

AperTO - Archivio Istituzionale Open Access dell'Università di Torino

One-pot synthesis of Ag-modified SrTiO₃: synergistic effect of decoration and doping for highly efficient photocatalytic NO_x degradation under LED

This is the author's manuscript

Original Citation:

Availability:

This version is available <http://hdl.handle.net/2318/1946365> since 2025-06-02T18:21:27Z

Published version:

DOI:10.1016/j.jece.2023.110368

Terms of use:

Open Access

Anyone can freely access the full text of works made available as "Open Access". Works made available under a Creative Commons license can be used according to the terms and conditions of said license. Use of all other works requires consent of the right holder (author or publisher) if not exempted from copyright protection by the applicable law.

(Article begins on next page)

One-pot synthesis of Ag/SrTiO₃: synergistic effect of decoration and doping for highly efficient photocatalytic NO_x degradation under LED

Marcela Frias Ordoñez ^a, Giuseppina Cerrato ^b, Alessia Giordana ^b, Alessandro Di Michele ^c, Ermelinda Falletta^{a,d*},
and Claudia L. Bianchi ^{a,d}.

^a Department of Chemistry, University of Milan, Via Golgi 19, 20133 Milano, Italy

^b Department of Chemistry, University of Turin, Via Pietro Giuria 7, 10125 Turin, Italy

^c Department of Physics and Geology, University of Perugia, Via Pascoli 1, 06123 Perugia, Italy

^d Consorzio Interuniversitario Nazionale per la Scienza e Tecnologia dei Materiali (INSTM), via Giusti 9,
50121 Florence, Italy

Corresponding author: ermelinda.falletta@unimi.it

Abstract

Strontium titanate is an ideal cubic-structure perovskite oxide that has shown great potential, especially within hydrogen production and water pollutants degradation field. However, its photocatalytic activity is limited under UV irradiation due to its wide energy band gap (~ 3.2 eV). In order to overcome this limitation, modification strategies such as metal-doping or metal-decoration have emerged. By applying these two approaches, a rational designing of 5 wt.% Ag-SrTiO₃ materials was performed for extending photocatalytic efficiency towards the degradation of NO_x under LED light. As a result, the Ag-decorated SrTiO₃ synthesized by wet-impregnation method photodegraded by nearly 77% within 3h under LED light. Whilst, through an eco-friendly, simpler, and valuable one-pot synthesis it was fabricated a highly efficient Ag-SrTiO₃ material that achieved complete photodegradation of NO_x, exhibiting a photoactivity 4 times higher than bare SrTiO₃. This new material was doubly modified by Ag⁺ doping within the SrTiO₃ lattice and by Ag nanoparticles-decoration on the SrTiO₃ surface. The photocatalytic enhancement can be attributed to the synergetic

23 effect of Ag-doping and Ag-decoration on SrTiO₃, which contributed to a narrow band gap (2.80 eV), as well as the
24 formation of heterojunctions that promotes the separation of photogenerated charges, respectively. The morphological,
25 optical, structural, and physicochemical properties of the synthesized materials were fully investigated. The double
26 modified photocatalysts showed good stability during recycling tests, maintaining high performances after three
27 cycles. Eventually, active species were identified using various scavengers by trapping holes and radicals generated
28 during the photocatalytic degradation process.

29 **Keywords:** One-step procedure; Ag-modified SrTiO₃ materials; NO_x photodegradation; Degradation mechanism

30 1. Introduction

31 Air pollution has become the front stage of global environmental issues since it threatens human health, the
32 proliferation of diseases, and environmental disasters. The rapid growth of population, anthropogenic activities, and
33 over-consumption of natural resources have been the leading causes of the increment of harmful pollutants levels over
34 the last decades [1-3]. Some pollutants of significant public health concern include carbon monoxide, nitrogen oxides,
35 sulfur oxides, ammonia, and volatile organic compounds (VOCs)[4]. Nitrogen oxides (NO_x) are primary pollutants
36 directly linked to acid rains, photochemical smog, and respiratory problems. As a result, many countries and scientific
37 communities have striven to mitigate and control NO_x emissions to the atmosphere by using several techniques, such
38 as selective catalytic reduction (SCR) and non-selective catalytic reduction (NSCR), thermal and photocatalytic
39 decomposition [5-7]. The latter is carried out at ambient conditions taking advantage of natural or artificial light in the
40 presence of semiconductor materials. Therefore, photocatalysis is a green and cost-effective technology for the
41 abatement of NO_x [4,8].

42 Perovskite oxides belong to the third generation of photocatalysts, which have been extensively explored for energy
43 conversion and storage applications due to their unique properties[9]. Recently, their potential for environmental
44 remediation, particularly organic and inorganic pollutants degradation, has been studied[10]. Among them, strontium
45 titanate (SrTiO₃) stands out as a cubic-like structure perovskite with high thermal stability in comparison to TiO₂, which
46 makes it attractive for its application on solar cells, gas sensors, and capacitors[11]. It is also a semiconductor with a
47 wide energy band gap (3.2 eV), implying a high-efficiency photodegradation under UV irradiation [12]. However, UV

48 light is only 4% of the total solar spectrum [13]. Hence, as well as for TiO₂, modifying strategies for extending the
49 photocatalytic performance within the visible region have been required, such as forming heterojunctions or
50 incorporating defects into the lattice [14-15]. For instance, the design of a heterostructure between plasmonic
51 nanoparticles (Ag, Au, or Pt) and SrTiO₃ (STO) surface has been studied by Chen et al. [16]. The authors reported the
52 successful loading of Ag clusters (3-5 nm) over SrTiO₃ surface (~0.5 – 5 μm) via photoreduction method. Then, the
53 fabricated Ag-SrTiO₃ composites (0.1 wt.%, 0.5wt.%, 1wt.%, 2wt.%, and 5wt.% Ag) were tested towards photocatalytic
54 NO oxidation. The investigation determined 1%Ag-STO sample with the best photocatalytic performance, achieving
55 54% NO oxidation within 90 min under visible light irradiation. The improvement on the photocatalytic activity was
56 attributed to the localized surface plasmonic resonance properties (LSPR) of Ag nanoparticles and the Schottky barrier
57 formed at the interface of the metal and the semiconductor, promoting the separation of photogenerated charge carriers
58 and prolonging their lifetime [17-18]. On the other hand, Zhang et al. synthesized the optimal 0.5% Ag-SrTiO₃
59 nanocomposite by one-pot solvothermal method. This composite was characterized by Ag spheres of 10-20 nm
60 dispersed on the top of SrTiO₃ surface. However, the 0.5% Ag-SrTiO₃ material showed only a 30% of NO photocatalytic
61 degradation under visible light within 30 min[19].

62 Taking into account these considerations, this work aimed to employ two synthetic approaches for the fabrication of
63 Ag-SrTiO₃ materials and compared their photocatalytic efficiencies towards the NO_x abatement under LED light at
64 ambient conditions. A conventional decoration procedure of the SrTiO₃ surface with pre-synthesized Ag nanoparticles
65 (two-step process) and an innovative one-pot synthesis for both doping and decoration of SrTiO₃ were selected. The
66 one-pot approach results very promising due to its economical, environmental and operational simplicity advantages.
67 Moreover, the extreme conditions of heat treatment (900°C) were chosen since Ag offers a high melting point nearby
68 961°C and SrTiO₃ is characterized by high thermal stability[20-21]. Finally, the reusability of the most performing
69 material has been adequately investigated, showing high stability for up to three runs, and active species were
70 identified using various scavengers by trapping holes and radicals generated during the photocatalytic degradation
71 process.

72

73

74 **2. Materials and methods**

75 *2.1. Chemicals*

76 Strontium acetate ($C_4H_6O_4Sr$), titanium (IV)isopropoxide ($C_{12}H_{28}O_4Ti$, 97%), silver nitrate ($AgNO_3$, ACS reagent,
77 $\geq 99.0\%$), citric acid ($C_6H_8O_7$, ACS reagent, $\geq 99.5\%$), hydrogen peroxide (H_2O_2 , 30%), KNO_3 ($\geq 99.0\%$),
78 polyvinylpyrrolidone (PVP), HNO_3 (ACS reagent, 70%), and acetone (ACS reagent, $\geq 99.0\%$) were purchased from
79 Sigma Aldrich, Italia. All chemicals were used as received without further purification. An Ag^+ enriched solution was
80 purchased from Argor-Heraeus SA (Mendrisio, Switzerland), the solution composition is described elsewhere [22].

81 *2.2. Preparation of $SrTiO_3$ by a one-pot method*

82 Strontium titanate (STO) was synthesized by a procedure previously described elsewhere [23]. First, a solution A was
83 prepared by adding titanium (IV) isopropoxide (5.4 mmol) dropwise into deionized water (20 mL) under vigorously
84 stirring. Then, citric acid (10.8 mmol) was added to the previous suspension in a 2/1 molar ratio with respect to Ti-ions.
85 Afterwards, hydrogen peroxide 30% (2.6 mL) was added until the complete dissolution of titanium hydroxide formed
86 an orange-colored solution [24]. A solution B was obtained by dissolving a stoichiometric amount of strontium acetate
87 (5.4 mmol) in a minimum amount of deionized water (7 mL). Then, citric acid (10.8 mmol) was added in a molar ratio
88 2/1 with respect to Sr^{2+} ions. Both solutions were mixed at room temperature under vigorous stirring for 3 hours.
89 Subsequently, the solution was heated at $70^\circ C$. Before complete evaporation, the orange solution was transferred into
90 a crucible and dried at $75^\circ C$ for 12^h and at $120^\circ C$ for 12 h. The sample was calcined at $900^\circ C$ adopting the following
91 ramp ($0.5^\circ C/min$ from 25 to $250^\circ C$, time 1 min; $2.5^\circ C/min$ from 250 to $900^\circ C$, time 60 min).

92 This material was used as it is and for the preparation of Ag-decorated $SrTiO_3$, as described below (section 2.4.).

93

94 *2.3. Preparation of Ag- $SrTiO_3$ (Ag/STO) by one-pot approach*

95 Ag- $SrTiO_3$ materials were synthesized by the one-pot method described above [23] with the difference that the solution
96 B was obtained by dissolving silver ions (0.5 mmol) and strontium acetate (4.9 mmol) in a minimum amount of
97 deionized water (7 mL). Then, citric acid (10.8 mmol) was added in a molar ratio 2/1 with respect to the sum of ions of

98 Ag⁺ and Sr²⁺ ions. Both solutions were mixed at room temperature under vigorous stirring for 3 hours. Subsequently,
99 the solution was heated at 70°C. Before complete evaporation, the orange solution was transferred into a crucible and
100 dried at 75°C for 12^h and at 120°C for 12 h. The samples were calcined at 900°C adopting the following ramp (0.5°C/min
101 from 25 to 250°C, time 1 min; 2.5°C/min from 250 to 900°C, time 60 min).

102 Depending on the Ag source, the final materials were labeled Ag/STO1 (from the Ag⁺ enriched solution) and Ag/STO2
103 (from pure AgNO₃ from Sigma Aldrich).

104 2.4. Preparation of Ag-decorated SrTiO₃ (Ag@STO) by multistep approach

105 Ag nanoparticles (5 wt.%) were deposited onto STO surface (Ag@STO) via a wet impregnation method on the basis of
106 the literature [22]. The STO (44 mmol) synthesized as described above (section 2.2.) was dispersed in 10 mL acetone.
107 Then, 13 mL of a 30 g/L solution of Ag nanoparticles, previously synthesized by an electrochemical method starting
108 from the silver-enriched wastewater source, were added [22, 25]. The mixture was stirred at 40°C for 24h. Afterwards,
109 the temperature was raised up to 120°C under stirring for 1 h in order to evaporate the solvent. The remained product
110 was dried overnight at 100°C and further calcined at 900°C adopting the following ramp: 300°C for 2h, 600°C for 1h,
111 and 900°C for 1h (9°C/min).

112 2.5. Materials characterization

113 The crystal structure and phase composition of the synthesized samples were determined by X-ray diffraction (XRD)
114 employing a Rigaku-Miniflex-600 diffractometer using Cu-K α radiation ($\lambda= 1.541874 \text{ \AA}$). All the diffractograms were
115 collected between 10° - 80° (2 θ) at 3° min⁻¹ with a step of 0.02° (2 θ). The calculation of the crystallize size of the pure
116 and Ag-modified SrTiO₃ materials was estimated by the Scherrer equation (Eq. 1) using the peak at 46.4° that
117 corresponds to 200 reflection plane, where D is the crystallite size, k is the shape factor, λ is the X-ray wavelength of
118 radiation for Cu(k α), β_{hkl} is the full-width at half maximum (FWHM) at (hkl) peak, and θ_{hkl} is the diffraction angle.

$$119 \quad D = \frac{k\lambda}{\beta_{hkl}\cos\theta_{hkl}} \quad (\text{Eq. 1})$$

120 Specific surface area (SSA) and porosity (pore volume) were determined by N₂ adsorption/desorption isotherms
121 collected at -196°C on a Micrometrics Tristar II 3020 (Micromeritics) with Brunauer-Emmett-Teller (BET) equation (two

122 parameters, $0.05 < p/p_0 < 0.3$). Before the analysis, the sample was treated at 150°C for 4h under vacuum-adsorbed
123 foreign species. UV-Vis diffuse reflectance spectra were carried out at room temperature on a Perkin Elmer Lambda 35
124 UV-Vis Spectrometer equipped with an integrating sphere. The data were converted using the Kubelka-Munk formula.
125 X-ray photoelectron spectroscopy (XPS) measurements were performed on a M-probe apparatus equipped with an Al
126 $K\alpha$ source ($h\nu = 1486.6$ eV). Survey scans were measured between 0-1100 eV binding energy range with 5 eV energy
127 resolution. The morphology and composition of the samples were evaluated by field emission gun electron scanning
128 microscopy (FE-SEM, model LEO 1525 Zeiss), coupled with a Bruker Quantax EDX. The acceleration potential voltage
129 was maintained at 15 keV and measurements were carried out using AsB detector (Angle selective Backscattered
130 detector) and In-lens detector. The HRTEM images were collected using a JEOL 3010-UHR instrument (acceleration
131 potential: 300 kV; LaB6 filament).

132 *2.6 Photocatalytic tests*

133 The sample was prepared by suspending 50 mg of each photocatalyst in 5 mL isopropanol. An ultrasonic bath was
134 used to sonicate the suspension in order to obtain a uniform dispersion. Then, a thin film was deposited on a glass
135 plate (3.3 cm x 11.5 cm) by drop casting. After the solvent was evaporated, the glass plate with the photocatalyst was
136 placed inside a 20L Pyrex reactor. The reactor was filled in with NO_x gas at the initial concentration of 500 ± 50 ppb (by
137 feeding 0.625% of NO_2 and 0.125% of NO , diluted with air). The tests were carried out at room temperature for 3 hours
138 under LED irradiation (350 mA, 9-48V, 16.8W), yielding an intensity of 2900 lx on the photocatalyst surface. An ENVEA
139 AC32e chemiluminescence detector directly connected to the reactor measured the NO_x concentration at 30, 60, and
140 180 min. Photolysis tests indicate 10% of pollutant degradation after 3h of light irradiation. Lastly, the most performing
141 photocatalyst (Ag/STO2) was subjected to recycling tests (three tests under the same experimental conditions without
142 any post-treatment) in order to assess the stability.

143

144 *2.7 Scavenging experiments*

145 In order to investigate the reaction pathway, the photocatalytic tests were carried out in the presence of scavenger
146 agents using the procedure reported in the literature [26,28]. Potassium iodide (KI), potassium dichromate ($\text{K}_2\text{Cr}_2\text{O}_7$),

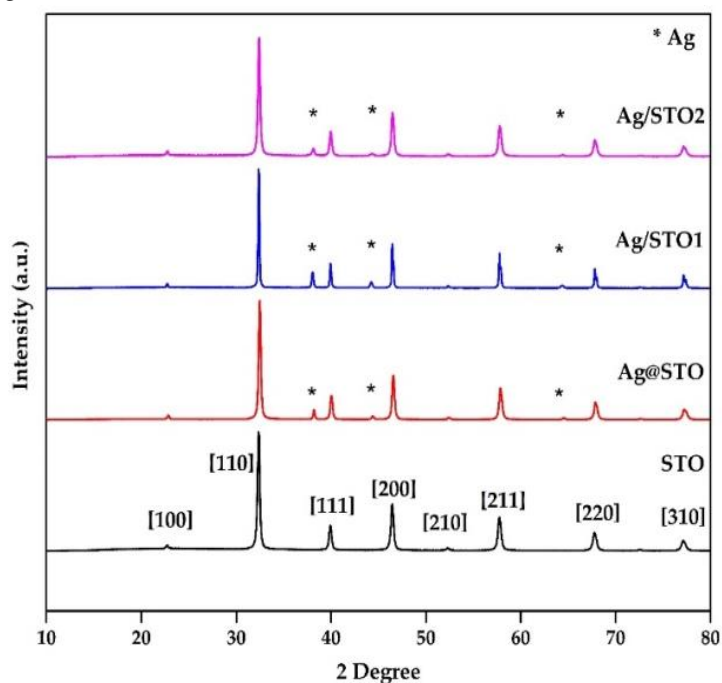
147 *tert*-butyl alcohol (TBA) and *p*-benzoquinone (PBQ) were employed as scavengers for h^+ , e^- , $\bullet OH$ and $O_2^{\bullet -}$, respectively.
148 Briefly, the photocatalyst (50 mg) was mixed with the proper radical scavenger (1 mmol) in deionized water (10 mL)
149 and ultrasonicated for 15 min. Afterwards, the suspensions were deposited onto glass plates, followed by a drying
150 process at 60°C until the water was completely evaporated. The plates were used to test the photocatalytic degradation
151 of NO_x under LED light, as described above.

152 3 Results and discussion

153 In order to investigate the effect of Ag-modification on STO, all the synthesized materials were extensively
154 characterized by several techniques, as described below.

155 3.1 Photocatalysts characterization

156 The crystal structures of the synthesized materials were analyzed using X-ray diffraction (XRD) and the obtained
157 results are showed in Figure 1.



165 **Figure 1.** XRD patterns of prepared samples.

166 The XRD data (Figure 1) evidenced the presence of peaks located at 22.7°, 32.4°, 39.9°, 46.4°, 52.3°, 57.7°, 67.7°, and 77.1°,
167 which are identified as (100), (110), (111), (200), (210), (211), (220), (310) planes of a cubic perovskite structure (space
168 group: P3m3) of strontium titanate. Compared to pristine STO, the samples Ag@STO, Ag/STO1, and Ag/STO2 exhibited

169 additional diffraction peaks located at around 38°, 44° and 64°, corresponding to metallic Ag on the surface of strontium
 170 titanate. More in detail, they can be indexed to the (111), (200) and (220) planes of cubic Ag phase. The average
 171 crystallite sizes were about 51 nm, 61 nm, 108 nm, and 54 nm for STO, Ag@STO, Ag/STO1 and Ag/STO2, respectively,
 172 as estimated from the half-peak width of the (200) diffraction peak (Table 1). It can be observed that the particle size
 173 was no dramatically affected by the synthesis procedure. However, the Ag/STO1 demonstrated a slight growth of the
 174 particle size. This could be associated to the presence of large amount of impurities (i.e. Cu, Fe, As, Ni, Pb, Zn) in the
 175 sample coming from the Ag⁺ enriched wastewater. In contrast with the scientific literature, from the XRD patterns of
 176 the three modified photocatalysts (Ag@STO, Ag/STO1 and Ag/STO2) is difficult to distinguish the characteristic shift
 177 associated to the introduction of a dopant into the host lattice (SrTiO₃), because all the diffractograms, also that of
 178 Ag@STO, exhibited a slight shift towards the higher angles.

179 **Table 1.** Average crystallite size, specific surface area (SSA), pore size, bandgap energy, rate constants of the
 180 prepared photocatalysts.

Photocatalyst	Average crystallite size (nm)	SSA (m ² /g)	Pore Size (nm)	Energy Bang (eV)	Rate Constant (min ⁻¹)
STO	51	16	33	3.2	0.002
Ag @STO	61	13	32	3.1	0.009
Ag/STO1	108	2	9	3.0	0.002
Ag/STO2	54	15	19	2.8	0.018

181

182

183 Since the photocatalytic performance is strongly linked to the surface properties of the material [29], the surface area
184 of all the photocatalysts was properly investigated (Figure 2 and Figure S1). According to the IUPAC classification, all
185 the materials exhibited isotherms type IV, characteristic of mesoporous materials [29]. The specific surface areas (SSA)
186 of all the samples are presented in Table 1. Among the Ag-SrTiO₃ samples, the Ag/STO2 photocatalyst showed the
187 biggest BET surface area (15 m²g⁻¹), which is beneficial for better adsorption of nitrogen oxide molecules and provides
188 a more significant number of reactive sites for enhancing photocatalytic activity [30-31]. On the other hand, the
189 Ag/STO1 sample exhibited the lowest BET surface area, suggesting that also in this case the large amount of impurities
190 present in the Ag-enriched wastewater might adversely affect the properties of the material. Moreover, this is in
191 accordance with the XRD results, showing very large crystallite size for this material. Lastly, the Ag@STO sample
192 showed a slight reduction in BET surface area (13 m²g⁻¹), which should be ascribed to a double calcination treatment
193 that the material was undergone[32-33].

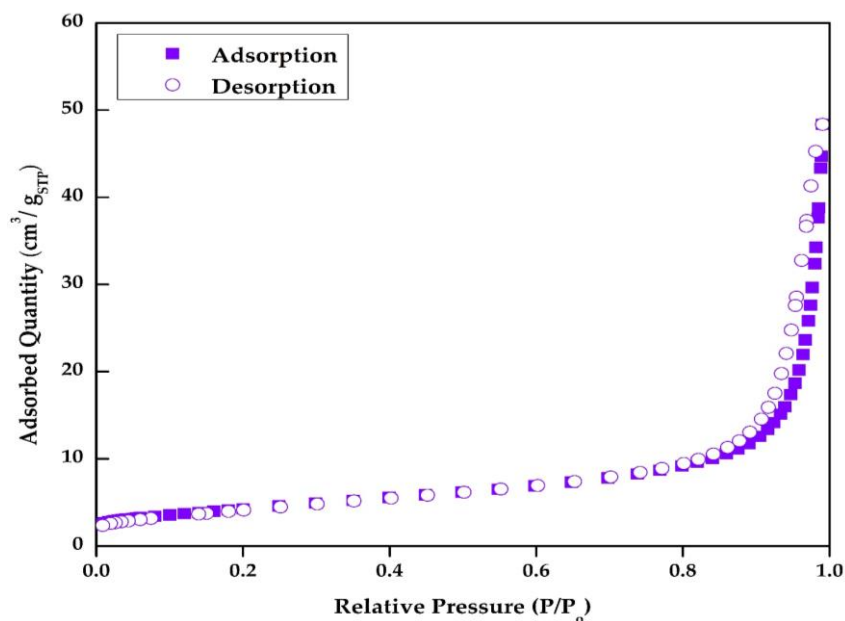
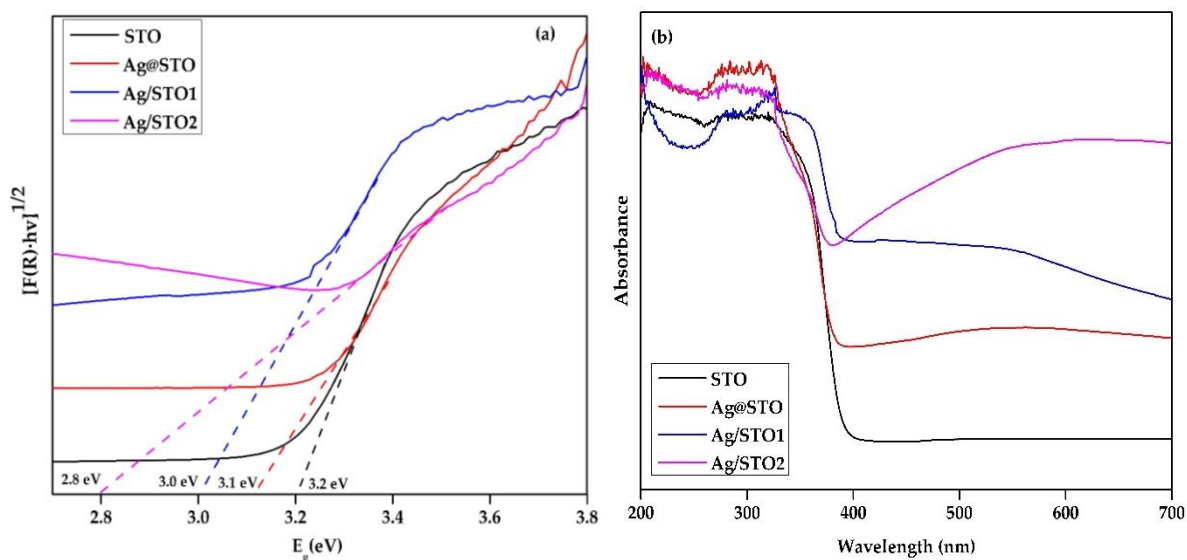


Figure 2. Nitrogen adsorption-desorption isotherm for Ag/STO2

194
195 The optical properties of the as-prepared samples were investigated by UV-Diffuse Reflectance Spectroscopy (UV-
196 DRS). The data were collected between 200-700 nm and transformed into absorption spectra by applying the Kubelka-
197 Munk function (Figure 3a) [34-35]. As a result, STO photocatalyst showed a wide energy band gap (Table 1), typical of
198 pristine SrTiO₃ that possesses the absorption edge in the UV light region (Figure 3b)[36-38]. After its surface decoration

199 with Ag nanoparticles (NPs), Ag@STO photocatalyst exhibits a higher light absorption within the visible region
 200 characteristic of the localized surface plasmon resonance (LSPR) effect. The latter suggests the promotion of electron-
 201 hole pair separation and hence a higher photocatalytic activity. Interestingly, by changing the synthetic procedure to
 202 one-step process the energy band gap and the absorbance got definitely modified. The Ag/STO1 and Ag/STO2
 203 displayed a nominal band gap of 3.0 and 2.8 eV, respectively. In particular, Ag/STO2 showed at the same time a greater
 204 absorbance in the visible region which, in accordance to the presence of Ag clusters, and low value of energy band gap,
 205 are highly beneficial for photocatalytic properties, exploiting the synergistic effect of both decoration and doping.
 206 Lastly, it has to be highlighted that the source of Ag might have negatively affected the response of the final Ag-SrTiO₃
 207 material, since there was no significant variation in the energy band gap.



208 **Figure 3.** (a) Tauc plots and (b) UV-Visible diffuse reflectance spectra of bare STO and Ag- SrTiO₃ samples.

209
 210
 211
 212
 213

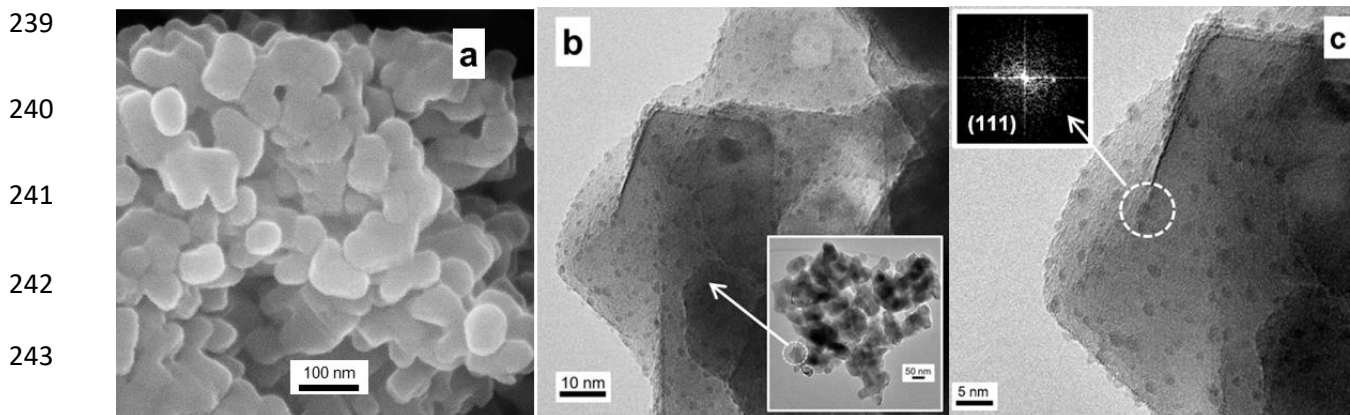
214 As it will be discussed below, Ag/STO₂ resulted to be the most performing photocatalyst, therefore, this material was
215 characterized more in depth.

216 To investigate the morphological features of Ag/STO₂, SEM and HRTEM analyses were performed (Figure 4). From
217 Figure 4a, it can be seen aggregated SrTiO₃ particles with not well-defined shape, and their size is in the 40-70 nm
218 range. The further HRTEM analysis was resorted to, in order to investigate the possible presence of Ag NPs, to confirm
219 the average size of STO particles and in some extent the general morphology as well. All the small nanoparticles can
220 be referred to Ag species on the basis of the EDX analysis (see the Figure S2), confirming the indications coming from
221 XRD investigations.

222 The Ag species were uniformly distributed onto the SrTiO₃ surface and exhibited a roundish morphology with an
223 average size of 2 nm. Their metallic character was also confirmed by the FFT calculations of the fringes (see the inset
224 to Figure 4c), in which the (1 1 1) family of metallic Ag is evidenced (ICDD file No. 004-0783). Moreover, the HRTEM
225 results provided evidence that SrTiO₃ particles displayed a cubic morphology. The amounts of Sr, Ti, Ag, and O
226 elements were confirmed being in the ratio expected from the theoretical stoichiometry. The Ag element was
227 homogeneous dispersed on SrTiO₃ and it was found in an atomic ratio of 0.09 respect to Sr. An excess of oxygen on the
228 Ag/STO₂ surface was observed, which can be explained by calcination treatment at 900°C in air.

229 Hence, bearing in mind these results and the UV-DRS experimental data, it can be implied that an amount of Ag would
230 be successfully introduced into SrTiO₃ lattice, responsible of the energy band gap modification, whereas the excess of
231 Ag would remain as metallic silver on SrTiO₃ surface, as demonstrated by the morphological characterization. This
232 synergistic effect of Ag-doping and Ag decoration will positively impact the photo response of Ag/STO₂.

233 Lastly, in order to compare the results obtained for Ag/STO₂ with those of Ag@STO, the latter was investigated by FE-
234 SEM. In Figure S3, roundish micro-sized particles (0.5 – 1.5 μm average size) were deposited on STO surface. It is clear
235 that these are Ag particles with a much larger dimension than those observed on Ag/STO₂ (less than 2 nm). Therefore,
236 by using a conventional decoration process (two-step), the Ag particles deposited onto SrTiO₃ surface exhibit these
237 features, that might negatively impact the surface area of the composite material and further limited its photocatalytic
238 performance[39].



245 **Figure 4.** (a) SEM image of Ag/STO2 and (b,c) TEM-HR-TEM images of Ag/STO2: the inset to Figure 4b represents a
246 low magnification overview, whereas the inset to Figure 4c refers to the FFT (simulating the electron diffraction) of the
247 evidenced nanoparticle.

248 To identify the surface chemical states of the Ag/STO2 photocatalyst, XPS analysis was performed[40]. The survey
249 spectrum (Figure S4) reports the typical photopeaks associated with Sr 3d (133.1 eV), Ti 2p (464.6 eV), O 1s (529.8 eV),
250 Ag 3d (368.0 eV), and C 1s (284.6 eV), the latter was ascribed to adventitious carbon during XPS measurement. The C
251 1s photopeaks are composed of three components (C-C, C-O, and C=O) with respective binding energies 284.6, 286.1,
252 and 287.5 eV (Figure S5)[41]. The high resolution (HR) Sr 3d spectrum (Figure 5a) showed the formation of a doublet
253 signal at 133.1 and 134.8 eV, associated with Sr 3d_{5/2} and Sr 3d_{3/2}, respectively. These values are in good agreement with
254 the reported values in the literature for the Sr²⁺ state in SrTiO₃ with a 1.7 eV splitting [42-43]. Regarding Ti, the two
255 main peaks of Ti 2p_{3/2} and Ti 2p_{1/2} centered at 458.7 and 464.6 eV, respectively, and a doublet splitting of 5.8 eV, verify
256 the existence of Ti⁴⁺ [42,44]. In Figure 5c, the characteristic binding energies of Ag 3d_{5/2} and Ag 3d_{3/2} of metallic silver
257 are at 368.0 and 374.0 eV for [45], whereas 3d Ag sub-peaks located at 368.9 and 374.9 eV were attributed to the presence
258 of silver ion, usually assigned to Ag₂O state [46]. Lastly, the O1s high-resolution region was described by two peaks at
259 529.0 and 531.7 eV, corresponding to lattice oxygen and hydroxyl groups, respectively [47-48].

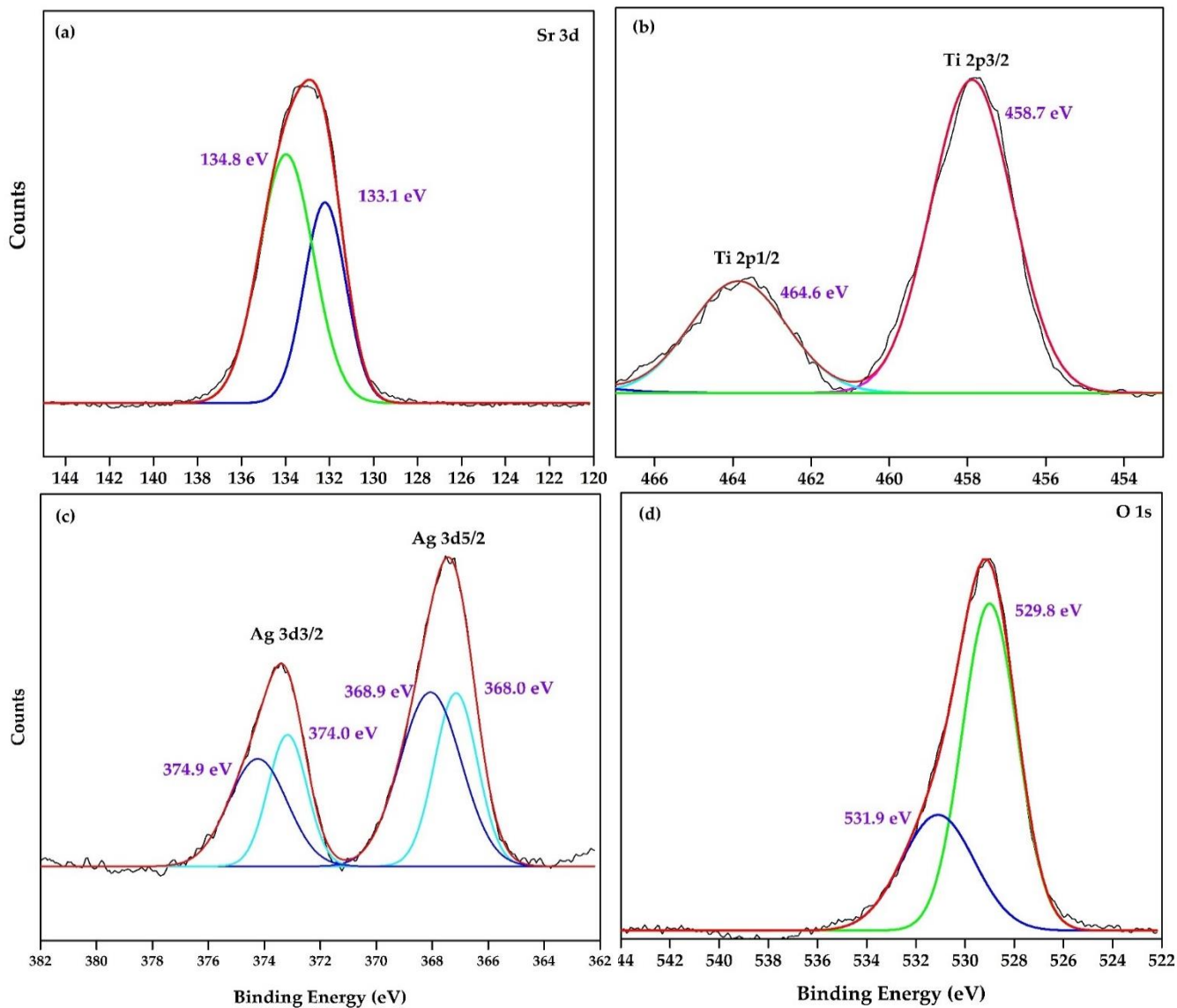


Figure 5. XPS HR spectra of (a) Sr 3d, (b) Ti 2p, (c) Ag 3d and (d) O 1s of Ag/STO2 photocatalyst.

260

261 *3.2 Photocatalytic activity*

262 The photocatalytic activity of the synthesized samples was tested towards the degradation of NO_x under LED light
 263 (400-700 nm) for 3 hours. For comparison, the photolysis of nitrogen oxide gases in the absence of any photocatalysts
 264 was investigated under the same experimental conditions. The results suggest that photolysis was negligible on the
 265 degradation process (Figure S6).

266 The photodegradation rates of nitrogen oxides as a function of the irradiation time over all the photocatalysts are
 267 plotted in Figure 6a. As expected, pristine STO sample was inactive under LED light due to their large energy bandgap

268 (3.2 eV)[49-50]. The STO surface's modification by Ag nanoparticles decoration (Ag@STO) leads to a significant increase
269 of the photocatalytic efficiency, achieving nearly 77% of NO_x photocatalytic abatement under LED light within 3 hours.
270 The enhancement was ascribed to the localized surface plasmon resonance (LSPR) effect that boosted the photoinduced
271 charge separation, despite the micrometric – sized of the Ag nanoparticles. Nonetheless, the small specific surface area
272 due to a double calcination treatment at 900C, and a large band energy value might have adversely inhibited the fully
273 photocatalytic degradation of nitrogen oxides under LED light after 180 min [22,29].

274 On the other hand, the performances of the materials synthesized by the one-pot procedure seem to be strongly
275 correlated to the purity of the Ag source. In fact, for Ag/STO1 and Ag/STO2 two different outcomes were obtained.
276 The Ag/STO1 photocatalyst showed a photocatalytic efficiency similar to bare STO. This result suggests that the large
277 amount of impurities present in the Ag⁺ enriched wastewater might play a deactivating role on the photocatalyst. On
278 the other hand, the Ag/STO2 photocatalyst synthesized with pure Ag demonstrated an extraordinary photocatalytic
279 efficiency greater than that obtained by Ag@STO. In fact, after 3 hours of LED irradiation, Ag/STO2 is capable to
280 completely degrade nitrogen oxides. This photocatalyst revealed a larger specific surface area (16 m²g⁻¹) in comparison
281 to Ag/STO1 [51-53]. Most importantly, it is evidenced the synergic effect of Ag⁺-doping and Ag-decoration, which is
282 noticed in a reduction in the band gap (2.80 eV) and the presence of small Ag clusters (average size 2 nm) that boosted
283 up the separation of charge carriers through LSPR effect [16,54]. The photocatalytic efficiency of Ag/STO2 was 4 times
284 greater than bare STO.

285 The corresponding first-order kinetics plot is shown in Figure 6b. It indicates that Ag/STO2 photocatalyst exhibited the
286 highest degradation rate (0.018 min⁻¹), which is almost 9 times greater than the bare SrTiO₃ and Ag/STO1 prepared in
287 similar experimental conditions (0.002 min⁻¹). Conversely, concerning the Ag-decorated photocatalyst (Ag@STO), the
288 rate (0.009 min⁻¹) is five times faster than bare STO, and it is two times slower than Ag/STO2. These results suggest that
289 the photocatalytic performance of the final Ag-SrTiO₃ materials is totally impacted by the synthesis method and the
290 source of Ag used. Nevertheless, one-pot solution synthesis carried out in the presence of pure Ag source stands out
291 as an attractive, feasible, one-step alternative for the fabrication of double-modified Ag-SrTiO₃ materials (Ag⁺-doped
292 and Ag-decorated) with extraordinary physicochemical, morphological, optical, and photocatalytic properties.

293 Lastly, the stability of the Ag/STO2 nanostructure was further evaluated by reusing the photocatalyst for degrading
294 NO_x gases under LED light. Three successive photocatalytic experimental runs were carried out under similar
295 experimental conditions. As shown in Figure 6c, the efficiency decreased by 5% only, demonstrating no significant loss
296 after three repeated runs. Therefore, the Ag/STO2 photocatalyst is a stable material for prolonged usage and practical
297 application for the photocatalytic degradation of NO_x under LED light.

298

299

300

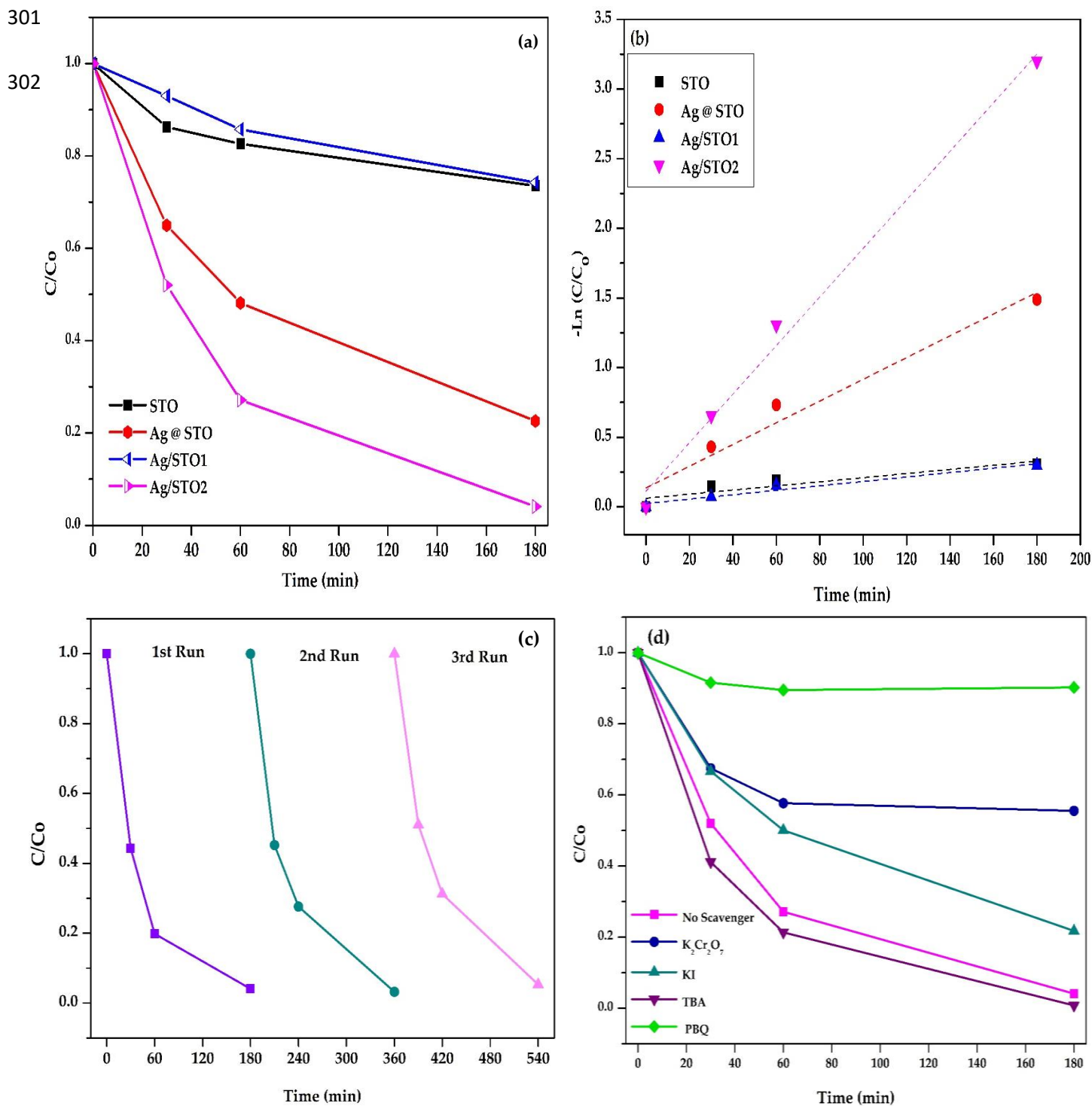


Figure 6. (a) NO_x photocatalytic removal efficiency C/C_0 as a function of time and (b) First-order kinetic plot of synthesized STO, Ag@STO, Ag/STO1, and Ag/STO2 materials; (c) reusability of Ag/STO2 photocatalyst for the NO_x degradation under LED irradiation; (d) NO_x photocatalytic removal efficiency under LED irradiation with the addition of potassium dichromate, potassium iodide, tert-butanol (TBA) and p-Benzoquinone as electrons, holes, •OH radicals and •O₂⁻ scavengers, respectively.

303 3.3 Mechanism of NO_x photodegradation

304 In order to explore the reaction mechanism of NO_x photodegradation, the scavenger tests were performed in order to
305 determine the active species mainly responsible for the degradation of nitrogen oxides under LED light. During the
306 photocatalytic process, the photogenerated holes (h⁺) and electrons (e⁻) can drive the redox reactions directly. However,
307 they can also trigger the production of other oxidizing species, such as hydroxyl radicals ([•]OH) and superoxide anion
308 radicals (O₂^{•-}), which indirectly start the degradation process of NO_x [5]. Potassium iodide, potassium dichromate, *tert*-
309 butyl alcohol (TBA), and p-benzoquinone (PBQ) were added to capture holes, electrons, [•]OH and O₂^{•-} radicals,
310 respectively [27,55,56]. As can be seen in Figure 6d, by the addition of TBA, the photocatalytic efficiency of Ag/STO2
311 demonstrated similar performance as in the absence of the scavenger. In contrast, the addition of PBQ thoroughly
312 depressed the photoactivity, which indicates that O₂^{•-} has a crucial role during the NO_x removal process. It is also clear
313 that the photodegradation of NO_x was also reduced by the addition of KI and K₂Cr₂O₇, showing a significant inhibition
314 but not a complete quenching of the reaction. These results confirm that photogenerated h⁺, e⁻ and O₂^{•-} play a key role
315 in photocatalytic NO_x degradation under LED light.

316 Similarly, the band edge positions of the photocatalysts are very important for the generation of active species during
317 the photocatalytic process. From the band gap shown in the Figure 3a and Table 1, it is possible to calculate the positions
318 of the minimum conduction band edge (CB) and maximum valence band edge (VB) of STO with respect to a normal
319 hydrogen electrode (NHE), according to the following equations:

$$320 \quad E_{CB} = X - E_e - \frac{1}{2}E_g \quad (\text{Eq. 2})$$

$$321 \quad E_{VB} = E_{CB} + E_g \quad (\text{Eq. 3})$$

322 X is the absolute electronegativity of the semiconductor, which can be calculated as the geometric mean of the absolute
323 electronegativity of the constituent elements of the semiconductor [57]. The E_g is the band gap energy of the
324 semiconductor, and E_e is the energy of the free electrons on the hydrogen scale (~ 4.5 eV). The electronegativity of Ag-
325 modified SrTiO₃ (X) was estimated to be 5.36 eV [58]. Therefore, the E_{CB} and E_{VB} for Ag/STO2 were estimated to be -
326 0.58 eV and 2.22 eV, respectively. These results indicate that the conduction band of Ag/STO2 is negative compared to
327 the reduction potential of O₂/O₂^{•-} (-0.33 eV). This suggests that it is easier for the photogenerated electrons to react with

328 O₂ to form O₂^{•-} placed in the conduction band of Ag/STO2. Correspondingly, the valence band of Ag/STO2 is positive
 329 compared to the redox potential for OH[•] (OH⁻/OH[•]) at 1.99 eV, indicating that the formation of OH[•] by the holes of
 330 SrTiO₃ is feasible [59-60]. The previous results are in line with the findings observed during the trapping experiments.
 331 The proposal reaction pathway is summarized in Figure 7.

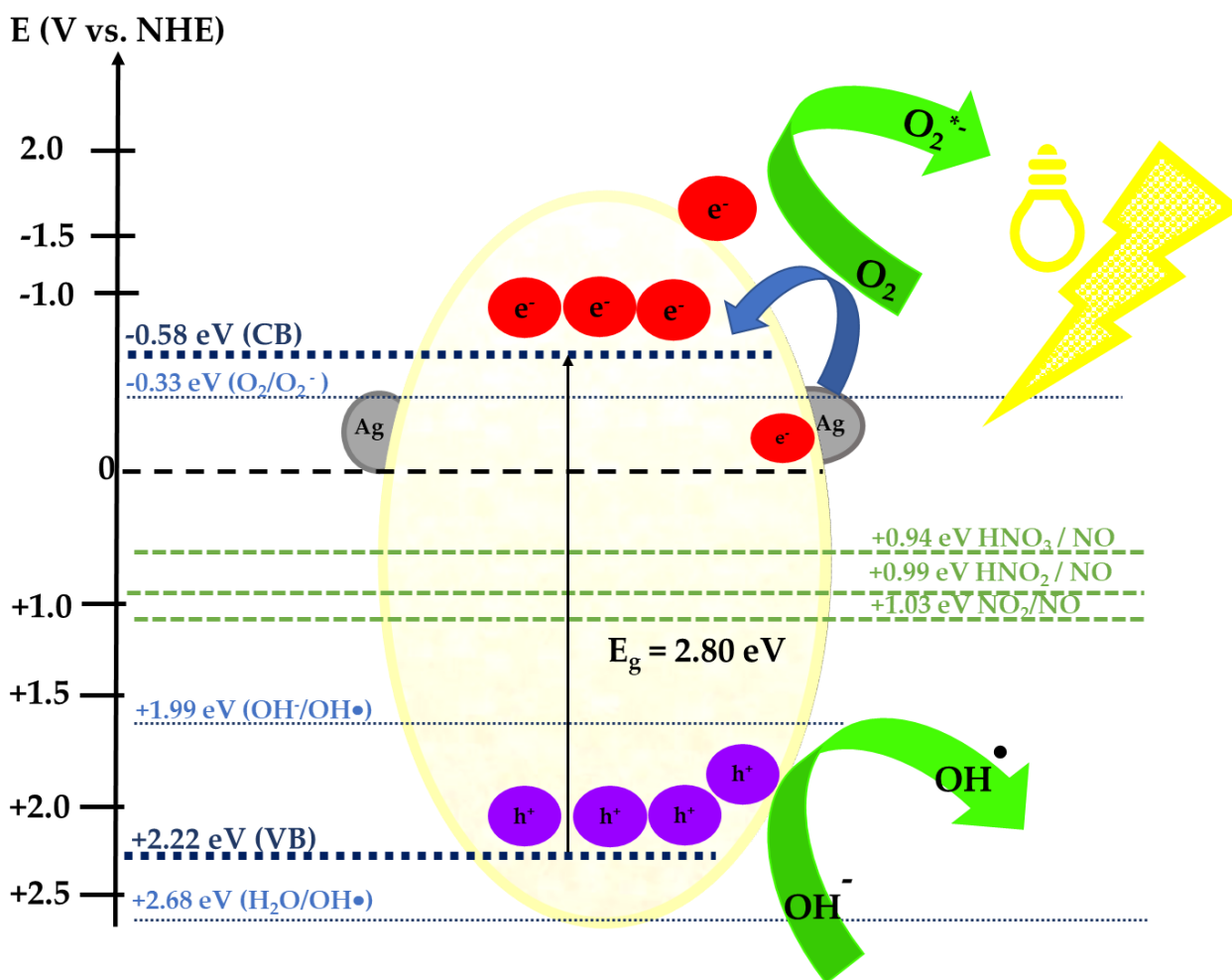
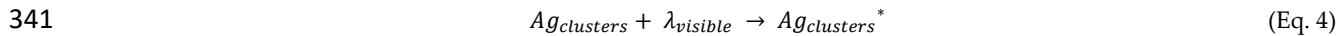


Figure 7. Scheme of photocatalytic degradation of NO_x under LED light.

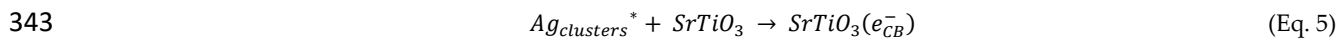
332
 333 When the Ag/STO2 photocatalyst is excited by LED light at the plasmonic frequency of Ag clusters, photoelectrons are
 334 produced and transferred to the conduction band of SrTiO₃ by overcoming the Schottky barrier at the
 335 metal/semiconductor interface (Eq. 4-5). Then, the hot electrons react with the adsorbed oxygen to produce superoxide
 336 anion and hydroxyl radicals (Eq. 6-8). The introduction of Ag into the SrTiO₃ lattice is reflected by the rise of the valence
 337 band with respect to the original SrTiO₃. The electrons present in the valence band can migrate to the E_f (fermi level) of

338 Ag nanoparticles, leaving behind holes that can react with hydroxyl groups to produce hydroxyl radicals (Eq. 9). Lastly,
339 the O_2^- and $\cdot OH$ radicals control the NO_x degradation process (Eq. 10-12)[19,27].

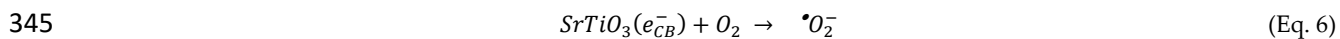
340



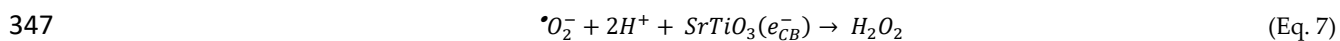
342



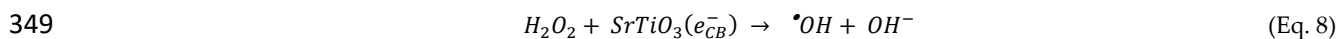
344



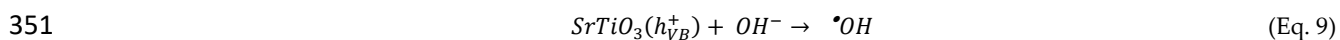
346



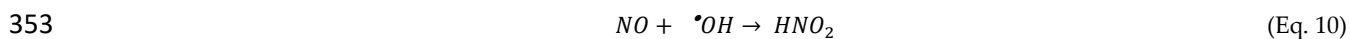
348



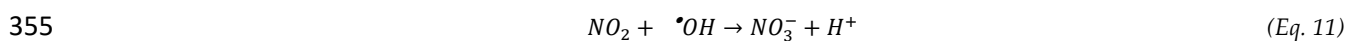
350



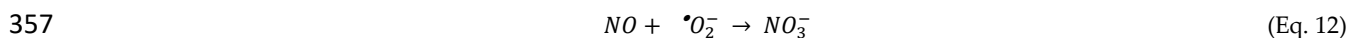
352



354



356



358

359 **Conclusions**

360 In this work, $SrTiO_3$ has been successfully modified using two strategies to extend its photocatalytic efficiency in the
361 degradation of NO_x under LED light: decoration by a two-steps approach, and doping and decoration by a one-step
362 method. By the two-step process, the $SrTiO_3$ surface was loaded with pre-synthesized Ag nanoparticles starting from
363 an Ag-enriched wastewater. The heterostructure material achieved nearly 77% NO_x removal after 3h under LED light.

364 Whilst, the one-pot process demonstrated to be a prominent method for the fabrication of an outstanding doubly Ag-
365 modified SrTiO₃ photocatalyst (Ag⁺ - doping and Ag-decorated). However, when the reaction was carried out by the
366 Ag⁺-enriched wastewater, the presence of large amount of impurities completely nullified the effect of the Ag, whereas
367 the photocatalyst prepared by a pure Ag source (Ag/STO2) showed excellent activity. Ag/STO2 photocatalyst exhibited
368 a narrow band gap and Ag metallic nanoparticles (size 2 nm) loaded on STO surface, besides a large specific surface
369 area (16 m²g⁻¹). The synergistic effect between Ag⁺-doping and Ag-decorated had essentially contributed to the
370 thorough NO_x photodegradation within 3h under LED light. Therefore, a one-pot synthesis is an eco-friendly, simpler,
371 and valuable process for the design of Ag-SrTiO₃ materials with double modification (within the SrTiO₃ lattice and
372 SrTiO₃ surface) that allows the extension and enhancement of photocatalytic efficiencies of SrTiO₃ towards NO_x
373 abatement under LED light.

374 Finally, the Ag/STO2 photocatalyst showed good photocatalytic activity without significant loss over the degradation
375 efficiencies of NO_x after three consecutive runs. Concerning the degradation mechanism, trapping experiments
376 confirmed that the photogenerated h⁺, e⁻ and O₂^{•-} species are the major reactive species responsible for NO_x
377 photodegradation in the Ag/STO2 surface.

378

379 CRediT authorship contribution statement

380 **Marcela Frias Ordoñez:** Methodology, Formal analysis, Investigation, Data Curation, Writing-Original Draft.

381 **Giuseppina Cerrato:** Formal analysis, Data curation, Writing- Review & Editing. **Alessia Giordana:** Formal analysis,

382 Data curation. **Alessandro Di Michele:** Formal analysis, Data curation. **Ermelinda Falletta:** Conceptualization,

383 Methodology, Data Curation, Validation, Supervision, Writing - Review & Editing, **Claudia L. Bianchi:**

384 Conceptualization, Data Curation, Supervision, Writing - Review & Editing, Project administration, Funding

385 acquisition.

386 Data availability

387 Data will be made available on request.

388 **Declaration of competing interest**

389 The authors declare that they have no known competing financial interests or personal relationships that could have
390 appeared to influence the work reported in this paper.

391 **Acknowledgements**

392 The authors thank Carolina Peverelli for materials preparations and photodegradation tests.

393 **Appendix A. Supplementary data**

394 **References**

- 395 [1] K. W. Tham, "Indoor air quality and its effects on humans—A review of challenges and developments in the
396 last 30 years," *Energy Build.*, vol. 130, pp. 637–650, 2016, doi: 10.1016/j.enbuild.2016.08.071.
- 397 [2] F. J. Kelly and J. C. Fussell, "Air pollution and public health: emerging hazards and improved understanding
398 of risk," *Environ. Geochem. Health*, vol. 37, no. 4, pp. 631–649, 2015, doi: 10.1007/s10653-015-9720-1.
- 399 [3] I. Khan, F. Hou, and H. P. Le, "The impact of natural resources, energy consumption, and population growth
400 on environmental quality: Fresh evidence from the United States of America," *Sci. Total Environ.*, vol. 754, p.
401 142222, 2021, doi: 10.1016/j.scitotenv.2020.142222.
- 402 [4] A. K. Priya, R. Suresh, P. S. Kumar, S. Rajendran, D. V. N. Vo, and M. Soto-Moscoco, "A review on recent
403 advancements in photocatalytic remediation for harmful inorganic and organic gases," *Chemosphere*, vol. 284,
404 no. May, p. 131344, 2021, doi: 10.1016/j.chemosphere.2021.131344.
- 405 [5] V. H. Nguyen *et al.*, "Photocatalytic NO_x abatement: Recent advances and emerging trends in the development
406 of photocatalysts," *J. Clean. Prod.*, vol. 270, no. x, p. 121912, 2020, doi: 10.1016/j.jclepro.2020.121912.
- 407 [6] T. H. Panigrahi, S. R. Sahoo, G. Murmu, D. Maity, and S. Saha, "Current challenges and developments of
408 inorganic/organic materials for the abatement of toxic nitrogen oxides (NO_x) – A critical review," *Prog. Solid
409 State Chem.*, vol. 68, no. 2, p. 100380, 2022, doi: 10.1016/j.progsolidstchem.2022.100380.
- 410 [7] A. Talaiekhosani, S. Rezanian, K. H. Kim, R. Sanaye, and A. M. Amani, "Recent advances in photocatalytic

- 411 removal of organic and inorganic pollutants in air," *J. Clean. Prod.*, vol. 278, p. 123895, 2021, doi:
412 10.1016/j.jclepro.2020.123895.
- 413 [8] M. S. S. Danish *et al.*, "Photocatalytic applications of metal oxides for sustainable environmental remediation,"
414 *Metals (Basel)*, vol. 11, no. 1, pp. 1–25, 2021, doi: 10.3390/met11010080.
- 415 [9] C. Sun, J. A. Alonso, and J. Bian, "Recent Advances in Perovskite-Type Oxides for Energy Conversion and
416 Storage Applications," *Adv. Energy Mater.*, vol. 11, no. 2, pp. 1–21, 2021, doi: 10.1002/aenm.202000459.
- 417 [10] R. Djellabi, M. F. Ordonez, F. Conte, E. Falletta, C. L. Bianchi, and I. Rossetti, "A review of advances in
418 multifunctional XTiO₃ perovskite-type oxides as piezo-photocatalysts for environmental remediation and
419 energy production," *J. Hazard. Mater.*, vol. 421, no. May 2021, p. 126792, 2022, doi:
420 10.1016/j.jhazmat.2021.126792.
- 421 [11] R. Sasikala *et al.*, "Strontium titanate perovskite embedded reduced graphene oxide photoanode for dye-
422 sensitized solar cell," *Opt. Mater. (Amst)*, vol. 136, no. January, p. 113464, 2023, doi:
423 10.1016/j.optmat.2023.113464.
- 424 [12] B. L. Phoon, C. W. Lai, J. C. Juan, P. L. Show, and G. T. Pan, "Recent developments of strontium titanate for
425 photocatalytic water splitting application," *Int. J. Hydrogen Energy*, vol. 44, no. 28, pp. 14316–14340, 2019, doi:
426 10.1016/j.ijhydene.2019.01.166.
- 427 [13] C. H. A. Tsang *et al.*, "Titanium oxide based photocatalytic materials development and their role of in the air
428 pollutants degradation: Overview and forecast," *Environ. Int.*, vol. 125, no. January, pp. 200–228, 2019, doi:
429 10.1016/j.envint.2019.01.015.
- 430 [14] R. Marschall, "Semiconductor composites: Strategies for enhancing charge carrier separation to improve
431 photocatalytic activity," *Adv. Funct. Mater.*, vol. 24, no. 17, pp. 2421–2440, 2014, doi: 10.1002/adfm.201303214.
- 432 [15] J. P. Jeon, D. H. Kweon, B. J. Jang, M. J. Ju, and J. B. Baek, "Enhancing the Photocatalytic Activity of TiO₂
433 Catalysts," *Adv. Sustain. Syst.*, vol. 4, no. 12, pp. 1–19, 2020, doi: 10.1002/adsu.202000197.
- 434 [16] Z. Chen, H. Yin, R. Wang, Y. Peng, C. You, and J. Li, "Efficient Electron Transfer by Plasmonic Silver in SrTiO₃

- 435 for Low-Concentration Photocatalytic NO Oxidation," *Environ. Sci. Technol.*, vol. 56, no. 6, pp. 3604–3612, 2022.
- 436 [17] N. Zhou, V. López-Puente, Q. Wang, L. Polavarapu, I. Pastoriza-Santos, and Q. H. Xu, "Plasmon-enhanced
437 light harvesting: Applications in enhanced photocatalysis, photodynamic therapy and photovoltaics," *RSC
438 Advances*, vol. 5, no. 37, pp. 29076–29097, 2015, doi: 10.1039/c5ra01819f.
- 439 [18] X. Zhang, Y. L. Chen, R.-S. Liu, and D. P. Tsai, "Plasmonic photocatalysis," *Reports Prog. Phys.*, vol. 76, no. 4, p.
440 046401, Apr. 2013, doi: 10.1088/0034-4885/76/4/046401.
- 441 [19] Q. Zhang, Y. Huang, L. Xu, J. J. Cao, W. Ho, and S. C. Lee, "Visible-Light-Active Plasmonic Ag-SrTiO₃
442 Nanocomposites for the Degradation of NO in Air with High Selectivity," *ACS Appl. Mater. Interfaces*, vol. 8,
443 no. 6, pp. 4165–4174, 2016, doi: 10.1021/acsami.5b11887.
- 444 [20] L. A. Navarro, X. Perpiñà, M. Vellvehi, D. Sánchez, and X. Jordà, "Sintering Process Analysis for Die-Attach of
445 Power Packages for High Temperature Applications based on Silver Nano-Particles Sintering Process Analysis
446 for Die - Attach of Power Packages for High Temperature Applications based on Silver Nano - Particles," no.
447 September, 2014.
- 448 [21] A. Fakharzadeh, R. Ebrahimi-Kahrizsangi, B. Nasiri-Tabrizi, and W. Jeffrey Basirun, "Effect of dopant loading
449 on the structural features of silver-doped hydroxyapatite obtained by mechanochemical method," *Ceram. Int.*,
450 vol. 43, no. 15, pp. 12588–12598, 2017, doi: 10.1016/j.ceramint.2017.06.136.
- 451 [22] M. Stucchi, D. Meroni, G. Safran, A. Villa, C. L. Bianchi, and L. Prati, "Noble Metal Promoted TiO₂ from Silver-
452 Waste Valorisation: Synergism between Ag and Au," *Catalysts*, vol. 12, no. 2, pp. 1–17, 2022, doi:
453 10.3390/catal12020235.
- 454 [23] L. Fabbrini *et al.*, "Sr_{1-x}Ag_xTiO_{3±δ} (x = 0, 0.1) perovskite-structured catalysts for the flameless combustion of
455 methane," *J. Catal.*, vol. 232, no. 2, pp. 247–256, 2005, doi: 10.1016/j.jcat.2005.03.012.
- 456 [24] M. Mori, M. Shibata, E. Kyuno, and S. Ito, "Reaction of hydrogen peroxide with titanium (IV) at different pH
457 values," *Bull. Chem. Soc. Jpn.*, vol. 29, no. 8, pp. 904–907, 1956.
- 458 [25] B. Yin, H. Ma, S. Wang, and S. Chen, "Electrochemical synthesis of silver nanoparticles under protection of

- 459 poly(N-vinylpyrrolidone)," *J. Phys. Chem. B*, vol. 107, no. 34, pp. 8898–8904, 2003, doi: 10.1021/jp0349031.
- 460 [26] G. Dong, L. Yang, F. Wang, L. Zang, and C. Wang, "Removal of nitric oxide through visible light photocatalysis
461 by g-C₃N₄ modified with perylene imides," *ACS Catal.*, vol. 6, no. 10, pp. 6511–6519, 2016.
- 462 [27] Q. Zhang *et al.*, "Perovskite LaFeO₃-SrTiO₃ composite for synergistically enhanced NO removal under visible
463 light excitation," *Appl. Catal. B Environ.*, vol. 204, pp. 346–357, 2017.
- 464 [28] S. Jin, G. Dong, J. Luo, F. Ma, and C. Wang, "Improved photocatalytic NO removal activity of SrTiO₃ by using
465 SrCO₃ as a new co-catalyst," *Appl. Catal. B Environ.*, vol. 227, pp. 24–34, 2018.
- 466 [29] J. Xiao *et al.*, "Simultaneously Tuning the Defects and Surface Properties of Ta₃N₅ Nanoparticles by Mg-Zr
467 Codoping for Significantly Accelerated Photocatalytic H₂ Evolution," *J. Am. Chem. Soc.*, vol. 143, no. 27, pp.
468 10059–10064, 2021, doi: 10.1021/jacs.1c04861.
- 469 [30] X. Guan, S. Zong, L. Tian, and M. Liu, "Efficient photocatalytic hydrogen production under visible-light
470 irradiation on SrTiO₃ without noble metal: dye-sensitization and earth-abundant cocatalyst modification,"
471 *Mater. Today Chem.*, vol. 26, p. 101018, 2022, doi: 10.1016/j.mtchem.2022.101018.
- 472 [31] S. Tonda, S. Kumar, O. Anjaneyulu, and V. Shanker, "Synthesis of Cr and La-codoped SrTiO₃ nanoparticles
473 for enhanced photocatalytic performance under sunlight irradiation," *Phys. Chem. Chem. Phys.*, vol. 16, no. 43,
474 pp. 23819–23828, 2014, doi: 10.1039/c4cp02963a.
- 475 [32] A. S. Al-Fatesh *et al.*, "Effect of pre-treatment and calcination temperature on Al₂O₃-ZrO₂ supported Ni-Co
476 catalysts for dry reforming of methane," *Int. J. Hydrogen Energy*, vol. 44, no. 39, pp. 21546–21558, 2019, doi:
477 10.1016/j.ijhydene.2019.06.085.
- 478 [33] S. Adhami *et al.*, "Influence of the specific surface area and silver crystallite size of mesoporous Ag/SrTiO₃ on
479 the selectivity enhancement of ethylene oxide production," *J. Chem. Technol. Biotechnol.*, vol. 94, no. 12, pp. 3839–
480 3849, 2019, doi: 10.1002/jctb.6182.
- 481 [34] P. Makuła, M. Pacia, and W. Macyk, "How To Correctly Determine the Band Gap Energy of Modified
482 Semiconductor Photocatalysts Based on UV-Vis Spectra," *J. Phys. Chem. Lett.*, vol. 9, no. 23, pp. 6814–6817, 2018,

- 483 doi: 10.1021/acs.jpcclett.8b02892.
- 484 [35] S. Landi, I. R. Segundo, E. Freitas, M. Vasilevskiy, J. Carneiro, and C. J. Tavares, "Use and misuse of the
485 Kubelka-Munk function to obtain the band gap energy from diffuse reflectance measurements," *Solid State
486 Commun.*, vol. 341, no. July 2021, pp. 1–7, 2022, doi: 10.1016/j.ssc.2021.114573.
- 487 [36] K. S. Kiran, B. S. Ashwath Narayana, and S. V Lokesh, "Enhanced photocatalytic activity of perovskite SrTiO₃
488 nanorods," *Solid State Technol.*, vol. 63, no. 4, pp. 1913–1920, 2020.
- 489 [37] U. Sulaeman, S. Yin, and T. Sato, "Solvothermal synthesis and photocatalytic properties of chromium-doped
490 SrTiO₃ nanoparticles," *Appl. Catal. B Environ.*, vol. 105, no. 1–2, pp. 206–210, 2011, doi:
491 10.1016/j.apcatb.2011.04.017.
- 492 [38] T. S. Jamil, H. A. Abbas, A. M. Youssief, E. S. Mansor, and F. F. Hammad, "The synthesis of nano-sized
493 undoped, Bi doped and Bi, Cu co-doped SrTiO₃ using two sol-gel methods to enhance the photocatalytic
494 performance for the degradation of dibutyl phthalate under visible light," *Comptes Rendus Chim.*, vol. 20, no. 2,
495 pp. 97–106, 2017, doi: 10.1016/j.crci.2016.05.022.
- 496 [39] L. F. Da Silva *et al.*, "An Understanding of the Photocatalytic Properties and Pollutant Degradation Mechanism
497 of SrTiO₃ Nanoparticles," *Photochem. Photobiol.*, vol. 92, no. 3, pp. 371–378, 2016, doi: 10.1111/php.12586.
- 498 [40] J. Lefebvre, F. Galli, C. L. Bianchi, G. S. Patience, and D. C. Boffito, "Experimental methods in chemical
499 engineering: X-ray photoelectron spectroscopy-XPS," *Can. J. Chem. Eng.*, vol. 97, no. 10, pp. 2588–2593, 2019,
500 doi: 10.1002/cjce.23530.
- 501 [41] M. G. Galloni, G. Cerrato, A. Giordana, E. Falletta, and C. L. Bianchi, "Sustainable Solar Light Photodegradation
502 of Diclofenac by Nano- and Micro-Sized SrTiO₃," *Catalysts*, vol. 12, no. 8, 2022, doi: 10.3390/catal12080804.
- 503 [42] M. A. Kafeshani, V. Mahdikhah, and S. Sheibani, "Facile preparation and modification of SrTiO₃ through Ni-
504 Cd co-doping as an efficient visible-light-driven photocatalyst," *Opt. Mater. (Amst.)*, vol. 133, no. September, p.
505 113080, 2022, doi: 10.1016/j.optmat.2022.113080.
- 506 [43] M. E. Pilleux, C. R. Grahmann, and V. M. Fuenzalida, "Hydrothermal Strontium Titanate Films on Titanium:

507 An XPS and AES Depth-Profiling Study," *J. Am. Ceram. Soc.*, vol. 77, no. 6, pp. 1601–1604, 1994, doi:
508 10.1111/j.1151-2916.1994.tb09763.x.

509 [44] U. Diebold and T. E. Madey, "TiO₂ by XPS," *Surf. Sci. Spectra*, vol. 4, no. 3, pp. 227–231, 1996.

510 [45] O. Schalm *et al.*, "A dataset of high-resolution synchrotron x-ray photoelectron spectra of tarnished silver-
511 copper surfaces before and after reduction with a remote helium plasma at atmospheric pressure," *Data Br.*,
512 vol. 35, p. 106872, 2021, doi: 10.1016/j.dib.2021.106872.

513 [46] S. W. Lee, S. Obregón, and V. Rodríguez-González, "The role of silver nanoparticles functionalized on TiO₂ for
514 photocatalytic disinfection of harmful algae," *RSC Adv.*, vol. 5, no. 55, pp. 44470–44475, 2015, doi:
515 10.1039/c5ra08313c.

516 [47] A. Bedon and A. Glisenti, "Chapter Developing Functionality in Perovskites from Abatement of Pollutants to
517 Sustainable Energy Conversion and Storage," 2020.

518 [48] J. Xie *et al.*, "Enhancing H₂ Evolution by Adjusting Oxygen Species at Perovskite SrTiO₃," *ACS Appl. Energy*
519 *Mater.*, vol. 5, no. 8, pp. 9559–9570, 2022.

520 [49] J. Wang, H. Li, H. Li, S. Yin, and T. Sato, "Preparation and photocatalytic activity of visible light-active sulfur
521 and nitrogen co-doped SrTiO₃," *Solid State Sci.*, vol. 11, no. 1, pp. 182–188, 2009, doi:
522 10.1016/j.solidstatesciences.2008.04.010.

523 [50] T. Ohno, T. Tsubota, Y. Nakamura, and K. Sayama, "Preparation of S, C cation-codoped SrTiO₃ and its
524 photocatalytic activity under visible light," *Appl. Catal. A Gen.*, vol. 288, no. 1–2, pp. 74–79, 2005, doi:
525 10.1016/j.apcata.2005.04.035.

526 [51] Y. Wu *et al.*, "Solvothermal fabrication of La-WO₃/SrTiO₃ heterojunction with high photocatalytic performance
527 under visible light irradiation," *Sol. Energy Mater. Sol. Cells*, vol. 176, pp. 230–238, 2018.

528 [52] S. Wan, M. Chen, M. Ou, and Q. Zhong, "Plasmonic Ag nanoparticles decorated SrTiO₃ nanocubes for
529 enhanced photocatalytic CO₂ reduction and H₂ evolution under visible light irradiation," *J. CO₂ Util.*, vol. 33,
530 pp. 357–364, 2019.

- 531 [53] X. Guan and L. Guo, "Cocatalytic effect of SrTiO₃ on Ag₃PO₄ toward enhanced photocatalytic water
532 oxidation," *Acs Catal.*, vol. 4, no. 9, pp. 3020–3026, 2014.
- 533 [54] H. A. Rafeie, R. M. Nor, M. S. Azmina, N. I. T. Ramli, and R. Mohamed, "Decoration of ZnO microstructures
534 with Ag nanoparticles enhanced the catalytic photodegradation of methylene blue dye," *J. Environ. Chem. Eng.*,
535 vol. 5, no. 4, pp. 3963–3972, 2017, doi: 10.1016/j.jece.2017.07.070.
- 536 [55] S. Jin, G. Dong, J. Luo, F. Ma, and C. Wang, "Improved photocatalytic NO removal activity of SrTiO₃ by using
537 SrCO₃ as a new co-catalyst," *Appl. Catal. B Environ.*, vol. 227, no. October 2017, pp. 24–34, 2018, doi:
538 10.1016/j.apcatb.2018.01.020.
- 539 [56] Y. Cui *et al.*, "Enhanced solar photocatalytic degradation of nitric oxide using graphene quantum dots/bismuth
540 tungstate composite catalysts," *Chem. Eng. J.*, vol. 420, no. P1, p. 129595, 2021, doi: 10.1016/j.cej.2021.129595.
- 541 [57] H. Dong, J. Sun, G. Chen, C. Li, Y. Hu, and C. Lv, "An advanced Ag-based photocatalyst Ag₂Ta₄O₁₁ with
542 outstanding activity, durability and universality for removing organic dyes," *Phys. Chem. Chem. Phys.*, vol. 16,
543 no. 43, pp. 23915–23921, 2014, doi: 10.1039/c4cp03494e.
- 544 [58] H. Bentour, M. El Yadari, A. El Kenz, and A. Benyoussef, "DFT study of electronic and optical properties of (S–
545 Mn) co-doped SrTiO₃ for enhanced photocatalytic hydrogen production," *Solid State Commun.*, vol. 312, no.
546 February, 2020, doi: 10.1016/j.ssc.2020.113893.
- 547 [59] V. Lalan, V. P. Mahadevan Pillai, and K. G. Gopchandran, "Enhanced electron transfer due to rGO makes Ag–
548 CaTiO₃@rGO a promising plasmonic photocatalyst," *J. Sci. Adv. Mater. Devices*, vol. 7, no. 3, p. 100468, 2022,
549 doi: 10.1016/j.jsamd.2022.100468.
- 550 [60] Z. H. Shah, Y. Ge, W. Ye, X. J. Lin, S. Zhang, and R. Lu, "Visible light activation of SrTiO₃ by loading Ag/AgX
551 (X = Cl, Br) for highly efficient plasmon-enhanced photocatalysis," *Mater. Chem. Phys.*, vol. 198, pp. 73–82, 2017,
552 doi: 10.1016/j.matchemphys.2017.05.002.

553