, Trèmoliéres M (1998) Quantitative study and modelling of the litter
- 391 decomposition in a European alluvial forest. Is there an influence of overstorey tree species on the
- decomposition of ivy litter (*Hedera helix* L.)? Acta Oecol 19:491–500
- Biggerstaff MS, Beck CW (2007) Effects of English ivy (*Hedera helix*) on seed bank formation and
- 394 germination. Am Mid Nat 15:250-257
- Blick RAJ, Burns KC (2011) Liana co-occurrence patterns in a temperate rainforest. J Veg Sci
  22:868–877
- 397 Brandes AFN, Lisi CS, Barros CF (2011) Dendrochronology of lianas of the Leguminosae family
- 398 from the Atlantic Forest, Brazil. Trees 24:1045-1060
- 399 Campanello PI, Garibaldi JF, Gatti MG, Goldstein G (2007) Lianas in a subtropical Atlantic Forest:
- 400 Host preference and tree growth. For Ecol Manage 242:250–259
- 401 Castagneri D, Vacchiano G, Lingua E, Motta R (2008) Analysis of intraspecific competition in two
- 402 subalpine Norway spruce (Picea abies (L.) Karst.) stands in Paneveggio (Trento, Italy). For Ecol
- 403 Manage 255:651–659
- 404 Clarke MM, Reichard SH, Hamilton CW (2006) Prevalence of different horticultural taxa of ivy
- 405 (Hedera spp., Araliaceae) in invading populations. Biol Cons 8:149-157
- 406 Daniels RF, Burkhart HE, Clason TR (1986) A comparison of competition measures for predicting
- 407 growth of loblolly pine trees. Can J For Res 16:1230–1237
- 408 Emmons LH, Gentry AH (1983) Tropical forest structure and the distribution of gliding and
- 409 prehensile tailed vertebrates. Am Nat 121:513–524
- 410 Garfi G, Ficarrotta S (2003) Influence of ivy (Hedera helix L.) on the growth of downy oak
- 411 (*Quercus pubescens* sl) in the Monte Carcaci Naure Reserve (central-western Sicily). Ecol Mediter
- 412 29:5-14
- 413 Gerwing JJ, Schnitzer SA, Burnham RJ, Bongres F, Chave J, DeWalt SJ et al (2006) A standard
- 414 protocol for liana censuses. Biotropica 38:256–261

- 415 Gianoli E, Saldana A, Jimènez-Castillo M, Valladares F (2010) Distribution and abundance of vines
- 416 along the light gradient in a southern temperate rain forest. J Veg Sci 21:66–73
- 417 Heuzé P, Dupoueyb J-L, Schnitzler A (2009) Radial growth response of Hedera helix to
- 418 hydrological Changes and climatic variability in the Rhine floodplain. River Res Applic 25:393–
- 419 404
- 420 Ichihashi R, Tateno M (2011) Strategies to balance between light acquisition and the risk of falls of
- 421 four temperate liana species: to overtop host canopies or not? J Ecol 99:1071–1080
- 422 Ladwig LM, Meiners SJ (2010a) Liana host preference and implications for deciduous forest
- 423 regeneration. J Torr Bot Soc 137:103–112
- 424 Ladwig LM, Meiners SJ (2010b) Spatiotemporal dynamics of lianas during 50 years of succession
- 425 to temperate forest. Ecology 91:671–680
- 426 Laurence WF, Pérez-Salicrup D, Delamônica P, Fearnside PM, D'Angelo S, Jerozolinski A, Pohl L,
- 427 Lovejoy TE (2001) Rain forest fragmentation and the structure of Amazonian liana communities.
- 428 Ecology 82 (1):105-116
- 429 Leicht-Young SA, Pavlovic NB, Frohnapple KJ, Grundel R (2010) Liana habitat and host
- 430 references in northern temperate forests. For Ecol Manage 260:1467–1477
- 431 Londré RA, Schnitzer SA (2006) The distribution of lianas and their change in abundance in
- 432 temperate forests over the past 45 years. Ecology 87:2973–2978
- 433 Macía MJ (2011) Spatial distribution and floristic composition of trees and lianas in different forest
- 434 types of an Amazonian rainforest. Plant Ecol 212:1159-1177
- 435 Madeira BG, Espírito-Santo MM, D'Angelo Neto S, Nunes YRF, Arturo Sánchez Azofeifa G,
- 436 Wilson Fernandes G, Quesada M (2009) Changes in tree and liana communities along a
- 437 successional gradient in a tropical dry forest in south-eastern Brazil. Plant Ecol 201:291-304
- 438 Metcalfe DJ (2005) Biological flora of the British Isles *Hedera helix* L. J Ecol 93:632–645

- 439 Morrissey RC, Gauthier MM, Kershaw JA, Jacobs DF, Seifert JR, Fischer BC (2009) Grapevine
- 440 (*Vitis* spp.) dynamics in association with manual tending, physiography, and host tree associations
- 441 in temperate deciduous forests. For Ecol Manage 257:1839-1846
- 442 Motta R, Nola P (2001) Growth trends and dynamics in sub-alpine forest stands in the Varaita
- 443 Valley (Piedmont, Italy) and their relationships with human activities and global change. J Veg Sci
- 444 12:219–230
- 445 Motta R, Nola P, Berretti R (2009) The rise and fall of the black locust (*Robinia pseudoacacia* L.)
- 446 in the "Siro Negri" Forest Reserve (Lombardy, Italy): lessons learned and future uncertainties. Ann
- 447 For Sci 66:410
- 448 Nesheim I, Økland RH (2007) Do vine species in neotropical forests see the forest or the trees? J
- 449 Veg Sci 18:395-404
- 450 Nogueira A, Costa FRC, Castilho CV (2011) Liana abundance patterns: the role of ecological filters
  451 during development. Biotropica 43:442–449
- 452 Nola P (1997) Interactions between Fagus sylvatica L. and Hedera helix L. A dendroecological
- 453 approach. Dendrochronologia 15:23–37
- 454 Page NV, Qureshi Q, Rawat GS, Kushalappa CG (2010) Plant diversity in sacred forest fragments
- 455 of Western Ghats: a comparative study of four life forms. Plant Ecol 206:237–250
- 456 Putz FE, Mooney HA (1991) The biology of vines. Cambridge University Press, Cambridge
- 457 Putz FE (1984) The natural history of lianas on Barro Colorado Island, Panama. Ecology 65:1713–
- 458 1724
- 459 Sartori F (1984) Les forêts alluviales de la basse vallée du Tessin (Italie du nord). In: Cramer J (ed)
- 460 Colloques phytosocologiques, la végétation des forêts alluviales, pp 201-216
- 461 Schnitzler A (1995) Community ecology of arboreal lianas in gallery forest of the Rhine valley,
- 462 France. Acta Oecol 16:219–236
- 463 Schnitzler A, Heuzé P (2006) Ivy (Hedera helix L.) dynamics in riverine forests: Effects of river
- 464 regulation and forest disturbance. For Ecol Manage 236:12-17

- Schnitzer SA, Bongers F (2002) The ecology of lianas and their role in forests. Trends Ecol Evol
  17:223–230
- 467 Schnitzer SA, Bongers F (2011) Increasing liana abundance and biomass in tropical forests:
- 468 emerging patterns and putative mechanisms. Ecol Lett 14:397–406
- 469 Schnitzer SA, DeWalt SJ, Chave J (2006) Censusing and measuring lianas: a quantitative
- 470 comparison of the common methods. Biotropica 38:581–591
- 471 Ter Braak CJF, Smilauer P (1998) CANOCO 4. Centre for Biometry, Wageningen, The
  472 Netherlands
- 473 Toledo-Aceves T, Swaine MD (2008) Above- and below-ground competition between the liana
- 474 Acacia kamerunensis and tree seedlings in contrasting light environments. Plant Ecol 196:233–244
- 475 Trèmolières M, Carbiener R, Exinger A, Turlot JC (1988) Un exemple d'interaction non
- 476 compètitive entre espèces ligneuses: le cas du lierre arborescent (Hedera helix L.) dans la foret
- 477 alluviale. Acta Oecol 9:187–209
- 478 Van der Heijden GMF, Phillips OL (2008) What controls liana success in Neotropical forests?
- 479 Global Ecol Biogeogr 17:372–383
- 480 Yuan CM, Liu WY, Tang CQ, Li XS (2009) Species composition, diversity, and abundance of
- 481 lianas in different secondary and primary forests in a subtropical mountainous area, SW China. Ecol
- 482 Res 24:1361–1370
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## **TABLES**

493 Table 1 Tree stand characteristics and percentage of trees hosting ivy in the study plots494

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	Density	Mean diameter	Basal Area	Host trees	
_	(n/ha)	(cm)	(m²/ha)	(%)	
	498	26.7	28.4	48	
	560 122	28.1	34.7	45	
	122	46.6 30.2	20.9	79 55	
-	320	33.9	24.5	52	
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**Table 2** Results of binary logistic regression to evaluate the host tree characteristics that determine

# 522 the presence of ivy on a tree

	Nagelkerke R <sup>2</sup>	Hosmer - Lemeshow Test	Percentage Correct	Variable	Beta	Wald	р	95.0% C.I. lower	95.0% C.I. upper
Model	0.303	0.569	69.5	Diameter Neighbors Species	0.060 -0.111	15.853 25.266 9.582	<0.001 <0.001 0.088	1.032 0.857	1.092 0.934
				Constant	-0.564	0.401	0.527		
Diamete	r is the diam	eter at brea	st height of	f the tree, ne	aighbors is	s the num	ber of tre	es in a cii	cle of
10 m arc	ound the tree	, and specie	es is tree sp	ecies. p valu	1 es < 0.05	are indic	ated in b	old	
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**Table 3** Results of General Linear Models to predict the number of ivy stems on host trees, the sum

556 of their diameters, and basal area

	Number of stems			Sum of diameters			Basal area		
	R <sup>2</sup> Adj	Beta	р	R <sup>2</sup> Adj	Beta	р	R <sup>2</sup> Adj	Beta	р
Model	0.402		<0.001	0.375		<0.001	0.286		<0.001
Intercept		-0.530	0.302		-0.629	0.373		-3.451	0.018
Diameter		0.717	<0.001		0.704	<0.001		1.435	<0.001
Neighbors		-0.025	0.008		-0.030	0.021		-0.055	0.040
Quercus		-0.534	0.041		0.626	0.080		1.355	0.065
Robinia		-0.090	0.741		-0.004	0.992		0.532	0.489
Acer		-0.541	0.068		0.265	0.514		1.340	0.108
Ulmus		-0.428	0.159		-0.075	0.857		0.873	0.307
Crataegus		-0.726	0.004		0.112	0.742		0.893	0.204

559 Diameter is the diameter at breast height of the tree, neighbors is the number of trees in a circle of

560 10 m around the tree, and Quercus, Robinia, Acer, Ulmus and Crataegus are tree species. p values <

- 561 0.05 are indicated in bold

## 583 FIGURE CAPTIONS

584

585 **Fig. 1** Ivy density expressed as the number of stems per hectare  $\ge 0.5$ , 1 and 2.5 cm in the four 586 study plots

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588 Fig. 2 Ivy basal area per hectare (considering stems  $\ge 0.5$  cm) in the four study plots

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590 Fig. 3 Redundancy analysis ordination biplots. In the upper RDA (Fig. 3a), all 374 trees were 591 included, and tree age was not considered. In the lower RDA (Fig. 3b), only the 207 trees for which 592 tree age was estimated were included. Full-line arrows represent the biplot scores of tree variables: 593  $TR_d$  = tree diameter;  $TR_age$  = tree age;  $NE_n$  = number of neighbors in a circle of 10 m around 594 the tree; NE s d = sum of diameters of neighbors; NE ra = neighborhood ratio. Dotted arrowsrepresent the biplot scores of ivy descriptors:  $HE_n$  = number of ivy stems on the tree;  $HE_s_d$  = 595 596 sum of diameters;  $HE_ba$  = basal area;  $HE_d_M$  = diameter of the biggest stem;  $HE_cl$  = class of 597 occupancy. The dots represent tree species: Acer, Crataegus, Quercus, Robinia, Ulmus and Others 598 599 Fig. 4 Relationship between trunk circumference available to ivy stem (i.e. the ratio between host trunk circumference and sum of ivy stem diameters) and its mean ring width over the last 5 years. 600

Both axes are log-scaled. The straight line is the linear regression fitted to the data, with confidence

602 intervals (dotted lines)









