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The role of fire disturbances, human activities and climate change for long-term forest dynamics in upper-montane forests of the central Dinaric Alps

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1 **The role of fire disturbances, human activities and climate change for long-term forest**
2 **dynamics in upper-montane forests of the central Dinaric Alps**

3

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30 Abstract

31 We present the first high-resolution Holocene pollen, plant-macrofossil, and charcoal records from the
32 upper-montane zone in the central Dinaric Alps. Drawing on these new records from well-dated
33 lacustrine sediments of ~~Zminje jezero~~Zminje Jezero (ca. 1500 m a.s.l.; Montenegro) and on independent
34 chironomid-inferred summer temperatures, we explore long-term ecosystem responses to variations in
35 climate, fire disturbances and land use, as well as legacy effects of past environmental changes.

36 A mixed spruce-fir (~~*Picea abies*—*Abies alba*~~) forest established in the upper-montane zone around 9500
37 cal BP, and beech (*Fagus sylvatica*) became co-dominant with the two conifers after 5000 cal BP.
38 Prehistoric land-use pressure was overall remarkably low, but increased since 2000 cal BP and was
39 highest after the Middle Ages.

40 We found a significant positive relationship between biomass burning and summer temperature,
41 indicating that fires were mostly climate driven. *P. abies*~~Spruce~~ was insensitive to summer temperature,
42 biomass burning and human impact, which supports the view that spruce forests may not be significantly
43 impacted by fire. In contrast, *A. alba*~~fir~~ and other disturbance-sensitive trees (*Tilia*, *Ulmus*, *Fraxinus*
44 *excelsior*-type) show significant negative responses to land-use pressure and positive responses to
45 summer temperature. This supports the notion that these species may be well-adapted to warmer-than
46 present summer temperatures and that their populations declined in recent millennia due to land-use
47 activities. Conversely, *F. sylvatica*~~beech~~ was sensitive to summer temperatures but was promoted by
48 low biomass burning, indicating that its expansion in the spruce-fir dominated forest was enhanced by
49 the onset of cooler and possibly also moister climatic conditions as well as by fire disturbances.

50

51 Keywords

52 Holocene, vegetation dynamics, mountain forests, fire history, vegetation-fire interactions, ~~Dinaric~~
53 ~~Alps~~Montenegro

54 **Introduction**

55 Assessing the responses of European mountain forests to changing environmental conditions is crucial
56 to develop adaptation and management strategies to possible ecosystem shifts that may cause the loss
57 of important ecosystem services, such as risk prevention (from avalanches, landslides or rockfall),
58 recreation, and the maintenance and promotion of biodiversity (Klopčič et al., 2017). Such assessments
59 are particularly important given that there is strong support for the hypothesis that climate change could
60 markedly modify disturbance regimes, with a likely increase of some disturbances (e.g. fire) in a
61 warming world (Seidl et al., 2017).

62 However, understanding natural ecological processes is challenging as processes often play out over
63 long time scales (centuries, millennia), particularly when long-lived species such as trees are involved.
64 The long time horizon of palaeoecological records permits to study ecosystem responses under
65 substantially different environmental conditions than the present ones, including temperatures analogous
66 to those predicted by future climate changes (approximately 1-2°C higher than present; Carter et al.,
67 2018; Morales-Molino et al., 2021; Samartin et al., 2017). Moreover, current ecosystems and ecological
68 processes often carry legacy effects of past environmental changes and anthropogenic impacts (Cagliero
69 et al., 2022; Feurdean et al., 2009; Grindean et al., 2019; Morales-Molino et al., 2022), which have often
70 altered species-environment relationships (Tinner et al., 2013). Thus, taking into account long-term
71 records may be important to gain more accurate environmental-change response assessments, especially
72 in Europe, where anthropogenic manipulation of nature started several millennia ago (Birks and Tinner,
73 2016).

74 The long-term history of mountain ecosystems varies across regions and is arguably contingent on
75 patterns of human settlement, land use, and socioeconomic development. In this context, the central
76 Dinaric Alps are an interesting area as its land-use history may differ from those of other European
77 mountain regions. There is little doubt that the Balkan Peninsula acted as a land bridge from the Near
78 East, across Asia Minor and towards Central and Western Europe (Forenbaher et al., 2013). While it
79 seems that the Neolithic spread mainly across the fertile plains along the Danube valley in the east
80 (Starčevo culture) and along the Adriatic coastal strip in the west (Impresso culture) around the 6th
81 millennium BCE (8000-7000 cal BP; Borić et al., 2019), there is evidence to suggest that the

82 mountainous hinterland of the central Dinaric Alps remained an agricultural frontier zone, eventually
83 even until the Late Neolithic (Forenbaier and Miracle, 2005). The frontier-zone hypothesis is consistent
84 with modelled anthropogenic deforestation (Kaplan et al., 2009), which suggests that during the past
85 3000 years relatively low forest clearance occurred when compared to other regions of Europe, and that
86 a large fraction of usable land in the region was not heavily exploited until the Middle Ages (600-1460
87 CE; 1350-490 cal BP). By contrast, charcoal production, mining activities, and grazing likely occurred
88 in the interior area of the Dinaric Alps since the Bronze Age (3500-600 BCE; ~~545500-2600-2550~~ cal
89 BP), and intensified during the Iron Age (600-200 BCE; ~~255600~~-2150 cal BP) and the Ottoman Empire
90 (1460-1800 CE; 490-150 cal BP) (Kranjc, 2009; Longman et al., 2018). However, it is not clear if these
91 activities were widespread across the mountain region, or whether human activities rather concentrated
92 in some areas. In the latter case, large forested regions in the interior range may have remained relatively
93 intact until the present (Nagel et al., 2017) or managed with low intensity silvicultural systems during
94 recent centuries (Boncina, 2011). Low historical forest clearance in the montane belt may have preserved
95 fir (*Abies alba* Mill.), beech (*Fagus sylvatica* L.), and spruce (*Picea abies* (L.) Karst.) old-growth forests
96 (Cagliero et al., 2022; Motta et al., 2011; Sabatini et al., 2018). As these tree species are widespread and
97 often form mixed forests in European mountains (Hilmers et al., 2019), ~~knowledge of~~ exploring their
98 long-term responses under low land-use pressure may be useful to improve knowledge on the role of
99 disturbances by fire and climate. However, the impact of Neolithic and of more recent cultures on the
100 interior mountain region is still weakly constrained. In contrast to other European mountain regions, the
101 area is both archeologically and palaeoecologically under-documented (Borić et al., 2019; Finsinger et
102 al., 2017; Vander Linden et al., 2014).

103 Thus, to provide new insights into the Holocene vegetation dynamics and into vegetation responses to
104 disturbances by fire and to changes in climate and land use we studied sediments from ~~Zminje~~
105 ~~jezero~~ Zminje Jezero (Fig. 1), a lake at 1535 m a.s.l., thus in the montane belt of the central Dinaric Alps.
106 Our main aims are to (1) assess the long-term Holocene vegetation dynamics of mixed montane forests,
107 (2) characterize the Holocene land-use and fire histories, (3) investigate the influence of the main drivers
108 of fire dynamics (climate and human impact), and (4) track responses of the dominant montane trees (A

109 *alba*, *F. sylvatica*, *P. abies*~~fir, beech, spruce~~, as well as of other disturbance-sensitive trees such as
110 *Ulmus*, *Tilia*, and *Fraxinus excelsior*) to variations in climate, fire and human impact.

111

112

113 **Material and methods**

114

115 *Study area and study site*

116 In the Dinaric Alps, vegetation is influenced by the interaction of mountainous reliefs, pedological
117 conditions, and proximity to the Adriatic Ssea (Horvat et al., 1974; Nagel et al., 2017). Mediterranean
118 evergreen ~~hard-leaf~~sclerophyllous vegetation dominated by *Pinus halepensis* Mill., *Quercus -ilex* L. and
119 *Olea europaea* L. occurs near the coast. Mixed ~~sub-Mediterranean~~submediterranean deciduous
120 woodlands with *Carpinus betulus* L., *Fraxinus ornus* L., deciduous *Quercus* L., *Ostrya carpinifolia*
121 Scop., *Carpinus orientalis* Mill., and *Castanea sativa* Mill. occur up to c. 700 m asl. The valley
122 floodplains are commonly occupied by *Alnus glutinosa* (L.) Gaertn., *Quercus robur* L., *Fraxinus* sp.
123 and *Ulmus minor* Mill.. In the montane belt (from 700 to 2000 m asl), forests are dominated by *Pinus*
124 *nigra* J.F. Arnold, *F. sylvatica*, and mixed beech-fir-spruce forests. Other species occurring in the
125 montane belt include *Pinus sylvestris* L., *O. carpinifolia*, *Acer* L., *Ulmus* L., *Tilia* L., and *Fraxinus*
126 *excelsior* L.. The subalpine vegetation belt is generally above c. 2000 m asl and is dominated by *Pinus*
127 *mugo* Turra, *Juniperus communis* L., and *Alnus viridis* (Chaix) DC.

128 ~~Zminje jezero~~Zminje Jezero (Snake lake; 43°09'21" N 19°04'14" E; 1535 m asl) is a small 1.2-ha large
129 and 9.5-m deep glacial-origin lake (Fig. 1) whose water is drained into the Crno Jezero (1460 m asl) by
130 a small stream. The lake is located on the northwest part of a high karst plateau in the Durmitor massif
131 (Annys et al., 2014) in a valley adjacent to the one currently hosting the last surviving glacier in
132 Montenegro, Debeli namet (Hughes et al., 2011). Given its elevation, ~~Zminje jezero~~Zminje Jezero ~~may~~
133 ~~be~~is within the limits of terminal moraines that were deposited ~~either~~ before ~~or~~ during the Lateglacial
134 Interstadial (Hughes et al., 2011).

135 The vegetation in the surroundings of the lake is dominated by a dense spruce-fir forest (Fig. 1) with
136 lesser amounts of beech, *Acer* sp., *Sorbus aucuparia* L., *Rhamnus fallax* Boiss., and *Vaccinium myrtillus*

137 L.. However, *P. sylvestris*, *Pinus nigra* subsp. *nigra*, *Betula ~~alba~~ pendula* ~~L.~~ Roth, and *Pinus heldreichii*
138 H. Christ are also well represented in the Durmitor massif. To the east, there is a town (Žabljak; 1450
139 m asl) with pastures and agriculture lands. The lake shores host a rich community of wetland plants,
140 including *Molinia caerulea* (L.) Moench, *Comarum palustre* L., *Carex* sp., *Sphagnum* sp., *Parnassia*
141 *palustris* L., *Filipendula ulmaria* (L.) Maxim., and *Eriophorum vaginatum* L.. The current tree line is at
142 c. 1950 m asl and is mainly formed by spruce with beech (Bui, 1975). Above the tree line are *P. mugo*
143 and *J. communis* shrublands, alpine meadows, and rocks.

144 Climate is continental with a mean annual air temperature of 5.1°C and a mean annual precipitation of
145 c. 1450 mm at Žabljak (Annys et al., 2014). In the northern and central parts of the mountain range,
146 bedrock consists of Mesozoic sedimentary rocks and thick Middle and Upper Triassic and Upper
147 Jurassic limestones, while in the southern part Upper Cretaceous flysch are predominant (Mirković,
148 1985).

149 ~~Zminje jezero~~ Zminje Jezero is located in a zone of special protection of the Durmitor National Park,
150 where only cutting of naturally collapsed trees (so-called “sanitary cuttings”), clearing of forest roads,
151 and tourism are allowed (Srdanović and Pavić, 2013; UNESCO and IUCN, 2018). Before the
152 establishment of the National Park in 1952 CE, the area was informally protected since 1907 CE.

153

154 *Field work and sediment-core correlation*

155 We studied the lake floor with an echo sounder connected to a Garmin GPS device (Fig. 1d) and
156 collected two overlapping sediment cores (sections 1-m long and 6 cm in diameter) from the deepest
157 part of the lake at a water depth of 9.3 m using a modified Livingstone piston corer and an UWITEC
158 gravity corer. All core sections were split longitudinally and photographed at ISEM (University of
159 Montpellier) to visually describe the sediments and align the sections with the aid of Corelyzer v2.1.1
160 (CSDCO/LacCore-University of Minnesota). To assess sediment composition, we used an X-ray
161 fluorescence (XRF) AVAATECH core scanner at 5 mm resolution (with the two following settings: 10
162 kV/0.09 mA for 15 s and 30 kV/0.09 mA for 30 s), and then normalised the values using the centre-log-
163 ratio method (CLR, Weltje et al., 2015). We correlated the sections based on marker layers and on

164 selected XRF records. Thereafter, one core half was cut into 1-cm slices and samples were stored in zip-
165 lock bags at 4°C.

166

167 *Chronology*

168 We modelled the depth-age relationship (Fig. 2) with RBacon v2.4.1 (Blaauw and Christen, 2011) using
169 41 control points (Table 1 [and Table S1](#)), including 19 AMS ¹⁴C dates from terrestrial plant macrofossils
170 calibrated using the IntCal20 dataset (Reimer et al., 2020), one pollen-inferred age (onset of the
171 Holocene; Giesecke et al., 2014), and 21 control points derived from a chronology of short-lived
172 radionuclides (²¹⁰Pb and ¹³⁷Cs; Fig. S1) that was built with the constant flux constant sedimentation
173 (CFCS) using the serac package (Brueel and Sabatier, 2020). We excised visually identified event-deposit
174 layers matching peaks of K, Ti, Rb, Zr, and Sr (Fig. 2), as they most probably represent sediment
175 deposition that occurred over very short time spans (Finsinger et al., 2021; Heiri et al., 2003).

176

177 *Pollen, spores and plant-macrofossils analyses*

178 We processed 78 ~~1~~¹one-cm⁻³ samples for pollen analysis following standard physical and chemical
179 treatments, including treatments with HCl, KOH, sieving with a 500-µm mesh and decanting, HF,
180 ~~a~~^aAcetolysis, and mounted the ~~Fuchsin~~^{fuchsin}-stained residues on slides with glycerol. To calculate
181 influxes (# cm⁻² yr⁻¹), we added *Lycopodium* tablets (Stockmarr, 1971). We identified and counted
182 pollen, stomata and the dung-fungi *Sporormiella* spore using determination keys and photographic
183 atlases (Beug, 2004; Cagliero et al., 2022; Davis, 1987; Finsinger and Tinner, 2005; Moore et al., 1998;
184 Reille, 1992) as well as the pollen and stomata reference collections at ISEM (University of
185 Montpellier). We counted at least 300 terrestrial pollen grains at ~~×400x~~^{×400x} magnification and calculated
186 percentages relative to the terrestrial pollen sum, which excludes pollen of obligate aquatic plants, [spores](#)
187 [of ferns and fungi](#).

188 The pollen diagram was divided into pollen assemblage zones using optimal partitioning by sums-of-
189 squares with square-root-transformed proportions of pollen types included in the terrestrial pollen sum,
190 and exclusion of rare taxa (abundance <5%; Birks and Gordon 1985). The number of statistically

191 significant zones was determined by comparison with the broken-stick model (Bennett, 1996) with
192 Psimpoll v4.26 (Bennett, 2008).

193 For plant-macrofossils analyses, we processed 203 samples. We measured sample volumes (~~e-~~5.6±1.0
194 cm³) by water displacement and then sieved them with a 100-µm mesh sieve. We identified and counted
195 macrofossils with a stereomicroscope at ~~×~~7.5-60x magnification using atlases (Birks, 2017; Katz et al.,
196 1965; Schoch et al., 1988) and the reference collection at ISEM (University of Montpellier) and stored
197 them at -18°C in Milli-Q water. Plant-macrofossil abundances are expressed as concentrations (# cm⁻³).
198 To assess land-use pressure, we used the abundances of *Secale cereale* pollen (a primary anthropogenic
199 indicator), secondary indicators (as the sum of adventives and apophytes), and cultural indicators (as the
200 sum of Cerealia-type and *Plantago lanceolata*-type pollen) (Behre, 1981; Tinner et al., 2003). These
201 indexes include pollen of cultivated plants, ruderals, and plants of meadows (Deza-Araujo et al., 2020).
202 *Veratrum album* was added to the apophyte category, as it is avoided by cattle due to its toxicity
203 (Spiegelberger et al., 2006). We also assessed land-use pressure in the wider region using the OJC index
204 (Mercuri et al., 2013), which is based on the abundance of woody crops (*Olea europaea* L., *Juglans*
205 *regia* L., and *Castanea sativa* Mill.) that grow at lower elevations and along the coast of the Adriatic
206 Sea. Despite their presence in uncultivated areas, an increase in their abundances is often due to broad-
207 scale cultivation (Deza-Araujo et al., 2020).

208

209 *Charcoal analyses*

210 We counted microscopic charcoal particles in 42 pollen slides spread along the sediment core (Finsinger
211 and Tinner, 2005; Tinner and Hu, 2003) to obtain a low-resolution record of regional fire activity. For
212 macroscopic charcoal analyses, we took contiguous 1-cm³ samples and treated them without the use of
213 an orbital shaker with a 5% NaPO₃ solution overnight and then with a 15% H₂O₂ solution for 24h
214 (Schlachter and Horn, 2010). Subsequently, samples were sieved through a 150-µm mesh (Higuera et
215 al., 2005) and charcoal particles were identified and counted under a stereomicroscope at ~~×~~7.5-60x
216 magnification. The area of charcoal particles was not estimated since it is generally highly correlated to
217 charcoal counts (Finsinger et al., 2014; Tinner et al., 1998). To account for variations in sedimentation
218 rate (Whitlock and Larsen, 2001), both microscopic and macroscopic charcoal counts were converted

219 to charcoal-accumulation rates (influx; # cm⁻² yr⁻¹), hereafter denoted mCHAR and MCHAR,
220 respectively.

221

222 *Fire episodes and fire-regime changes*

223 To reconstruct regional fire activity and landscape-scale biomass burning we used the long-term trends
224 of two complementary proxies, mCHAR and MCHAR. While their potential source areas are similar (c.
225 40-km distance from study sites; Adolf et al., 2018), the records may differ, with MCHAR potentially
226 reflecting fires within a smaller distance (Finsinger et al., 2017). To determine significant changes in
227 landscape-scale biomass burning, we identified change points based on the mean and the variance in
228 MCHAR and tested their independence from variations in the modelled sediment-accumulation rates
229 (Finsinger et al., 2018).

230 To detect fire episodes within a few km distance around the study site (Higuera et al., 2010), we analysed
231 the MCHAR record with tapas v0.1.2 (Finsinger and Bonnici, 2022). This analysis involved resampling
232 the record to a constant resolution of 33 years (i.e. the 3rd quartile of the sampling-interval distribution).
233 Thereafter, we decomposed the record into a low-frequency background and a high-frequency peak
234 component using a loess robust to outliers with a smoothing-window width of 2000 years. The suitability
235 of the record for peak detection was evaluated through the signal-to-noise index (Kelly et al., 2011).
236 Peaks were evaluated using the 95th percentile of the modelled noise distribution obtained with locally
237 defined Gaussian mixture models for each 2000-year portion of the record (Higuera et al., 2009). The
238 peaks were further screened (Higuera et al., 2010) before calculating the fire return intervals (FRI: years
239 between adjacent fire episodes).

240

241 *Ecosystem responses*

242 We used generalized additive models (GAMs) to investigate ecosystem responses to variations in
243 climate, fire and land use. GAMs are suited for this purpose as they can unfold nonlinear relationships
244 between a response variable and a smoothed function of the predictor variable (Hastie and Tibshirani,
245 1986). Specifically, we explored the responses of fires to variations in climate and land-use pressure,
246 and the responses of dominant forest canopy taxa (*P. abies*, *A. alba*, *F. sylvatica*, *F. excelsior-type*, *Tilia*,

247 and *Ulmus*) to variations in climate, fires and land-use pressure (Carter et al., 2020; Colombaroli et al.,
248 2010; Morales-Molino et al., 2021).

249 As climate data, we used chironomid-inferred summer (July) surface-air temperature anomalies from
250 Lago Verdarolo (1390 m asl) (Samartin et al., 2017) that is in a biogeographically comparable ecosystem
251 in the Tuscan-Emilian Apennines where current vegetation is dominated by fir and beech. Comparisons
252 between instrumental data and climate model runs (Samartin et al., 2017) show that at interannual to
253 multidecadal timescales summer temperature variations in the Tuscan-Emilian Apennines are well
254 correlated with grid cells in Montenegro (Pearson's correlation coefficients are 0.8-1 for multidecadal
255 variability and 0.6-0.7 for interannual variability).

256 We assumed a quasi-Poisson distribution, log-transformed both the response and the predictor variables
257 to stabilize their variances, and fitted GAMs using restricted maximum likelihood (REML) smoothness
258 selection (Carter et al., 2020) with mgcv v1.8-38 (Wood, 2017). We chose a base period from 8000 cal
259 BP to the present because secondary anthropogenic indicators were very abundant during the early
260 Holocene, when human impact was negligible (Carter et al., 2020).

261

262

263 **Results and interpretation**

264

265 *Chronology and sediment composition*

266 The sediments cover part of the Younger Dryas and the entire Holocene, from c. 12,340 cal BP to present
267 (Fig. 2). The reliability of the chronology in its most recent part is broadly confirmed (Fig. S2) by the
268 match between the expected and modelled ages of Pb peaks during the Roman period and the Middle
269 Ages (McConnell et al., 2018; Renberg et al., 2001). Sediment deposition times vary between 6-65 years
270 cm^{-1} (median = 21.6 years cm^{-1}).

271 At the base, from 615 cm to 575 cm (12,340-11,640 cal BP), the sediments consist of light-brown sandy
272 and silty clay (Fig. 2). From 575 cm to 479 cm (11,640-10,300 cal BP), there are 1-2-cm thick light grey
273 silty-clay layers that alternate with darker and organic-richer layers (gradual boundaries), and a c. 28-
274 cm thick sequence of five graded deposits (529-500 cm), each one characterized by a distinct fining-

275 upward trend. Above 479 cm, sediments are mainly composed of dark-brown organic gyttja, which is
276 interrupted by three minerogenic layers of fine grey detrital clay with sharp stratigraphic boundaries at
277 445-441 cm, 413-410 cm, and 283-281 cm depth. In keeping with the visual assessment of the sediments,
278 minerogenic elements (K, Ti, Rb, Zr, and Sr) are most abundant below 479 cm (10,300 cal BP) as well
279 as in conjunction with the minerogenic layers.

280

281 *Long-term vegetation dynamics*

282 The pollen record (Figs. 3-4) was divided into four statistically significant assemblage zones at 566.5,
283 455.5, and 359.5 cm (corresponding to 11,500, 9410, and 6400 cal BP) and three non-significant
284 subzones with boundaries at 200.5 cm and 58.5 cm (2090 and 430 cal BP).

285 Pollen assemblages from 12,340 to 11,500 cal BP (ZMN-1; 615-566.5 cm) are dominated by *Pinus* and
286 herb pollen (mainly *Artemisia*, Poaceae and Chenopodiaceae). Both stomata and plant macrofossils
287 attest the local presence of *Betula* and *Pinus*, the latter presumably as low-density stands of dwarf
288 mountain pine (*P. mugo*) or pine trees (*P. sylvestris*, *P. heldreichii*, or *P. nigra*). Cerealia-type pollen
289 are likely associated with wild Poaceae, as Cerealia-type pollen were also found elsewhere long before
290 the development of agriculture (Dörfler, 2013). Such assemblages are typical of the Central European
291 Younger Dryas (12,850-11,650 cal BP; Rasmussen et al., 2014) and are indicative of an open and arid
292 steppe-tundra with cold-tolerant pine woodlands. Rising arboreal pollen (from c. 45 to 65%) indicate a
293 gradual closure of the woodland during the Younger Dryas cold stage, including mixed oak woodlands
294 at lower altitude.

295 *Picea abies* stomata unequivocally indicate a rapid altitudinal shift of this tree species at the Younger
296 Dryas-Holocene transition, when a rapid climatic change including a warming of c. 2-4°C occurred
297 (Heiri et al., 2014; Tóth et al., 2015). The woodlands closed further from 11,500 to 9410 cal BP (ZMN-
298 2; 566.5-455.5 cm), as attested by an increase of arboreal pollen (from c. 65% to 80%). *F. excelsior*-
299 type, *Ulmus*, and *Tilia* pollen first appeared at 11,500 cal BP marking the expansion of meso-
300 thermophilous ~~thermophilus~~ trees at lower altitude, a characteristic feature of pollen records from the
301 Balkan Peninsula and the adjacent European Alps (Caf et al., 2022; Vescovi et al., 2007; Willis, 1994).
302 *Betula* populations were denser than during the Younger Dryas, as shown by both pollen and plant

303 macrofossils. However, the abundance of light-demanding pioneer trees and shrubs (*Ephedra*, *Juniperus*
304 and *Betula*) decreased in conjunction with the local expansion of spruce and fir populations starting
305 between 10,400 and 9600 cal BP. Finds of *Sporormiella* dung-fungi spores indicate the presence of ~~large~~
306 wild herbivores around the lake between 11,500 and 10,400 cal BP, as also observed on the Swiss
307 Plateau (Rey et al., 2017) and in the Maritime Alps (Finsinger et al., 2021).

308 The ~~local~~ shift to a *Picea-Abies* dominated forest was mostly concluded shortly after 9410 cal BP (onset
309 of ZMN-3; 455.5-359.5 cm), as also observed at Prokoško Jezero (1670 m asl, Bosnia Herzegovina;
310 Dörfler, 2013). Plant-macrofossils show that both *Pinus* and *Betula* were still intermixed in the spruce-
311 fir dominated forest until around 8000 cal BP, when a low-density population of beech likely
312 established, as suggested by the continuous pollen curve. At lower altitudes, vegetation was
313 characterized by mixed oak forests with a higher share of *Corylus* between 9410 and 6000 cal BP.
314 Around 7250 cal BP, *Ostrya*-type pollen increased indicating the expansion of *O. carpinifolia*, *C.*
315 *orientalis*, or of both species.

316 This expansion was followed by an increase of *Fagus* and *C. betulus* pollen starting around 6400 cal BP
317 (onset of ZMN-4a; 359.5-200.5 cm). Although the *Fagus* expansion at ~~Zminje jezero~~ Zminje Jezero
318 occurred approximately 2500-1000 years later than in pollen records from the Central Dinaric mountain
319 area (Prokoško Jezero, Crveni Potok), it broadly coincided with the expansion of *Ostrya* and *Carpinus*
320 *betulus*, as often seen in pollen records from the region (Prokoško Jezero, Lake Ohrid, Lake Prespa, and
321 Crveni Potok; Dörfler, 2013; Finsinger et al., 2017; Panagiotopoulos et al., 2013; Wagner et al., 2009).

322 The admixture of *Fagus* in the ~~local~~ *Picea-Abies* dominated forest involved a c. 2000-years long build-
323 up phase (6400-4600 cal BP), as attested by rising pollen abundance, after which *Fagus* plant
324 macrofossils and stomata regularly occur. As both *C. betulus* and *Fagus* are shade-tolerant and favoured
325 by moist habitats on upland soils (Houston Durrant et al., 2016; Sikkema et al., 2016), an increase in
326 humidity during the growing season may have occurred (Finsinger et al., 2017). In keeping with this,
327 plant macrofossils of more light-demanding and better drought-resistant *Pinus* are absent in this zone,
328 though ~~occasional~~ stomata occasionally finds indicate their scattered presence. *Secale cereale* pollen first
329 appear in this zone and occur mostly between 5800 and 3800 cal BP, ~~unambiguously documenting~~
330 cultivations suggesting its presence as a weed among other cereals during the ~~Copper~~ Bronze Age (Behre,

331 1992). Some secondary anthropogenic-indicator pollen types (e.g. *P. lanceolata*-type, *Melampyrum*)
332 increased as well. However, tree pollen abundance was highest in this time interval, suggesting that
333 land-use pressure was minor. Similarly, a slightly lower tree-pollen abundance between 3800 and 2100
334 cal BP suggests moderate presence of forest openings that are, however, not matched by rising
335 abundance of anthropogenic-indicator pollen types.

336 By contrast, anthropogenic-indicator pollen document higher land-use pressures from Roman times
337 onwards (c. after 2100 cal BP; ZMN-4b and ZMN-4c). *Secale cereale* pollen is mostly present around
338 the Roman period and regularly occurs since the Ottoman Empire (ZMN-4c; 58.5-0 cm; 430 cal BP to
339 present). Similarly, both cultural and secondary indicators are slightly higher during the Roman period
340 and substantially rise since 430 cal BP in conjunction with a substantial tree-pollen decrease. Thus,
341 pollen overall document two land-use phases characterized respectively by moderate land-use pressure
342 (Roman period) and increasing and reaching highest land-use pressure since the Ottoman Empire. These
343 two land-use phases were interrupted by a phase with minor land-use pressure during the Middle Ages
344 (ZMN-4b; 200.5-58.5 cm; 2090-430 cal BP).

345 The frequent occurrence of *C. sativa* and *O. europaea* pollen since 1800 cal BP is consistent with a
346 human-mediated spread of chestnut populations since Roman times (Conedera et al., 2004) and the
347 cultivation of olive trees in Roman territories (Mercuri et al., 2013), including Montenegro, which was
348 conquered by the Romans during the 2nd century BCE. Further, as in other records in the region (Dörfler,
349 2013; Finsinger et al., 2017), pollen of *Juglans* first appear at c. 1500 cal BP and occur more regularly
350 after 500 cal BP.

351 Although *Abies* pollen abundance was substantially lower after 2100 cal BP, a mixed *Abies-Picea-Fagus*
352 was still present locally until 430 cal BP, as confirmed by stomata and plant macrofossils. However,
353 lower pollen percentages and the absence of plant macrofossils of *Fagus* in conjunction with rare finds
354 of *Abies* macrofossils since 430 cal BP suggest a major shift towards a *Picea*-dominated forest. Forest
355 cover increased during the last century, perhaps due to the protection of the area since 1907 CE.

356

357 *Charcoal-inferred fire history*

358 Regional fire activity, inferred based on microscopic charcoal (mCHAR), was highest over the past 2000
359 years and was moderately high between 12,000 and 10,000 cal BP (Figs. 3-4). By contrast, landscape-
360 scale biomass burning macroscopic charcoal (MCHAR trend) ~~documents was~~ highest burned biomass
361 between 11,400 and 6000 cal BP, and overall decreasing ~~biomass burning~~ from 6000 cal BP to the
362 present (Figs. 3-4 and S4). The signal-to-noise index (Fig. S5) shows that the macroscopic charcoal-
363 accumulation rate (MCHAR) record is suitable for peak-detection analysis (Kelly et al., 2011). A total
364 of 35 fire episodes were identified over the past 12,500 years, with a median fire-return interval (mFRI)
365 of 260 years (FRI range: 79-790 years). Longest FRI values (>530 years) occur before 8000 and after
366 2000 cal BP. Charred *A. alba* and *P. abies* needles document the occurrence of local fires in the conifer-
367 dominated mountain ecosystem, mostly between 5600 and 500 cal BP (Figs. 3-4).

368

369 *Response of fire to variations in climate and land-use pressure*

370 Landscape-scale Biomass-biomass burning (~~as inferred from~~ MCHAR trend) and regional fire activity
371 (~~as inferred from~~ mCHAR) show different responses to climate and land-use pressure. Whereas biomass
372 burning significantly increases in response to summer temperatures, fire activity significantly decreases
373 (Fig. 5a; Table 2). Similarly, whereas biomass burning significantly decreases in response to land-use
374 pressure, regional fire activity shows a hump-shaped response including an increase for the lower range
375 of anthropogenic indicator pollen values (Fig. 5b).

376

377 *Response of forest canopy species to variations in climate, fire and land use*

378 *Picea*'s responses to variations in climate, land use and biomass burning are not statistically significant
379 (Fig. 5c-e). *Fagus* shows a significant negative response, and *Abies*, *F. excelsior*-type, *Ulmus*, and *Tilia*
380 show significant positive responses to both summer temperature and biomass burning (Fig. 5c and e;
381 Table 2). *Abies*, *F. excelsior*-type and *Ulmus* show significant negative responses to land-use pressure
382 (Fig. 5d).

383

384

385 **Discussion**

386

387 *Regional vs local land-use*

388 The investigation of long-term vegetation dynamics in relation to land-use pressure and disturbances
389 requires a detailed understanding of land-use history. This is particularly important for southern Europe,
390 where forest clearances and agriculture often involved the use of fire. Moreover, land-use phases were
391 often interspersed by periods of abandonment and secondary forest establishment as the result of the
392 development and expansion of more permanent land-use practices, such as animal husbandry,
393 ploughing, crop cultivation, and woodland management (Birks and Tinner, 2016). However, identifying
394 the start and inferring the intensity of land-use pressure using palaeoecological records is challenging,
395 particularly in regions where independent documentation by archaeological evidence is poor (Deza-
396 Araujo et al., 2020), as is the case for the Durmitor.

397 Stomata and plant macrofossils indicate the local presence of taxa, mostly within some decametres
398 around a site (Ammann et al., 2014; Birks, 2017). Thus, the continuous presence of plant macrofossils
399 and stomata of trees (*Picea*, *Abies*, and *Fagus*) ~~in the sediments of Zminje jezero (Fig. 3)~~ indicates that
400 tree cover persisted in the surroundings of Zminje Jezero ~~the lake~~ at least throughout the past 9000 years
401 (Fig. 3). In keeping with this, high arboreal pollen ~~values of arboreal pollen~~ (generally >80%) and
402 AP/NAP ratio values (generally >4; Fig. 4) qualitatively indicate the presence of closed forest (Deza-
403 Araujo et al., 2020). A persistently high forest cover was also found in the internal Dinaric Alps, at
404 Crveni potok (Finsinger et al., 2017), but contrasts with results from Prokoško Jezero (Dörfler, 2013)
405 where arboreal pollen values decreased to 50% during the past 2000 years. This contrasting evidence
406 would support the hypothesis that some of the forested regions in the interior range of the Dinaric Alps
407 have been left relatively intact until the present (Nagel et al., 2017).

408 Conversely, if land-use pressure was regionally heterogeneous, pollen reconstructions based on pollen
409 may not necessarily covary with those based on plant macrofossils and stomata. Pollen from the wider
410 region ~~including from lower elevations~~ may reach small lakes such as ~~Zminje jezero~~ Zminje Jezero even
411 if their relevant pollen source area is relatively small (some kilometres at most; Sugita, 1994). ~~Thus,~~
412 ~~reconstructions based on plant macrofossils and stomata may not necessarily covary with those based~~
413 ~~on pollen if land use pressure was not homogeneous across the region.~~ Indeed, despite persistence of

414 closed forest in the surroundings of Zminje Jezero, anthropogenic indicators unambiguously document
415 increasing land-use pressure in the region since the Roman period. Cultural indicators, secondary
416 indicators, as well as OJC-index values rise around 2000 cal BP, when the region currently belonging
417 to Montenegro was conquered by the Romans (2nd century BCE). At that time, land-use pressure
418 probably reached the Durmitor, as attested by the construction of roads, including the one connecting
419 Nikšić and Komini. The further rise of anthropogenic indicators around 500 cal BP (15th century CE) is
420 coherent with the first local settlements in the Durmitor. For instance, the so-called stećci sites (medieval
421 tombstones and graveyards from the 13th-17th century; Erdeljan, 2018) attest to the local presence of
422 settlements on the Durmitor during the Middle Ages and the Ottoman Empire, and churches and
423 monasteries from the 15th-19th century CE document the presence of local settlements during the
424 Ottoman Empire~~the presence of *Secale cereale* pollen unambiguously documents that.~~ Thus, these
425 results add a finer-scale scale documentation of past land-use pressure in comparison to modelled
426 anthropogenic deforestation patterns (Kaplan et al., 2009), which suggest that relatively low forest
427 clearance occurred during the past 3000 years and that the region was not heavily exploited until the
428 Middle Ages.

429 Similarly, regional heterogeneity of land-use pressure may also have occurred in pre-Roman times.
430 Anthropogenic indicators do not show distinct signs of Bronze and Iron Age land use, despite the
431 probable connection between coastal and inland settlements (Bulatović et al., 2020) and Illyrian and
432 Celtic Iron Age settlements (Cozzolino et al., 2020). However, *Secale cereale* pollen finds at Zminje
433 Jezero suggest its presence as a weed among other cereals (Behre, 1992) since about 6000 cal BP, as
434 also observed in archaeobotanical samples from Neolithic sites in the mountain hinterland (e.g. at
435 Okolište, Bosnia-Herzegovina; de Vareilles et al., 2022).~~cultivation started in the region at least around~~
436 ~~6000 cal BP, during the Copper Age.~~ While this is about 1000 years later than the age of Neolithic
437 settlements on the shores of the lowland Lake Ohrid (Hafner et al., 2021) and 2000 years later than the
438 earliest Starčevo and Butmir-culture Neolithic settlements in the inland of Bosnia-Herzegovina (Vander
439 Linden et al., 2014), the *Secale* pollen finds support the view that Neolithic novelties, albeit with a
440 certain time lag, spread in the hinterland of Montenegro (Borić et al., 2019). Although Neolithic sites
441 have not been found yet in the Durmitor area~~On the other hand, there is some evidence for~~ of Neolithic

442 occupation sites inland, such as the Odmuť cave at 20-35 km from the Durmitor distance (Odmuť rock
443 shelter; Borić et al., 2019), suggesting that inland mountain areas may be archeologically under surveyed
444 (Vander Linden et al., 2014). Nevertheless
445 ~~On the one hand, the sparse occurrence of *Secale* pollen~~ palaeoecological records may agree with the
446 hypothesis that the mountain hinterland remained an agricultural frontier zone for much of the Neolithic
447 (Forenbaher and Miracle, 2005) and possibly even until the Copper-Roman Age. ~~On the other hand,~~
448 ~~there is some evidence for Neolithic occupation sites inland, such as the Odmuť cave 20-35 km from the~~
449 ~~Durmitor (Borić et al., 2019), suggesting that inland mountain areas may be archeologically under~~
450 ~~surveyed (Vander Linden et al., 2014). Similarly, anthropogenic indicators do not show distinct signs of~~
451 ~~Bronze and Iron Age land use, despite the probable existence of connections between coastal and inland~~
452 ~~settlements (Bulatović et al., 2020) and Illyrian and Celtic Iron Age settlements (Cozzolino et al., 2020).~~
453 ~~By contrast, anthropogenic indicators unambiguously document increasing land use pressure since the~~
454 ~~Roman period. Cultural indicators, secondary indicators, as well as OJC index values rise around 2000~~
455 ~~cal BP, when the region currently belonging to Montenegro was conquered by the Romans (2nd-century~~
456 ~~BCE). At that time, land use pressure probably reached the Durmitor, as attested by the construction of~~
457 ~~roads, including the one connecting Nikšić and Komini. The further rise of anthropogenic indicators~~
458 ~~around 500 cal BP (15th-century CE) is coherent with the first local settlements in the Durmitor. For~~
459 ~~instance, the so-called stećci sites (medieval tombstones and graveyards from the 13th-17th century;~~
460 ~~Erdoljan, 2018) attest to the local presence of settlements on the Durmitor during the Middle Ages and~~
461 ~~the Ottoman Empire, and churches and monasteries from the 15th-19th-century CE document the presence~~
462 ~~of local settlements during the Ottoman Empire.~~

463

464 *Regional vs local fires*

465 We found contrasting responses of fire to climate and land-use pressure over the past 8500 years (Fig.
466 5a-b), pointing to different drivers of fire dynamics at landscape and regional scales. Landscape-scale
467 Biomass-biomass burning, as inferred based on MCHAR, responded positively and significantly to
468 summer temperature, as expected based on the strong influence (summer) temperature has on fire
469 activity (Jain et al., 2022; Westerling et al., 2006) and as often observed with sedimentary charcoal

470 records (Daniau et al., 2012; Power et al., 2008). As fuel was not a limiting factor in the moist and
471 productive mountain forests, higher mean July temperatures perhaps coupled with prolonged dry
472 seasons (Holocene Thermal Maximum; Samartin et al., 2017) likely improved fuel flammability (Pausas
473 and Paula, 2012). By contrast, surprisingly regional fire activity, as inferred based on mCHAR,
474 responded negatively to summer temperature. A disconnect between fire-history reconstructions can
475 arise both when comparing paleofire records from sites located near each other (Finsinger et al., 2018;
476 Gavin et al., 2006), as well as different paleofire proxies from the same site (Finsinger et al., 2017) if
477 fire occurrence was not spatially homogeneous. While charcoal records from additional sites may be
478 necessary to better constrain the spatial heterogeneity of fire occurrence in the region, it is likely that
479 mCHAR captures fire activity at greater distance from the site than MCHAR (Finsinger et al., 2017).
480 This interpretation is supported by the positive response of fire activity (mCHAR) to land-use pressure
481 (Fig. 5b), which likely occurred in a wider region, including lower-elevation areas where ~~sub-~~
482 ~~Mediterranean~~submediterranean deciduous woodlands occur. The slightly earlier rise of regional fire
483 activity, which started around the Iron Age (Fig. 4), may be connected to land use in the region that
484 remained undetected by the anthropogenic indicators from ~~Zminje-jezero~~Zminje Jezero. Biomass
485 burning, instead, responded negatively to land-use pressure (Fig. 5c). It is unlikely that this relationship
486 is the result of feedbacks arising from human-induced deforestation (e.g. lower fuel availability,
487 fragmentation; Bowman et al., 2011) because biomass burning decreased predominantly at the
488 Neolithic/Bronze Age transition, when land-use pressure was low. The overall decreasing biomass
489 burning suggests that humans did not extensively use fire in the Durmitor. While this overall contrasts
490 with paleofire evidence from other southern European mountains (e.g. the Apennines; Morales-Molino
491 et al., 2021), the lower biomass burning rates in recent millennia fit with the notion that fires are currently
492 rare in the Dinaric Alps (Nagel et al., 2017). However, charred *Picea* and *Abies* needles unambiguously
493 document the occurrence of fires in the mixed spruce-fir-beech mountain forests of the Durmitor.

494

495 *Species-response curves and legacies of past environmental changes*

496 Species-response curves are useful to assess factors that have driven long-term vegetation dynamics,
497 thereby unfolding legacies of past environmental changes (Carter et al., 2020; Colombaroli et al., 2010;

498 Morales-Molino et al., 2021). We found similar species responses to summer temperature and biomass
499 burning (Fig. 5c and 5e). Given the strong and positive relationship among these two predictors (Fig.
500 5a), it is difficult to disentangle their roles separately. However, as (summer) temperature currently has
501 a strong influence on fire activity (Jain et al., 2022; Westerling et al., 2006), the response curves may
502 yield valuable insights into the long-term fire ecology of the analysed species (*P. abies*, *F. sylvatica*, *A.*
503 *alba*, *Ulmus*, *Fraxinus excelsior*, and *Tilia*), which are widespread and often dominant in European
504 mountain forests. The right-skewed responses of *Fagus* to biomass burning and temperature (Fig. 5c
505 and 5e) suggest that it was favoured under low biomass-burning and cooler conditions. Overall, these
506 response curves support the notion that *Fagus* became more competitive in mixed fir-spruce forests with
507 the onset of cooler and moister summers and fire disturbance, as observed in the ecologically similar
508 Northern Apennines (Morales-Molino et al., 2021). Indeed, *Fagus* requires moist summers for its
509 germination (Giesecke et al., 2017) and is more sensitive than *Abies* to drought due to its shallow root
510 system (Cheddadi et al., 2016; Tinner et al., 2013). *Fagus* is a shade-tolerant species and its regeneration
511 can establish and grow with low light levels under the forest canopy for years before rapidly expanding
512 and achieving dominance when light conditions improve after disturbances (Gardner and Willis, 1999).
513 Thus, fire occurrence after 5000 cal BP may have promoted the recruitment of *Fagus* and its admixture
514 into the mixed *Picea-Abies* forest. This matches with evidence suggesting that during the initial phase
515 of stand establishment, *Fagus* is favoured by mixed-severity fire and intermediate human disturbances
516 (Bradshaw and Lindbladh, 2005; Giesecke et al., 2007; Tinner and Lotter, 2006) thanks to its capacity
517 to occupy newly exposed soils and canopy openings (Carter et al., 2018) more rapidly than its two main
518 competitors (*Abies* and *Picea*). ~~However~~Despite, lacking ecological adaptations to fire, and being
519 identified as *Fagus* is a fire-sensitive species that is affected by higher biomass burning and temperatures
520 (Carter et al., 2018; Morales-Molino et al., 2021; Tinner et al., 2000), *Fagus* can be advantaged by fire
521 incidence as long as it is not excessive (Maringer et al., 2020). In keeping with this, the change-point
522 analysis shows a decrease in biomass burning during the establishment of *Fagus* (Fig. 4), corroborating
523 the view that beech establishment occurs in conjunction with a shift to a lower fire activity (Bobek et
524 al., 2019; Feurdean et al., 2017).

525 Further, similar to observations from the ecologically similar Northern Apennines (Morales-Molino et
526 al., 2021), we found that *Abies* and disturbance-sensitive deciduous taxa such as *Ulmus*, *F. excelsior*,
527 and *Tilia* were promoted by warmer-than-present summer temperatures (up to ca. 2.5°C; Fig. 5c).
528 However, our findings also suggest that these taxa responded positively to increasing biomass burning
529 (Fig. 5e). This contrasts with several records showing that both high-severity infrequent fires and low-
530 severity frequent fires led to their rapid decline and to local extinctions in recent millennia (Morales-
531 Molino et al., 2021; Tinner et al., 2000, 2013). On the one hand, the striking positive response of these
532 disturbance-sensitive taxa to biomass burning could simply arise from the positive relationship between
533 landscape-scale biomass burning and summer temperatures (Fig. 5a). On the other hand, the significant
534 negative response of *Abies*, *Ulmus*, and *F. excelsior* to land-use pressure (Fig. 5d) as well as to regional
535 fire activity (Fig. S6 and Table S2) indicates that their populations declined even in the wider region
536 due to land use, probably with the use of fires. Indeed, the decline of *Abies* coincides with the rise in
537 land-use pressure (pasturing and small-scale cultivation) and of regional fire activity since 2000 cal BP.
538 Conversely, the local persistence of *Abies* in the surroundings of ~~Zminje jezero~~Zminje Jezero until
539 present-day suggests that land-use pressure was not widespread across the montane forest belt, leaving
540 relatively undisturbed patches of forests in a landscape mosaic (Fig. 1). Indeed, *Abies* can be resistant
541 to fire disturbances when fires are rare (Henne et al., 2013) and its populations strongly reduce under
542 excessive human-induced fire activity (Feurdean and Willis, 2008; Finsinger et al., 2021; Tinner et al.,
543 2013), which was not the case around ~~Zminje jezero~~Zminje Jezero but may have been the case to the
544 East of Žabljak and elsewhere in the upper-montane zone of the central Dinaric Alps.

545 *Picea*'s response curves suggest that neither temperature nor biomass burning and land use significantly
546 influenced its populations (Fig. 5c-e). Indeed, *Picea* remained relatively stable even under higher-than-
547 present summer temperatures (Fig. 4), as also observed with palaeoecological records from the
548 Carpathians (Carter et al., 2018). Its insensitivity to summer temperatures may simply imply that past
549 climatic conditions did not exceed the species' tolerance range. Moreover, the lack of a significant
550 response to biomass burning agrees with evidence showing that *Picea* can persist thanks to vigorous
551 post-fire regeneration from wind-dispersed seeds (Bobek et al., 2018) even with median FRI as low as
552 260 years as at ~~Zminje jezero~~Zminje Jezero (Brown and Giesecke, 2014; Feurdean et al., 2017; Finsinger

553 et al., 2018). Although fires may have been fuelled by *Picea*'s traits that promote fire spread (high resin
554 content, flammable litter, and a ladder-canopy structure), its well-established populations persisted at
555 [Zminje jezero](#)Zminje Jezero despite fire occurrence in the mixed spruce-fir-beech forests.

556 The species-response curves may also be useful to explain the relatively recent shift from a *Picea-Abies-*
557 *Fagus* mixed forest to one dominated by *Picea* that occurred in the surroundings of [Zminje jezero](#)Zminje
558 Jezero around 430 cal BP (15th century CE; Fig. 4). Fir populations, which are more sensitive to land-
559 use pressure than *Picea* and *Fagus* (Fig. 5d), could have been reduced due to the establishment of the
560 local pastoral katun society (c. 1477 CE; Turkish Land Registry). Around 1400 CE, the Balkan Peninsula
561 was under the influence of the Republic of Venice and from 1496 CE Montenegro fell under the domain
562 of the Ottoman Empire (Rastoder, 2003). In the mid-16th century, local land-owners started seizing
563 monastery estates and hunting grounds in the lower parts of Montenegro. This resulted in the retreat of
564 considerable parts of the population in mountain areas (Rastoder, 2003), including the Durmitor, whose
565 forests probably started to be more intensively exploited. An alternative, or additional explanation could
566 involve the lower sensitivity of spruce to declining winter temperatures (Cheddadi et al., 2016). In this
567 scenario, the development of a *Picea*-dominated forest could be linked to the onset of cooler conditions
568 during the Little Ice Age (1450-1850 CE). Indeed, due to its shallow root system *Picea* is more
569 vulnerable to drought than *Abies* (Henne et al., 2011), and spruce-dominated stands in the Dinaric Alps
570 mostly occur in habitats with significantly colder and moister microclimates (Nagel et al., 2017).

571

572

573 **Conclusions**

574 Vegetation of forest ecosystems is often the result of complex interactions between vegetation dynamics
575 and variations in climate, disturbances (e.g. fire), and human activities. However, understanding the
576 ecological processes underlying these interactions is challenging as processes often play out over long
577 time scales (centuries, millennia), particularly for forests where long-lived species such as trees are
578 involved. The long-term perspective of palaeoecological records permits to study ecosystem responses
579 under substantially different environmental conditions than the present ones.

580 In this study, we sought to investigate the Holocene vegetation and fire history in the upper-montane
581 zone of the central Dinaric Alps, where fires are currently rare and forests mainly include beech-
582 dominated stands and mixed forests dominated by varying amounts of beech, fir, and occasionally
583 spruce (Nagel et al., 2017). ~~We performed the first detailed palaeoecological study from this region~~
584 ~~(including pollen, plant macrofossils, microscopic and macroscopic charcoal analysis) on well dated~~
585 ~~sediments from Zminje jezero (Durmitor massif, Montenegro) and assessed the response of the~~
586 ~~dominant canopy taxa (*P. abies*, *A. alba*, *F. sylvatica*, *F. excelsior* type, *Tilia*, and *Ulmus*) to variations~~
587 ~~in climate, fire and land-use pressure.~~
588 ~~Vegetation in the upper montane zone was dominated by *Pinus* and *Betula* during the Younger Dryas~~
589 ~~and until 9500 cal BP, when a *Picea-Abies* forest established. This mixed forest remained relatively~~
590 ~~stable until 5000 cal BP, when *Fagus* expanded in the mixed forest. Biomass burning gradually~~
591 ~~decreased during the past 8500 years in the upper montane zone, though fire episodes did occur there~~
592 ~~throughout that period of time. Land-use pressure was overall remarkably low in the montane zone until~~
593 ~~the Roman period (2000 cal BP) but increased substantially after the Middle Ages. By contrast, regional~~
594 ~~fire activity rose since c. 2500 cal BP (Iron Age) as a result of higher land-use pressure in the wider~~
595 ~~region, probably including areas at lower elevation.~~ While records from different elevations may be
596 necessary to better support ~~these our results~~ interpretation, our results support the view that at least some
597 parts of the central Dinaric Alps remained an agricultural frontier zone during the Neolithic (Forenbaier
598 and Miracle, 2005) and that land-use pressure was low until the Middle Ages (Kaplan et al., 2009).
599 Response curves indicate that fires in the upper-montane zone were mostly driven by variations in
600 climatic conditions, as expected based on the strong influence (summer) temperature has on fire activity
601 today (Jain et al., 2022; Westerling et al., 2006). *Picea* was insensitive to variations in summer
602 temperature, biomass burning and human impact, which supports the view that spruce forests may not
603 be significantly impacted by fire (Carter et al., 2018). In contrast, *Abies* and other disturbance-sensitive
604 deciduous trees (*F. excelsior*-type, *Tilia*, and *Ulmus*) show a significant positive relationship with
605 summer temperatures and a significant negative relationship with land-use pressure. This supports the
606 notion that these tree species may be well-adapted to warmer-than-present summer temperatures and
607 that their populations declined in recent millennia due to land-use activities (Morales-Molino et al.,

608 2020; Tinner et al., 2000, 2013). Conversely, *Fagus* emerges as a species sensitive to summer
609 temperatures whose expansion in the *Picea-Abies* dominated forest was enhanced by the onset of cooler
610 and possibly moister climatic conditions as well as by fire disturbances.

611

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620

621

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625

626 **Data**

627 All data will be made publicly available upon acceptance. The palaeoecological data (pollen, spores,
628 stomata, plant macrofossils, charcoal, and XRF records) will be uploaded to the Neotoma database
629 (DOI: 10.21233/XYZZ-XYZZ) through the European Pollen Database.

630 **Figures (with captions) and Tables**

631 **Fig. 1** a-c) Location of the study site at decreasing spatial scales. b-c) Map showing the distribution of
632 land cover types (CORINE Land Cover 2018) in the Durmitor National Park and surroundings. d)
633 bathymetric map of Zminje jezeroZminje Jezero with contour lines at 1-m intervals and coring locations
634 (red circles).

635
636 **Fig. 2** Depth-age model for the Zminje jezeroZminje Jezero sediment core. The active control points
637 used to constrain the model (Table 1) are based on ^{14}C , ^{210}Pb , and ^{137}Cs dates as well as on a pollen-
638 inferred age estimate (onset of the Holocene). Passive control points based on local maxima in the XRF-
639 Pb record (McConnell et al., 2018; Renberg et al., 2001) are shown in green (see also Fig. S2). Grey
640 horizontal bands indicate turbidite layers that were excluded from the model. Lithological and XRF
641 geochemical profiles of the sediment core are shown on the right. The inset shows the distribution of
642 sediment deposition time within the sequence (median = 21 years cm^{-1})

643
644 **Fig. 3** Vegetation and fire history from Zminje jezeroZminje Jezero sediments. Left: synthetic pollen
645 diagram with selected pollen and spore percentages of the terrestrial pollen sum (grey curves; empty
646 curves show 10x exaggerations of the percentage values), presence/absence of stomata (black circles),
647 and plant-macrofossil concentrations (black vertical bars, dimensionless) for selected genera and for all
648 woody taxa of the Zminje jezeroZminje Jezero. The continuous horizontal and dashed lines indicate
649 statistically significant and non-significant pollen-assemblage zones, respectively. See Fig. S3 for a
650 more detailed plant-macrofossil diagram. To the right: macroscopic-charcoal accumulation rate
651 (MCHAR) and microscopic-charcoal accumulation rate (mCHAR) as black bars, their long-term trends
652 (red and black lines), fire episodes (black crosses), and presence of charred conifer needles (red
653 diamonds).

654
655 **Fig. 4** Comparison of the main proxies for vegetation dynamics, fire and land-use history from Zminje
656 jezeroZminje Jezero, and chironomid-inferred July-air temperature (T_{July}) anomalies from Lago
657 Verdarolo (Samartin et al., 2017). Microscopic charcoal (mCHAR, thick dark-grey line) documenting
658 regional fire activity and macroscopic charcoal (MCHAR, black filled area) documenting landscape-
659 scale biomass burning (red continuous line), local fire episodes (black crosses), periods of biomass
660 burning as determined by the change-point analysis (grey-shaded areas), and charred *Picea* and *Abies*
661 needles (red diamonds). Temporal changes of dominant forest canopy taxa and of anthropogenic
662 indicator pollen types (filled polygons: pollen percentages; black vertical bars: plant macrofossils
663 concentrations; grey circles: stomata). Grey vertical shaded areas represent pollen zone boundaries (see
664 Fig. 3).

665
666 **Fig. 5** Generalized Additive Models showing (a-b) the responses of landscape-scale biomass burning
667 and regional fire activity to climate and land-use pressure, and (c-e) the responses of dominant forest
668 canopy taxa to climate, land-use pressure, and biomass burning.

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