

AperTO - Archivio Istituzionale Open Access dell'Università di Torino

**Hydrological, thermal and chemical influence of an intact rock glacier discharge on mountain stream water**

**This is a pre print version of the following article:**

*Original Citation:*

*Availability:*

This version is available <http://hdl.handle.net/2318/1944272> since 2023-11-24T08:23:05Z

*Published version:*

DOI:10.1016/j.scitotenv.2023.162777

*Terms of use:*

Open Access

Anyone can freely access the full text of works made available as "Open Access". Works made available under a Creative Commons license can be used according to the terms and conditions of said license. Use of all other works requires consent of the right holder (author or publisher) if not exempted from copyright protection by the applicable law.

(Article begins on next page)

1 **Hydrological, thermal and chemical influence of an intact rock glacier discharge**  
2 **on mountain stream water**

3 Bearzot, F.<sup>1,2,\*†</sup>, Colombo, N.<sup>3,4,†</sup>, Cremonese, E.<sup>5</sup>, Morra Di Cella, U.<sup>5</sup>, Drigo, E.<sup>6</sup>, Caschetto, M.<sup>1</sup>, Basiricò,  
4 S.<sup>1</sup>, Crosta, G.B.<sup>1</sup>, Frattini, P.<sup>1</sup>, Freppaz, M.<sup>4</sup>, Pogliotti, P.<sup>5</sup>, Salerno, F.<sup>7</sup>, Brunier, A.<sup>5</sup>, Rossini, M.<sup>1</sup>

5

6 <sup>1</sup>Department of Earth and Environmental Sciences, University of Milano-Bicocca, Piazza della Scienza 1,  
7 20126, Milan, Italy.

8 <sup>2</sup>Faculty of Science and Technology, Free University of Bozen-Bolzano, Piazza Università 1, 39100, Bozen,  
9 Italy.

10 <sup>3</sup>Water Research Institute, National Research Council of Italy, Rome, Italy.

11 <sup>4</sup>Department of Agricultural, Forest and Food Sciences, University of Turin, Grugliasco, Italy.

12 <sup>5</sup>Environmental Protection Agency of Valle d'Aosta, Climate Change Unit, Saint-Christophe, Italy.

13 <sup>6</sup>Geologist freelance.

14 <sup>7</sup>Institute of Polar Sciences, National Research Council of Italy, Milan, Italy.

15 \*Corresponding author: *francesca.bearzot@unibz.it*

16 †These authors equally contributed to the paper.

17 **Abstract**

18 Rock glaciers are the most prominent permafrost-related mountain landforms. This study investigates the  
19 effects of the discharge from an intact rock glacier on the hydrological, thermal and chemical dynamics of a  
20 high-elevation stream in the NW Italian Alps. Despite draining only 39% of the watershed area, the rock glacier  
21 sourced a disproportionately large amount of discharge to the stream, with the highest relative contribution to  
22 the catchment streamflow occurring in late summer - early autumn (up to 63%). However, ice melt was  
23 estimated to be only a minor component to the discharge of the rock glacier, due to its insulating coarse debris  
24 mantle. The sedimentological characteristics and internal hydrological system of the rock glacier played a  
25 major role in its capability to store and transmit relevant amounts of groundwater, especially during the  
26 baseflow periods. Besides the hydrological influence, the cold and solute-enriched discharge from the rock  
27 glacier significantly lowered the stream water temperature (especially during warm atmospheric periods) as  
28 well as increased the concentrations of most solutes in the stream. Furthermore, in the two lobes forming the  
29 rock glacier, different internal hydrological systems, likely driven by different permafrost and ice content,  
30 caused contrasting hydrological and chemical behaviours. Indeed, higher hydrological contributions and  
31 significant seasonal trends in solute concentrations were found in the lobe with higher permafrost and ice  
32 content, due to different flowpaths. Our results highlight the relevance of rock glaciers as water resources,  
33 despite the minor ice melt contribution, also suggesting their potential, increasing hydrological importance in  
34 the light of climate warming.

35

36 **Keywords:** rock glaciers, hydrology, permafrost, discharge, water chemistry, climate change

## 37 **1. Introduction**

38 Billions of people depend on water supply from mountain regions (Viviroli et al., 2020), where climate  
39 change is driving a shift in precipitation phase from snow to rain and a trend toward earlier snowmelt (Huss et  
40 al., 2017). At the same time, mountain glaciers around the world are rapidly receding (e.g., Hugonnet et al.,  
41 2021; Zemp et al., 2015), shifting the predominant dynamics of the mountain cryosphere from glacial to  
42 periglacial (e.g., Haeberli et al., 2017; Seppi et al., 2015). This shift could have relevant implications for  
43 mountain water resources (e.g., Arenson et al., 2022; Wagner et al., 2021a). For instance, the projection of ice  
44 loss rates indicates that in the next decades more subsurface ice may remain compared to glacier surface ice,  
45 due to their different response times to atmospheric changes (Arenson et al., 2022; Haeberli et al., 2017). In  
46 turn, permafrost degradation and subsurface ice loss may partially offset water shortages by increasing the  
47 water storage capacity of mountain terrains due to the increase in the unfrozen sediment thickness (Rogger et  
48 al., 2017).

49 Rock glaciers are the most prominent permafrost-related mountain landforms (Haeberli et al., 2006). Intact  
50 rock glaciers consist of a seasonal frozen layer (active layer) covering ice-supersaturated debris and/or pure  
51 ice (Jones et al., 2019), differing from relict rock glaciers which do not longer contain ice (Haeberli et al.,  
52 2006). Intact rock glaciers can store water in solid (ice/snow) and liquid form (groundwater) (Wagner et al.,  
53 2021a), potentially representing a relevant water source under future climate warming conditions (Arenson et  
54 al., 2022; Jin et al., 2022; Jones et al., 2018). Rock glacier water storage occurs at long-term (ice storage from  
55 multiannual to millennial timescales), intermediate-term (snow storage and release of water on seasonal  
56 timescales) and short-term timescales (water diurnal drainage) (Jones et al., 2019). Despite the increasing  
57 attention gained by intact rock glaciers as water sources, their contribution to the runoff of high-elevation  
58 catchments, and its seasonal variations, remain largely unexamined.

59 Discharge from intact rock glaciers usually originates from snowpack and internal ice melting, rainfall, and  
60 groundwater (Krainer et al., 2007). Commonly, the highest discharge rates occur during the spring - early  
61 summer snowmelt, and gradually decrease through summer and autumn with the lowest flow rates recorded in  
62 the winter months (Krainer and Mostler, 2002). In this framework, the identification and quantification of the  
63 contribution of the different water sources composing the discharge of intact rock glaciers is still unclear (cf.,  
64 Krainer et al., 2007; Williams et al., 2006), above all the permafrost-ice melt rate and its contribution to the

65 water budget. In addition, irrespective of the water source, due to their cold discharge (Millar et al., 2013;  
66 Krainer et al., 2007; Krainer and Mostler, 2002), rock glaciers are also emerging as potential climate refugia,  
67 enabling the near- and long-term persistence of cold habitats and related biodiversity (Brighenti et al., 2021a).  
68 However, how cold waters emerging from intact rock glaciers can influence the thermal regime of high-  
69 elevation surface waters remains poorly investigated.

70 The degradation of permafrost and ice melting in rock glaciers can also affect the chemical characteristics  
71 of surface water (Colombo et al., 2018b). High content of solutes, such as ions, nutrients, and trace elements  
72 have been reported in intact rock glacier outflows (Brighenti et al., 2019; Colombo et al., 2019; 2018c;  
73 Williams et al., 2007; 2006). This occurrence has been also shown to impact the ecological conditions of  
74 downstream water bodies (Mania et al., 2018; Thies et al., 2013) and even the suitability of surface waters  
75 downstream of rock glaciers for use as safe, potable water sources (Ilyashuk et al., 2018; 2014). Despite the  
76 increasing number of studies devoted to the analysis of intact rock glacier hydrochemistry (Jones et al., 2019;  
77 Colombo et al., 2018a), it is evident that further scientific investigation is required, particularly focused on the  
78 quantification of the hydrochemical influence of intact rock glaciers on downstream water bodies.

79 The aim of this research is to assess the hydrological, thermal and chemical influence of an intact rock  
80 glacier discharge on high-elevation stream water in the NW Italian Alps (Valtournenche Valley, Aosta Valley).  
81 The specific objectives are to: (i) quantify the rock glacier hydrological influence (including the internal ice  
82 melt rate), and its seasonal variations, on the hydrologic stream regime; (ii) investigate how the rock glacier  
83 discharge thermally and chemically influences the stream water.

84

## 85 **2. Study area**

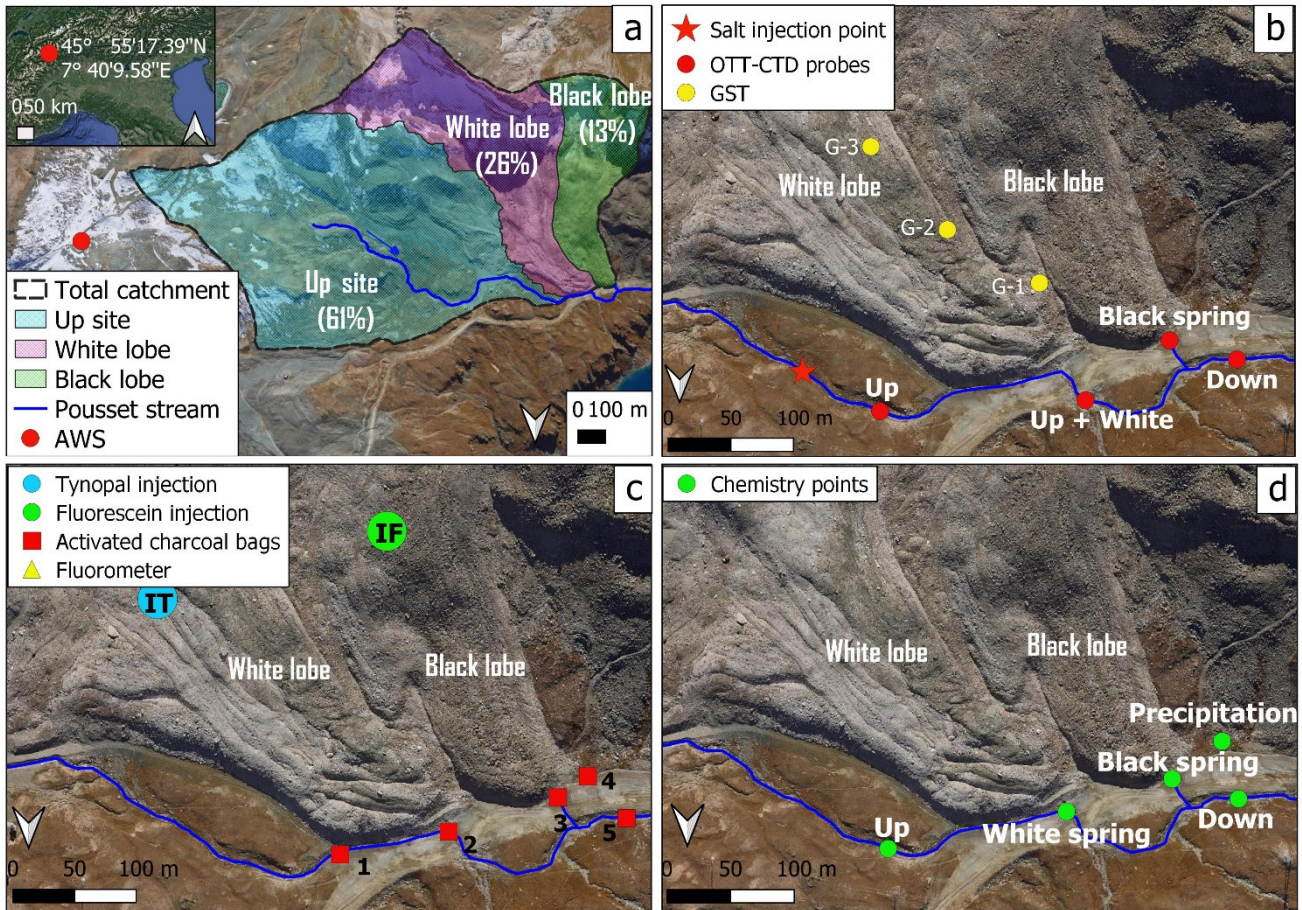
86 The study area is located in the NW Italian Alps, at the head of the Valtournenche Valley (Aosta Valley),  
87 in the south-eastern side of the Cervinia basin (Fig. 1a). The study is focused on the intact Gran Sometta Rock  
88 Glacier, a tongue-shaped landform originating from the rock walls of the Gran Sometta Peak (3165 m a.s.l.).  
89 The rock glacier is approximately 400-m long, between 150- and 300-m wide, with a thickness of 20–30 m,  
90 extending in elevation from 2630 to 2770 m a.s.l. (Bearzot et al., 2022).

91 The surface of the rock glacier consists of longitudinal ridges in the extensive central part, and a complex  
92 of transverse ridges and furrows in the compressive terminal part (Fig. 1). The debris cover is composed of

93 pebbles and angular blocks, in most places lacking any visible fine-grained matrix. Two adjacent lobes  
94 constitute the main body of the rock glacier, called “Black” and “White”, given their different colours, driven  
95 by the lithological composition. Indeed, the Black lobe is mainly composed of calcschists, with a secondary  
96 presence of dolomitic marbles and metabasites. The White lobe is primarily composed of dolomitic marbles,  
97 with a secondary presence of calcschists and metabasites. The remaining part of the study area is lithologically  
98 similar to the rock glacier. Bare cliffs and debris talus deposits are dominant outside the rock glacier, while  
99 soil cover with alpine meadows can be found in restricted areas, especially along the main stream draining the  
100 catchment, the Pousset stream.

101 The two rock glacier lobes differ in terms of internal structure and surface velocities (Bearzot et al., 2022).  
102 The internal structure of the White lobe is composed by two high-resistivity bodies in the central (30-m  
103 thickness) and terminal (20-m thickness) parts of the tongue, which are interpreted as substantial frozen ground  
104 occurrences. These parts are overlaid by a less resistive layer representing the active layer and, below, by  
105 resistivity values  $< 5 \text{ kohm m}$  indicating unfrozen conditions. This lobe moves downstream at an average  
106 annual surface velocity of  $0.5 \text{ m y}^{-1}$ . The Black lobe displays a different internal structure, showing a  
107 continuous layer (20-m thickness) with high resistivity values ( $\sim 100 \text{ kohm m}$ ) under the active layer, indicating  
108 a higher frozen ground content with respect to the White lobe. This lobe moves downstream at a velocity of  
109 about  $1 \text{ m y}^{-1}$ , twice as fast as the White lobe, likely due to its higher frozen ground content and, possibly,  
110 slightly different topographic conditions (Bearzot et al., 2022).

111 Two springs are located at the fronts of the lobes, discharging into the main stream. In the studied watershed  
112 ( $1.32 \text{ km}^2$ ), three hydrological catchments were outlined using the surface topography (2m×2m-cell digital  
113 terrain model), named “Up site” (upstream of the rock glacier), “White lobe”, and “Black lobe” (the last two  
114 related to the studied rock glacier lobes), draining 61% ( $0.80 \text{ km}^2$ ), 26% ( $0.34 \text{ km}^2$ ), and 13% ( $0.17 \text{ km}^2$ ) of  
115 the watershed area, respectively (Fig. 1a).



116

117 *Figure 1. (a) Overview map of the study area, showing the sub-basin divisions with the respective planar-area contribution to the total*  
 118 *catchment (%), the location of the Cime Bianche AWS, and the location of the study area in northern Italy (inset) (coordinate system:*  
 119 *WGS84). Detailed view of the rock glacier measuring and sampling points: (b) three ground surface temperature (GST) data loggers,*  
 120 *salt injection point, four sites equipped for the measurement of the discharge through the salt-dilution method, and locations of the*  
 121 *probes; (c) five activated charcoal points, one field fluorometer, and two dye injection points; (d) isotopic and water chemistry*  
 122 *sampling sites (the sampling point “Snow” is not shown since it is placed further upstream).*

123

124 **3. Data and methods**

125 **3.1 Meteorological conditions and snow-cover duration**

126 A high-elevation automatic weather station (AWS) is located about 1 km West from the rock glacier front,  
 127 the Cime Bianche AWS (3100 m a.s.l.) (Fig. 1a). For the time span 2010–2021, the mean annual air  
 128 temperature was  $-2.6 \pm 0.9$  °C. In summer (June, July, and August), the mean air temperature was around  
 129  $+4.4 \pm 1.8$  °C. The snowpack generally developed by early-October, and the melt out of snow occurred around  
 130 July. The mean maximum snow depth was  $1.8 \pm 0.3$  m. The liquid precipitation during the snow-free season

131 (no snow on the ground) was  $262\pm 138$  mm on average. To support the investigations in the years 2019-2020-  
132 2021, daily air temperature, snow-depth and precipitation data from the Cime Bianche AWS were analysed.

133 Snow-cover duration in the catchment and on the rock glacier was estimated using two approaches with  
134 different spatial and temporal resolutions in order to adequately assess the presence/absence of snow in the  
135 main stream catchment and in-between the coarse debris on the rock glacier surface. To provide an estimate  
136 of the snow-covered area evolution in the entire basin, the Fractional Snow Cover (FSC) product (Dumont et  
137 al., 2021) of the Copernicus High Resolution Snow & Ice Monitoring Service ([https://land.copernicus.eu/pan-](https://land.copernicus.eu/pan-european/buophysical-parameters/high-resolution-snow-and-ice-monitoring)  
138 [european/buophysical-parameters/high-resolution-snow-and-ice-monitoring](https://land.copernicus.eu/pan-european/buophysical-parameters/high-resolution-snow-and-ice-monitoring)) was used. To analyse long-  
139 lasting snow among the coarse blocks on the rock glacier surface, miniature temperature sensors were installed  
140 to monitor snowmelt based on ground surface temperature (GST). Three temperature data loggers (Geotest  
141 AG, Switzerland) were installed on the rock glacier surface (*G-1*, *G-2*, and *G-3*, Fig. 1b), measuring GST up  
142 to a depth of 10 cm. Data were acquired every two hours between August 2018 and August 2021. The melt-  
143 out date of snow was calculated based on the daily standard deviation of GST, following Schmid et al. (2012).

144

### 145 **3.2 Discharge measurements and water temperature**

146 Discharge measurements were performed using NaCl as a chemical tracer (Hudson and Fraser, 2008).  
147 Tracer tests were carried out two times in 2019 (11<sup>th</sup> July and 3<sup>rd</sup> October), and approx. on a bi-weekly basis  
148 during the field summer seasons 2020 (3<sup>rd</sup> July 2020 – 30<sup>th</sup> September, 7 observations) and 2021 (9<sup>th</sup> July –  
149 14<sup>th</sup> October, 9 observations). One injection point and three downstream measuring points were set (Fig. 1b).  
150 The downstream points were chosen to quantify the hydrological contribution from the rock glacier to the  
151 entire catchment runoff. Specifically, the point *Up* was located downstream of the injection point and upstream  
152 of the rock glacier, along the main stream; it was selected to quantify the water flow from the upstream  
153 catchment. The point *Up+White* was placed downstream of the White lobe tributary, in the main stream, to  
154 quantify the water flow given by the sum of the White lobe and *Up* site contributions. The third measuring  
155 point, *Down* (downstream), enclosed all the contributions from upstream (i.e., White and Black lobes, and *Up*  
156 site). In these selected points, three OTT-CTD probes (OTT HydroMet) were installed in 2020 and 2021 to  
157 continuously record physical and chemical parameters (water pressure, electrical conductivity - EC, and  
158 temperature) every 30 minutes throughout the investigated period and every 10 seconds only when the tracer



159 test was performed. A fourth probe was placed in the spring at the front of the Black lobe (*Black spring*, Fig.  
160 1b). The White lobe spring (*White spring*) was not instrumented due to the very low superficial flow rate and  
161 tortuosity among the boulders, which did not allow the proper positioning of the probe. Several problems  
162 affected the installed OTT-CTD probes, mainly due to the low water level and occasionally fine sediment  
163 deposition, often resulting in erroneous readings. The water temperature data series was the only one that  
164 provided usable, continuous values, although not covering entirely the periods of investigation. Therefore, the  
165 discharge patterns were investigated only using the salt dilution approach (discrete measurements) and the  
166 chemical characteristics were analysed only based on grab samples.

167

### 168 3.3 Tracer tests

169 Tracer tests were performed to investigate the internal hydrological mechanisms and the potentially  
170 different storage and flow dynamics of the two lobes. The injections were carried out in a debris covered zone  
171 and in absence of snow-cover. Tynopal and fluorescein were injected on the surface of the rock glacier at the  
172 injection points *IT* and *IF* (Fig. 1c) on 23<sup>rd</sup> July 2021, at 11.00 am and 11.40 am, respectively. *IT* was injected  
173 at ca. 2710 m a.s.l. on the White lobe, and *IF* at 2735 m a.s.l. on the Black lobe. 200 g of tynopal and 50 g of  
174 fluorescein were injected and rinsed using 1000-l water tanks at the *IT* and *IF* points, respectively. After the  
175 tracer injections, the springs were monitored using activated charcoal and a field fluorometer capable to  
176 measure, in continuous, three tracers (Tynopal CBS-X, Na-fluorescein and any molecule in the rhodamine  
177 family) and turbidity (Schnegg, 2002).

178 Activated charcoal (10 g) was inserted in perforated bags, allowing water to seep through them. The dye  
179 tracer in the seeping water is absorbed cumulatively over the period of the bag immersion in water (Winkler  
180 et al., 2016), providing evidence of the dye tracer passage at that point. Five sampling points were selected to  
181 place the activated charcoal bags (Fig. 1c). For the duration of the experiment (from 16<sup>th</sup> July to 6<sup>th</sup> September  
182 2021), two activated charcoal bags were placed at each sampling point. One activated charcoal remained  
183 throughout the experiment (captor C, 16<sup>th</sup> July – 6<sup>th</sup> September 2021) while the second (captor A, 16<sup>th</sup> July –  
184 23<sup>rd</sup> July 2021) was replaced with another one (captor B, 23<sup>rd</sup> July – 2<sup>nd</sup> August 2021) on the day of the dye  
185 tracer injections (23<sup>rd</sup> July). The extraction was performed by placing 1 g of activated charcoal in 10 ml of a

186 5% solution of KOH in methanol. The resulting eluate was then injected into the measuring cell of the  
187 fluorometer for dye tracer detection.

188 Differently from the analysis of the activated charcoal bags, which allowed a qualitative interpretation, the  
189 field fluorometer enabled the quantification of the amount of dye tracer flowing at the downstream point (Fig.  
190 1c). A GGUN-FL24 fluorometer was used, with a 5 min and 1 min recording rate in the week before injection  
191 (16<sup>th</sup> – 23<sup>rd</sup> July 2021) and after injection (23<sup>rd</sup> July – 2<sup>nd</sup> August 2021), respectively. The dye tracer recovery  
192 estimation was performed by integrating through time the measured concentration values multiplied by the  
193 discharge (Gaspar, 1987). The recovered mass of fluorescein was estimated from the fluorometer  
194 measurements using the QTRACER2 software that solves the equations from user-generated data input files  
195 through integration of consolidated hydraulic models (U.S. EPA, 2002). The recovered tracer concentration  
196 was estimated for the week following the injection, i.e., from 23<sup>rd</sup> July to 2<sup>nd</sup> August 2021. Only night data  
197 were considered (from 8pm to 6am) since a malfunction in the light radiation shield prevented the use of the  
198 complete data series.

199

### 200 **3.4 Volumetric rock glacier variations**

201 Measuring the actual contribution from permafrost ice melt to the water budget of rock glaciers has never  
202 been done (to the best of our knowledge) due to the extreme difficulties posed by the rock-glaciated  
203 environmental conditions. Here, the volumetric variations of the Gran Sometta Rock Glacier were used as a  
204 proxy for the interannual ice changes (cf., Halla et al., 2021). The surface changes (horizontal and vertical)  
205 were detected with unmanned aerial vehicle surveys on a yearly basis, around mid-August each year from  
206 2018 to 2021 (Bearzot et al., 2022). The Structure-from-Motion technique (Westoby et al., 2012) was used to  
207 generate dense point clouds of the rock glacier area and to produce orthomosaics and Digital Surface Models  
208 (DSMs) through the Agisoft Metashape software (v. 1.5.5 and later versions) (further details in Supplement,  
209 Text S1 and Tab. S1). DSMs, orthomosaics generation, point clouds comparison, and systematic error  
210 assessment were analysed in Bearzot et al. (2022). After image alignment, all photogrammetric blocks (one  
211 for each year) were processed with the same parametrisation to keep constant the impact on the final outputs.  
212 Based on the multitemporal analysis of the annual DSMs, the annual volumetric changes were computed. The

213 mean, standard deviation (stdev), and residual errors of the elevation change rates were computed on stable  
214 areas outside the rock glacier, over the three time intervals (2018-2019, 2019-2020, and 2020-2021).

215

### 216 3.5 Chemical and isotopic analyses

217 To understand the chemical influence of the rock glacier discharge on the main stream, assess the  
218 hydrochemical differences between the two lobes, and disentangle the contribution of different water sources  
219 to catchment water, EC, pH, major ions ( $\text{SO}_4^{2-}$ ,  $\text{HCO}_3^-$ ,  $\text{NO}_3^-$ ,  $\text{Cl}^-$ ,  $\text{PO}_4^{3-}$ ,  $\text{NO}_2^-$ ,  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{Na}^+$ ,  $\text{K}^+$ ,  $\text{NH}_4^+$ ),  
220 trace elements (Al, Ti, Cr, Mn, Mo, Co, Ni, Cu, Zn, As, Cd, Sn, Sb, Ba, Pb, Fe, V, Ag, Se, Hg), and water  
221 isotopes ( $\delta^{18}\text{O}$ ,  $\delta^2\text{H}$ ,  $^3\text{H}$ ,  $\delta^{13}\text{C}_{\text{DIC}}$ ) were analysed. Water samples for EC, pH, major ions, and trace elements  
222 were collected during three consecutive seasons, 2019-2020-2021 (June-July – October), approximately on a  
223 bi-weekly basis; water samples for isotopic analyses were collected only in 2021 (Fig. 1d). Samples were  
224 collected along the main stream, at sites *Up* and *Down*, and at the rock glacier springs, *White spring* and *Black*  
225 *spring*. Precipitation and snow samples were collected only for isotopic analyses, in 2021. For the precipitation,  
226 a collector was installed close to the Black lobe front and sampled at the same time of the chemical sampling  
227 (if precipitation occurred). One snow sample was collected on 9<sup>th</sup> July 2021 from a long-lasting snow patch;  
228 snowmelt water dripping from the snow patch was collected before interacting with the ground. Through the  
229 three-year period, 5 sampling surveys were performed in June-July, 8 in August, and 10 in September-October.  
230 Samples for  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$  were collected at all sampling dates and sites, whereas selected dates and sites were  
231 chosen for  $^3\text{H}$  and  $\delta^{13}\text{C}_{\text{DIC}}$ . An overview of the analysed parameters for each sampling date and site is reported  
232 in Tab. S2.

233 pH was measured using a GLP21 (Crison Instruments). Electrical conductivity (EC) at +20 °C was  
234 measured using a Metrohm 712 Conductometer.  $\text{HCO}_3^-$  was analysed with titration method. The  
235 concentrations of major ions were determined by ion chromatography, using a Dionex ICS1000 for cations  
236 and an Aquion for anions. Trace elements were determined by inductively coupled plasma-sector field mass  
237 spectrometry (Varian 720). Hg was determined using a DMA80 (Milestone). Analytical precision for major  
238 anions was < 10%, and for major cations and trace elements was < 5%. Charge balance error for each sample  
239 was always below 10%. Further information about the field sampling, sample treatment, and laboratory

240 analyses is reported in Supplement, Text S2, whereas the limit of detection (LOD) and quantification (LOQ)  
241 of each analyte are reported in Tab. S3.

242  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$  were analysed with a Cavity Ring Down Spectroscopy (Model L2130-I, Picarro). Isotopic  
243 composition was expressed as a  $\delta$  (per mil) ratio of the sample to the Vienna Standard Mean Ocean Water  
244 (VSMOW), where  $\delta$  is the ratio of  $^{18}\text{O}/^{16}\text{O}$  and  $^1\text{H}/^2\text{H}$ . Analytical precision was 0.1 ‰ for  $\delta^{18}\text{O}$  and 1 ‰ for  
245  $\delta^2\text{H}$ .  $^3\text{H}$  samples were analysed through Liquid Scintillation Counting (LSC), using Hidex 300SL, Tri-Carb-  
246 2500TR, and Tri-Carb-3100TR. The  $^3\text{H}$  LOD was 0.8 TU (tritium units).  $\delta^{13}\text{C}_{\text{DIC}}$  (ratio of  $^{12}\text{C}/^{13}\text{C}$  of dissolved  
247 inorganic carbon-DIC) was analysed through gas chromatography combustion isotope ratio mass spectrometry  
248 (GC-IRMS), using a GC-Isolink II IRMS system, Trace 1310, Conflow4 - IRMS Delta V Plus (Thermo Fisher  
249 Scientific). Analytical precision was < 0.5‰. Further details about the isotopic analyses and calibrations are  
250 reported in Supplement, Text S2. All isotopic analyses were performed by the IT2E Isotope Tracer  
251 Technologies Labs S.r.l.

252 To understand the influence of the rock glacier outflow on stream chemistry, differences among the  
253 sampling sites *Up*, *White spring*, *Black spring*, and *Down* were evaluated with the non-parametric one-way  
254 ANOVA (Kruskal-Wallis Test) using Wilcoxon scores, since several data series were non-normally distributed  
255 (Shapiro-Wilk Test) and/or variances were significantly uneven between groups (Levene's Test), even after  
256 data transformation. In addition, a simple two-component mixing-analysis was performed using EC, in order  
257 to assess the relative contribution of sites *Up* and  $\text{RG}_{\text{Schem}}$  (average values between *White spring* and *Black*  
258 *spring*) to *Down* (cf., Haag et al., 2000; Penna and van Meerveld, 2019). Findings of this last analysis are  
259 intended to further support the hydrological investigations.

260 A Principal Component Analysis (PCA), applying variable standardisation (scaling and centring), together  
261 with a correlation analysis (Spearman) were performed to visualise the differences in the relations among the  
262 water parameters characterising the springs at the two rock glacier lobes. In addition, based on the melt-out  
263 date of snow (obtained from the GST data elaboration), the variable “Days after snowmelt” was calculated for  
264 each sampling date, at the *White spring* and *Black spring* sites. This was used as a proxy to describe the seasonal  
265 patterns of water parameters while accounting for the contrasting end of the snowmelt period among different  
266 years (cf., Brighenti et al., 2021b).

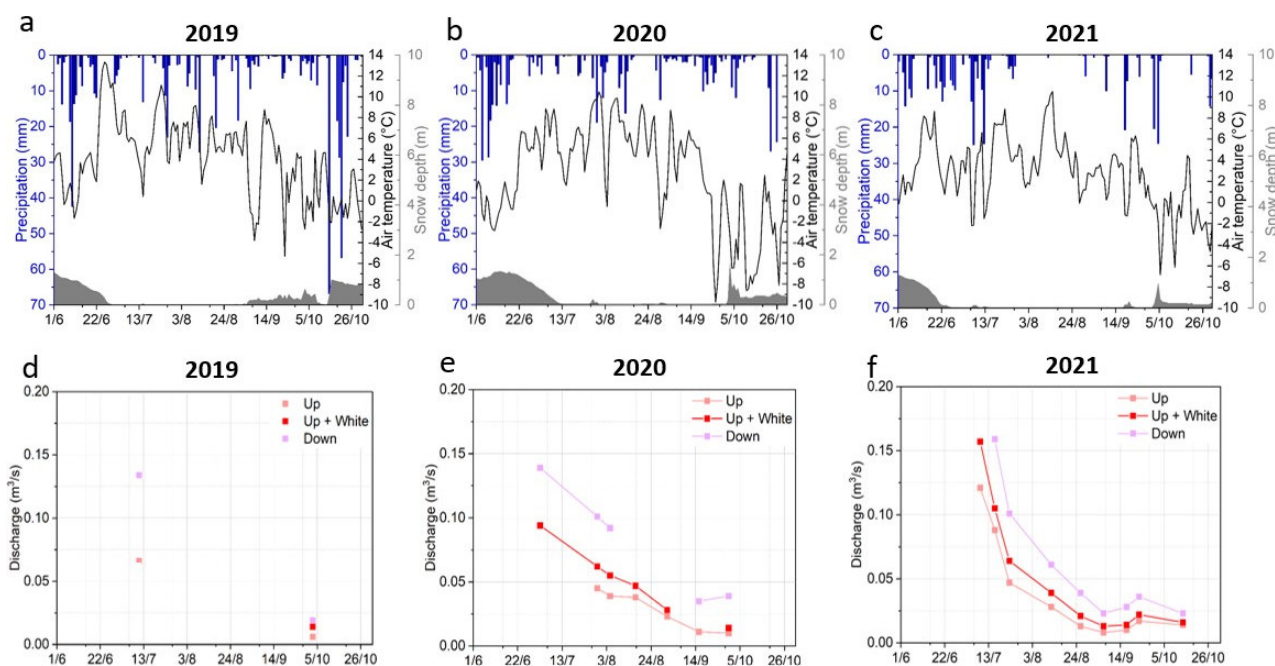
267 All analyses were implemented in the R software (R Development Core Team, 2022), at a significance  
268 level of  $p < 0.05$ .

269

## 270 **4. Results**

### 271 **4.1 Meteorological conditions and snow-cover duration**

272 The meteorological conditions and the snow-cover duration during the investigated periods are shown in  
273 Fig. 2. To focus our analyses on the months that generally had positive air temperature and scarce or absence  
274 snow cover, and to simplify the presentation of the data, only the meteorological and snow-cover  
275 characteristics from 1<sup>st</sup> June to 31<sup>st</sup> October were considered. The meteorological conditions during the  
276 investigated period were rather similar, although a moderately warmer mean air temperature was recorded in  
277 2019 (+4.1 °C), with respect to 2020 (+2.5 °C) and 2021 (+2.6 °C). The cumulative liquid precipitation during  
278 the snow-free periods was 190 mm in 2019, 186 mm in 2020, and 189 mm in 2021. The fractional snow cover  
279 in the catchment, derived from remote sensing analysis, was less than 10% in mid-July and decreased to zero  
280 in late July-early August, with no relevant difference between the three investigated years (Fig. S1). The melt-  
281 out date of snow (obtained from the GST) occurred earlier in 2021 (25<sup>th</sup> June), with respect to 2019 (28<sup>th</sup> June)  
282 and 2020 (11<sup>th</sup> July).



283

284 Figure 2. Daily values of total precipitation (blue bars), air temperature (black line), and snow depth (grey area) in 2019 (a), 2020  
 285 (b), and 2021 (c) at the Cime Bianche AWS. Discharge rates estimated at the sites Up, Up+White, and Down in 2019 (d), 2020 (e) and  
 286 2021 (f). Data gaps in 2020 (e) are due to sensor malfunction.

287

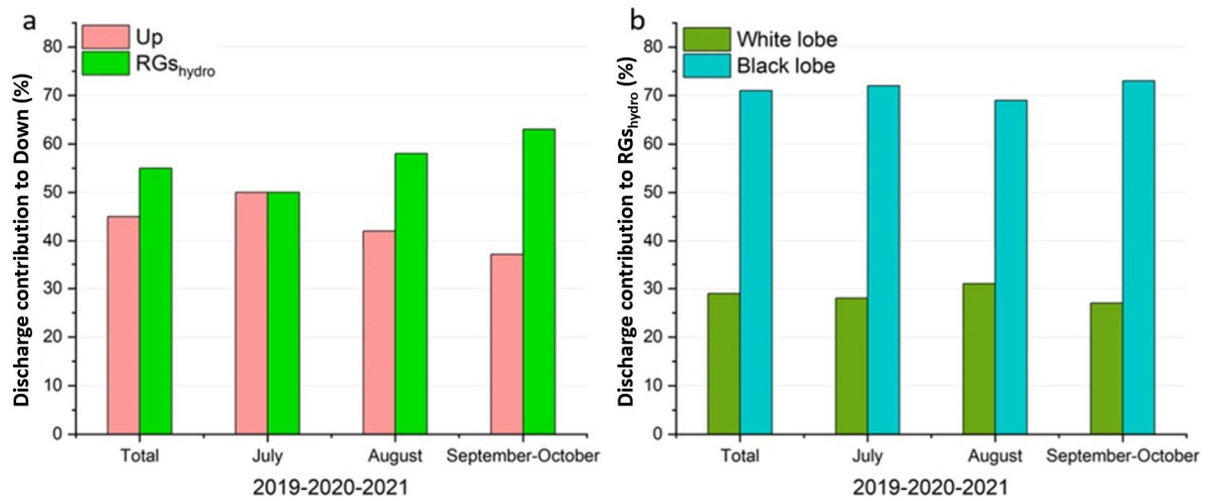
## 288 4.2 Hydrological dynamics and water sources

### 289 4.2.1 Discharge patterns

290 The results of the discharge measurements are shown in Fig. 2. In 2019, 2020, and 2021, the maximum  
 291 discharge rates occurred at *Down* site in July, with the highest value recorded in 2021 ( $0.16 \text{ m}^3 \text{ s}^{-1}$ ). From July  
 292 onwards, discharges decreased at all three monitoring sites (*Up*, *Up+White*, and *Down*), to reach their  
 293 minimum values in September-October. The lowest value,  $0.019 \text{ m}^3 \text{ s}^{-1}$ , was measured in October 2019 at site  
 294 *Up*.

295 Given the fact that the discharge measurements were discrete, single measurement values were aggregated  
 296 on a monthly basis (July, August, and September-October) in order to provide an overview of the seasonal  
 297 evolution of the rock glacier hydrological contribution to the Pousset stream over the entire research period.  
 298 The relative  $\text{RG}_{\text{hydro}}$  contribution (composed of White + Black lobe contributions) to *Down* site showed a  
 299 gradual increase during the snow-free season, from 50% in July to 58% in August, finally peaking in  
 300 September-October at 63%; the overall relative hydrological contribution of  $\text{RG}_{\text{hydro}}$  was 55% (Fig. 3a). The

301 Black lobe provided a higher relative contribution (71%), with respect to the White lobe, with rather stable  
 302 values across the snow-free season (Fig. 3b).



303  
 304 *Figure 3. Total and monthly relative discharge contributions (%), over the entire research period (2019-2020-2021), of Up and*  
 305 *RGS<sub>hydro</sub> (Black + White lobes) to the site Down (a) and of White and Black lobes to RGS<sub>hydro</sub> (b).*

306

#### 307 4.2.2 Response of the dye tracers

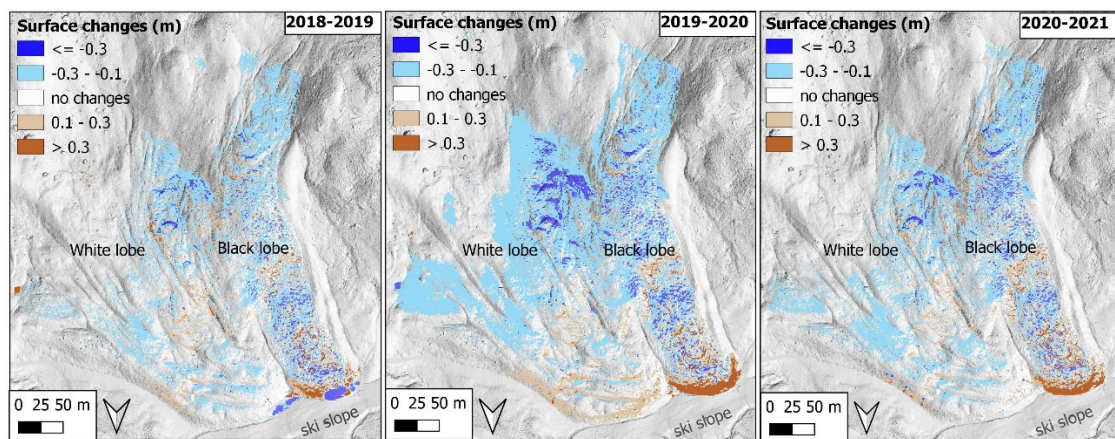
308 Fluorescein was tracked with a strong signal by two activated charcoal bags, while tynopal was not detected  
 309 at any of the control points (Figs. 1c, S2). In the week following the injection (23<sup>rd</sup> July – 2<sup>nd</sup> August 2021,  
 310 captors B), two activated charcoal bags detected the fluorescein at *Black spring* (n°3 in Fig. 1c) and *Down* (n°5  
 311 in Fig. 1c). This was confirmed by the analysis of the activated charcoal bags kept on site throughout the  
 312 campaign, i.e., from 16<sup>th</sup> July to 6<sup>th</sup> September 2021 (captors C). These results also showed that there was no  
 313 water exchange between White and Black lobes. In addition, since the activated charcoal bag on the left-hand  
 314 orographic site of the Black lobe (n°4 in Fig. 1c) did not track fluorescein, it is reasonable to assume that this  
 315 lobe does not release water laterally but only frontally, discharging directly (and only) into the Pousset stream.

316 During the sampling period, the field fluorometer at the site *Down* did not detect the transit of the tynopal  
 317 tracer but only the fluorescein. Fluorescein arrival was detected 32 hours after the injection, although this  
 318 transit time is likely overestimated since only night data were considered. In the week following the injection,  
 319 a tracer recovery rate of 7.2% was estimated (3.6 g out of 50 g injected were recovered). The estimated mean  
 320 tracer velocity was 3.4 m hour<sup>-1</sup>.

321

### 322 4.2.3 Rock glacier volumetric variations

323 The negative and positive rock glacier surface elevation changes were predominantly related to the ridge  
324 and furrow topography. The highest amount of these variations matched with high horizontal displacements  
325 in the Black lobe toe. Conversely, smaller elevation changes occurred in the White lobe, where slower  
326 movements took place. The mean horizontal and vertical errors were within the order of  $-0.44$  and  $0.24$  cm  
327 (stdev ranges  $0.45$ – $0.74$ ) and between  $-3.00$  and  $9.44$  cm (stdev ranges  $7.84$ – $12.61$ ), respectively. The relative  
328 residual errors were  $0.70$ ,  $0.76$ , and  $0.71$  cm for 2018–2019, 2019–2020, and 2020–2021, respectively. The  
329 uncertainty of image coregistration was also quantified along several profiles over the same stable terrain  
330 outside the rock glacier and the mean difference between the DSMs between  $-0.03$  m and  $0.03$  m. Although  
331 these areas are considered stable, possible microtopography alteration could occur due to weathering and snow  
332 cover presence. The mean annual volumetric changes of the rock glacier surface were  $-70$  mm yr<sup>-1</sup>,  $-124.6$   
333 mm yr<sup>-1</sup>, and  $-53.9$  mm yr<sup>-1</sup> in 2018–2019, 2019–2020, and 2020–2021 (Fig. 4), respectively. Since only two  
334 discharge measurements were performed in 2019, only the time intervals 2019–2020 and 2020–2021 were  
335 analysed in conjunction with rock glacier discharge rates (details in discussion).



336  
337 Figure 4. Surface vertical changes of the Gran Sometta rock glacier in 2018–2019, 2019–2020, and 2020–2021.

338

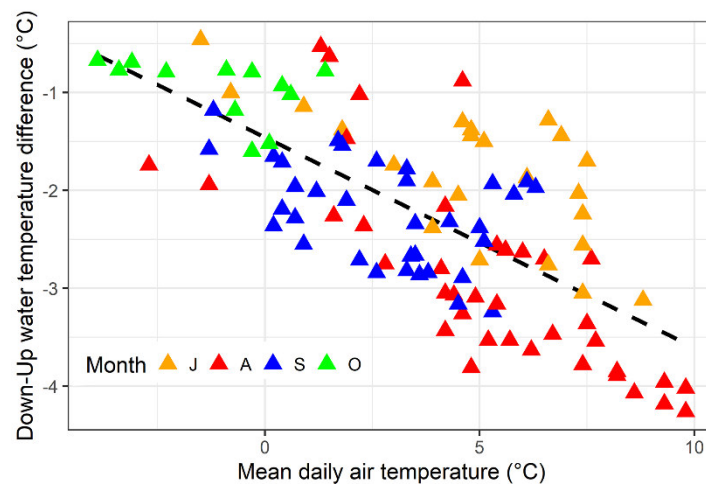
### 339 4.3 Thermal, chemical and isotopic characteristics of water

#### 340 4.3.1 Water temperature

341 The mean daily water temperatures at sites *Up*, *Up+White*, *Black spring*, and *Down* are shown in Fig. S3.  
342 The *Black spring* temperature remained constant throughout the investigated period at  $+0.5 \pm 0.1$  °C. This  
343 contributed to cooling down the Pousset stream; indeed, the mean temperature difference between *Down* and

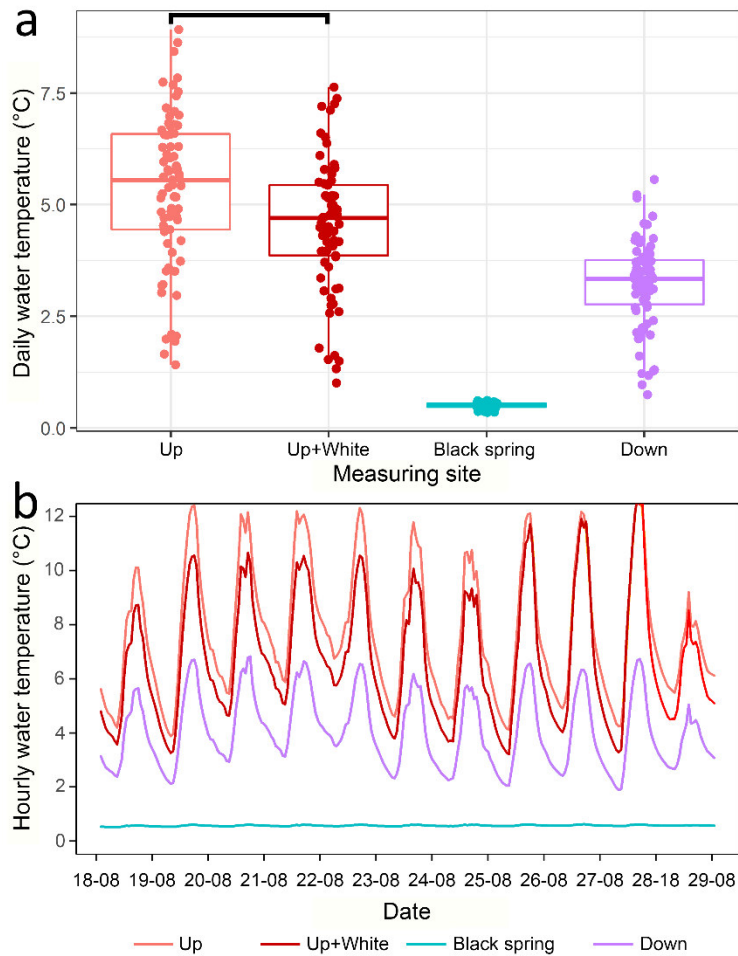


344 *Up* sites was  $-2.2 \pm 1.0$  °C. On average, the maximum temperature difference between *Down* and *Up* sites was  
 345 observed in August ( $-2.9 \pm 1.0$  °C), while the minimum temperature difference was observed in October  
 346 ( $-1.0 \pm 0.3$  °C). To investigate the cooling effect of the rock glacier springs, the relationship between the mean  
 347 daily water temperature difference between *Down* and *Up* and the mean daily air temperature was explored.  
 348 The results showed that a higher mean daily air temperature corresponded to a higher difference in water  
 349 temperature ( $r = -0.71$ ,  $p < 0.05$ ; Fig. 5).



350  
 351 *Figure 5. Scatterplot between the mean daily air temperature (Cime Bianche AWS) and the mean daily water temperature difference*  
 352 *between Down and Up sites ( $r = -0.71$ ,  $p < 0.05$ ). Month: J = July, A = August, S = September, and O = October.*

353  
 354 The Black lobe caused the strongest cooling effect on the stream (difference between *Down* and *Up*:  $-2.2$   
 355 °C; Fig. 5a), with *Up+White* and *Up* that were not significantly different, despite a decrease in stream water  
 356 temperature was also evident at *Up+White* site (difference between *Up+White* and *Up*:  $-0.8$  °C; Fig. 6a). The  
 357 water temperature variance also decreased from *Up* to *Down*, both on seasonal (*Up*:  $2.9 \pm 1.7$  °C, *Down*:  $0.9 \pm 1.0$   
 358 °C) and daily bases (*Up*:  $6.5 \pm 2.6$  °C, *Down*:  $2.5 \pm 1.6$  °C), with the strongest thermal smoothing effect exerted  
 359 by the Black lobe (Fig. 6b).



360

361 Figure 6. (a) Boxplots (with jittered data points) showing the mean daily water temperature at the four measuring sites *Up*, *Up+White*,  
 362 *Black spring*, and *Down* during all investigated period; the black segment indicates the sites that are not significantly different ( $p >$   
 363  $0.05$ ). (b) Hourly water temperature at sites *Up*, *Up+White*, *Black spring*, and *Down* between 18<sup>th</sup> and 29<sup>th</sup> August 2020 (here taken  
 364 as example).

365

### 366 4.3.2 Water chemistry

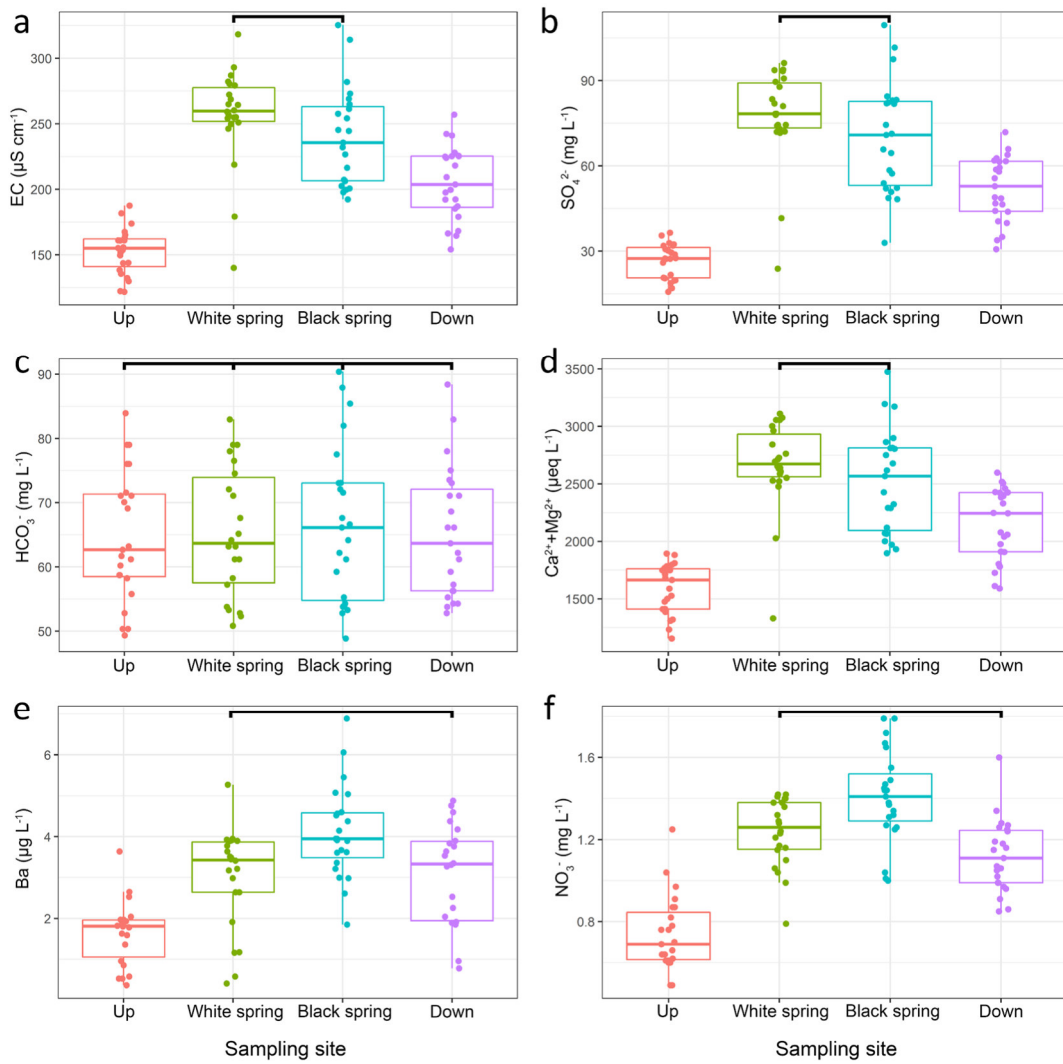
367 A statistical summary of water chemistry for all sampling sites is reported in Tab. S4. On average, pH was  
 368 basic (ca. 8) at all sampling sites.  $\text{SO}_4^{2-}$ ,  $\text{HCO}_3^-$ ,  $\text{Ca}^{2+}$ , and  $\text{Mg}^{2+}$  were the dominant ions. Ba was the only trace  
 369 element consistently above LOQ for all samples, while all other trace elements were very close to or below  
 370 LOQ and LOD for most of the analysed samples, thus they were excluded from further evaluation.

371 To understand the influence of the rock glacier on stream chemistry, the following parameters were  
 372 selected: (i) EC, as an integrator of all dissolved ionic species; (ii)  $\text{SO}_4^{2-}$  and (iii)  $\text{HCO}_3^-$ , as dominant anions;  
 373 (iv)  $\text{Ca}^{2+}$  and  $\text{Mg}^{2+}$ , as dominant cations, here considered together as  $\text{Ca}^{2+}+\text{Mg}^{2+}$  (expressed in  $\mu\text{eq L}^{-1}$ ); (v)  
 374 Ba; (vi)  $\text{NO}_3^-$  as an indicator of biotic processes (cf., Colombo et al., 2019; Williams et al., 2007). The

375 differences in the parameters previously indicated are shown in Fig. 7. All parameter values and concentrations  
376 at *Down* site were significantly higher than at *Up*, except for  $\text{HCO}_3^-$ , which did not significantly differ among  
377 all sites, according to Kruskal-Wallis Test with Wilcoxon scores. EC,  $\text{SO}_4^{2-}$ , and  $\text{Ca}^{2+}+\text{Mg}^{2+}$  were the highest  
378 at *White spring* and *Black spring*, although not significantly different between them. Ba and  $\text{NO}_3^-$   
379 concentrations were the highest at *Black spring*. Considering EC as a tracer into a two-component mixing-  
380 analysis, the relative contribution of the two lobes ( $\text{RG}_{\text{Schem}}$ ) to *Down* increased from 49% in June-July to 62%  
381 in September-October; the overall  $\text{RG}_{\text{Schem}}$  relative contribution was 55% (Fig. S4).

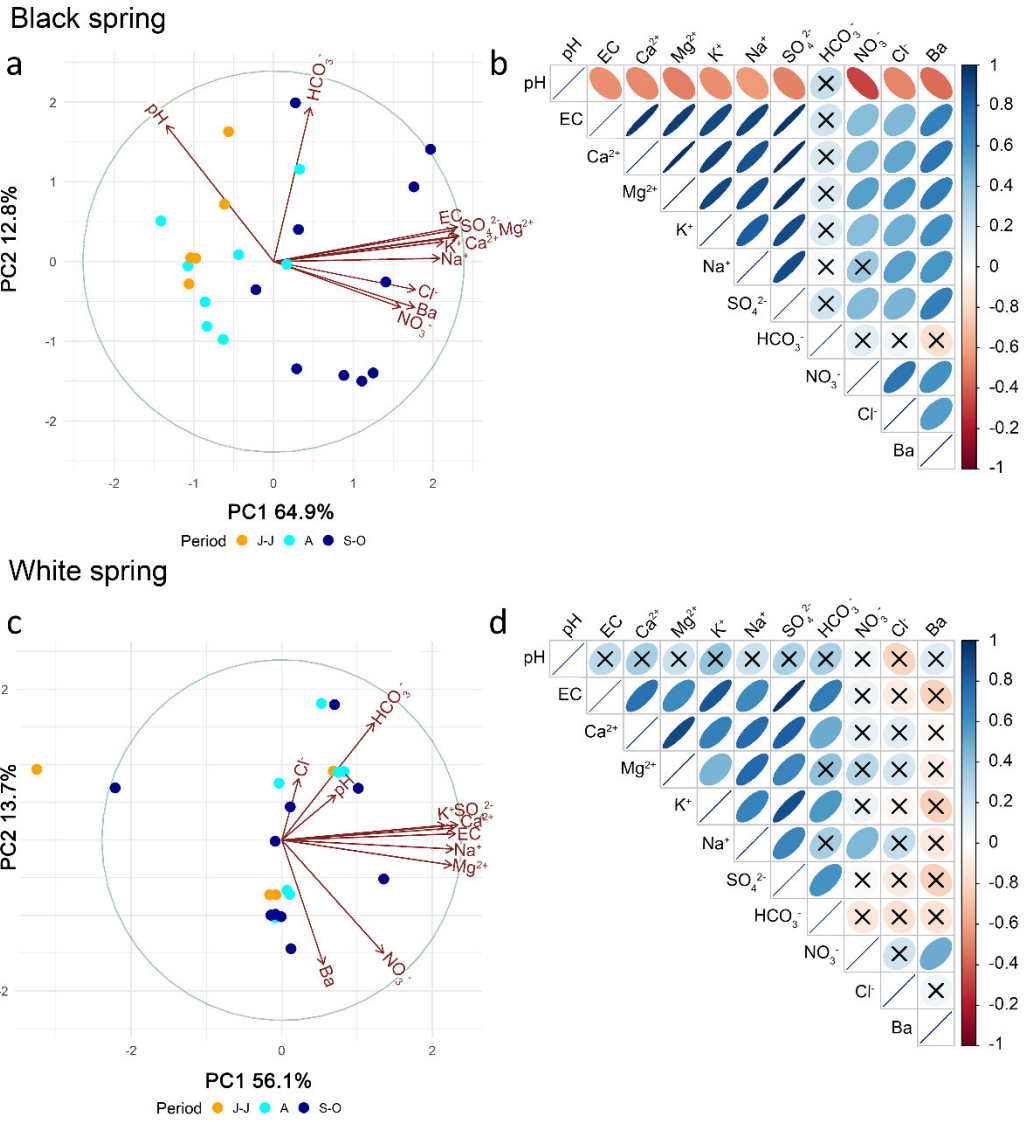
382 At the *Black spring* site, the first two axes of the PCA explained 77.7% of the total variance within the  
383 dataset (Fig. 8a). The PC 1 was primarily a gradient of EC, major ions (except  $\text{HCO}_3^-$ ), Ba, and, secondarily,  
384 pH (which also influenced the PC 2). These variables were generally positively correlated with each other,  
385 while pH was negatively correlated (Fig. 8b). Differently,  $\text{HCO}_3^-$  plotted along the PC 2 and it was not  
386 correlated to any variable. In addition, the samples collected at *Black spring* in September-October plotted  
387 almost exclusively on the right side of the first ordination, indicating that these dates were characterised by  
388 higher EC, major ions (except for  $\text{HCO}_3^-$ ) and Ba concentrations, and lower pH values, with respect to the  
389 samples collected in June-July and August. At the *White spring* site, the first two axes of the PCA explained  
390 69.8% of the total variance within the dataset (Fig. 8c). The main differences with respect to the PCA  
391 performed at *Black spring* was that at *White spring* no evident seasonal variations emerged from the data and  
392 correlations were generally weaker and even not significant (except for  $\text{HCO}_3^-$ ) (Fig. 8d).

393 The differences in seasonal variations were further confirmed by looking at the trends of the main chemical  
394 parameters (Fig. 9). Indeed, the *Black spring* exhibited evident increases in EC, major ions (except for  $\text{HCO}_3^-$   
395 ) and Ba values as a function of the days elapsed after the snowmelt period, while pH decreased, instead.  
396 Differently, the *White spring* did not show significant seasonal trends, except for  $\text{Mg}^{2+}$ , which slightly  
397 increased.



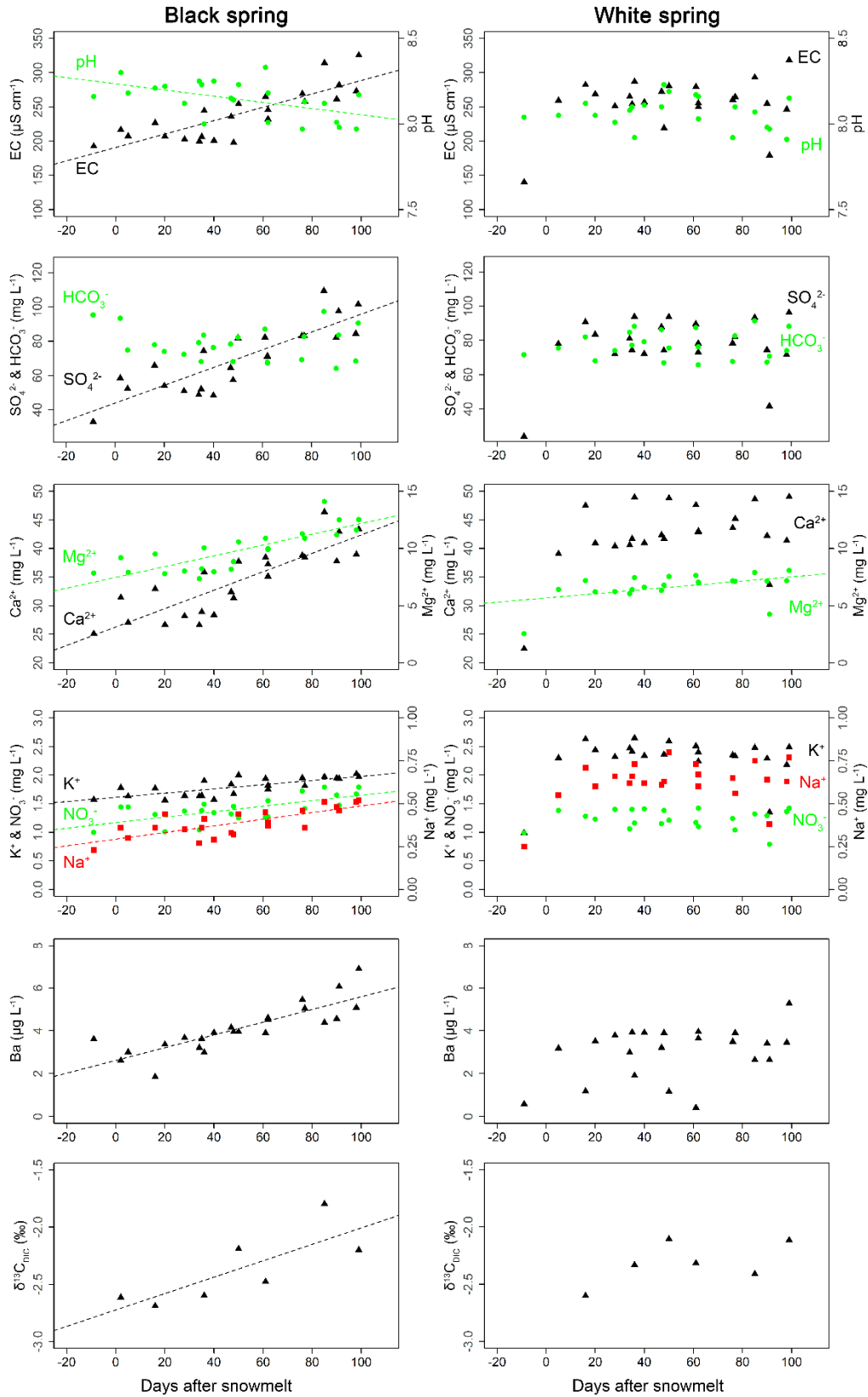
398

399 *Figure 7. Boxplots (with jittered data points) showing the differences, at the four sampling sites Up, White spring, Black spring, and*  
 400 *Down, among selected tracers: EC (a),  $\text{SO}_4^{2-}$  (b),  $\text{HCO}_3^-$  (c),  $\text{Ca}^{2+} + \text{Mg}^{2+}$  (d), Ba (e), and  $\text{NO}_3^-$  (f). The black segments indicate the*  
 401 *sites that are not significantly different ( $p > 0.05$ ).*



402

403 *Figure 8. PCA plots and correlation matrices for Black spring (a and b) and White spring (c and d). In panels a and c, J-J = June-*  
 404 *July, A = August, and S-O = September-October. In the correlation matrices in panels b and d, Spearman values are represented by*  
 405 *different colours (blue = positive correlation, red = negative correlation) and ellipse orientation (left = negative correlation, right =*  
 406 *positive correlations). Colour intensity and ellipse size represent the correlation strength (pale and wide = weak, intense, and narrow*  
 407 *= strong). Crossed ellipses correspond to correlations that are not significant ( $p > 0.05$ ).*



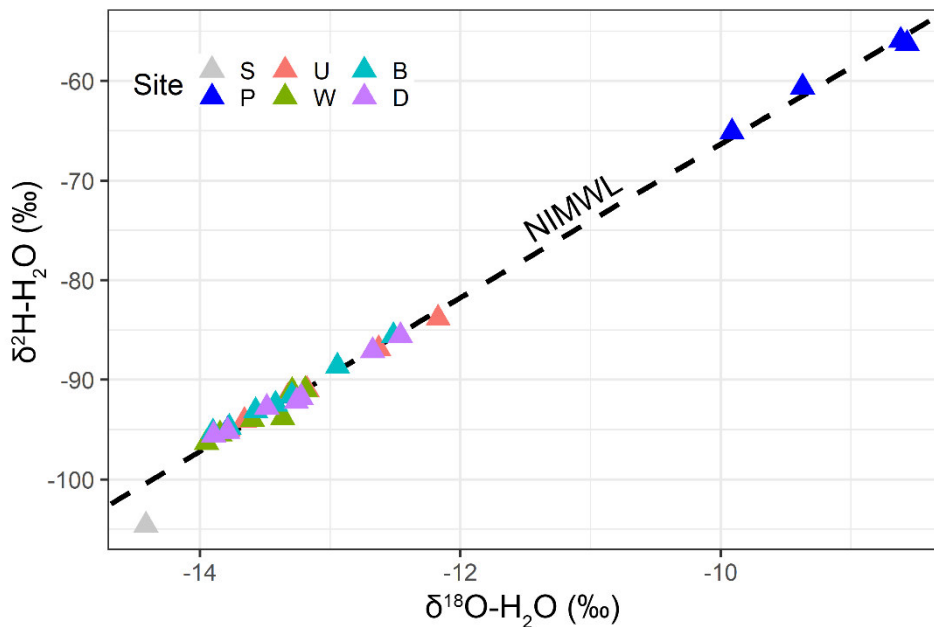
408

409 *Figure 9. Seasonal trends of water parameters as a function of the days after the melt-out date of snow (derived from GST) at Black*  
 410 *spring (left panels) and White spring (right panels). Interpolation lines represent significant linear regressions ( $p < 0.05$ ).*

411

412 **4.3.3 Water and DIC isotopic signatures**

413 The results of the isotopic analyses are presented in Tab. S4 and S5. *Precipitation* showed the most enriched  
414 values, with  $-9.1\pm 0.7\text{‰}$  for  $\delta^{18}\text{O}$  and  $59.5\pm 4.3\text{‰}$  for  $\delta^2\text{H}$ . On the opposite, the single sample for *Snow*  
415 evidenced the most depleted signal in the dataset, with values of  $-14.4\text{‰}$  for  $\delta^{18}\text{O}$  and  $-104.6\text{‰}$  for  $\delta^2\text{H}$ . *Up*  
416 and *Down* sites together with *White spring* and *Black spring* signatures arranged in between *Precipitation* and  
417 *Snow*, with a variability between  $-13.5\pm 0.3\text{‰}$  (*White spring*) and  $-13.2\pm 0.6\text{‰}$  (*Up*) for  $\delta^{18}\text{O}$ , and between  
418  $-93.6\pm 2.2\text{‰}$  (*White spring*) and  $-90.5\pm 3.9\text{‰}$  (*Up*) for  $\delta^2\text{H}$ . The stable water isotopic data are consistent with  
419 isotope compositions for the precipitation representative of the Northern Italian Meteoric Water Line  
420 (NIMWL) (Longinelli and Selmo, 2003; Fig. 10).



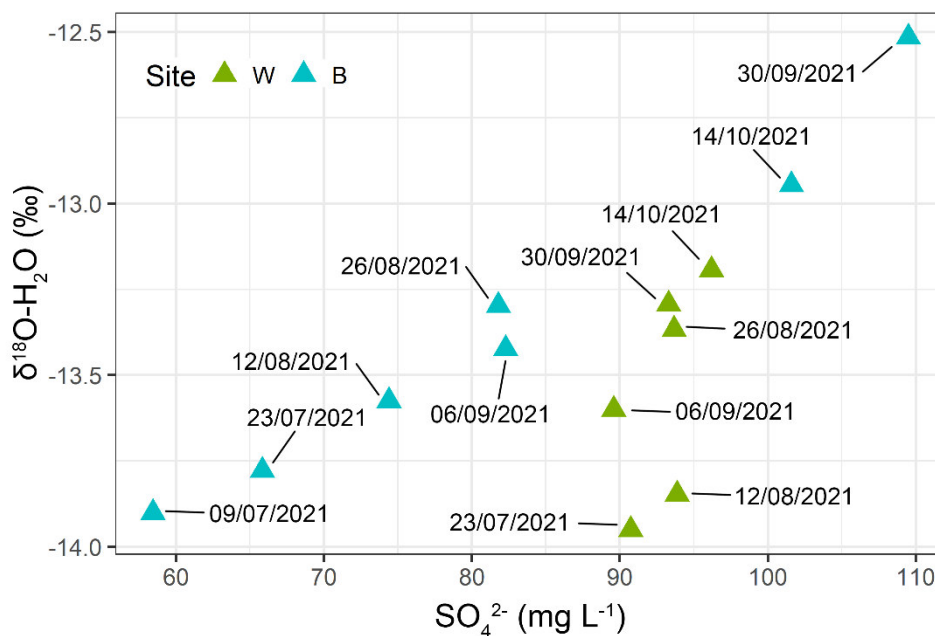
421  
422 *Figure 10. Dual-plot isotope distribution at the investigated sites, in 2021. Site: S = Snow, P = Precipitation, U = Up, W = White*  
423 *spring, B = Black spring, and D = Down. NIMWL: Northern Italian Meteoric Water Line.*

424  
425 Coupling chemistry and isotopes, it is possible to observe for the *Black spring* a strong positive correlation  
426 from July to October between  $\delta^{18}\text{O}$  and  $\text{SO}_4^{2-}$  (here used as a tracer of weathering processes in cryospheric  
427 environment; cf. Colombo et al., 2019; Salerno et al., 2016), while just a slight increment for the *White spring*  
428 occurred (Fig. 11); by specifying this, the isotope “delta”, expressed as the difference between the maximum  
429 and the minimum  $\delta^{18}\text{O}$  measured was 1.39 and 0.76‰, for the *Black spring* and *White spring*, respectively.  
430 Even *Up* and *Down* sites recorded higher delta values than the *White spring* of 1.61 and 1.44‰, respectively.

431 From the dual-plot isotope distribution, it is possible to assert how signals related to the White lobe resulted  
 432 more conservative than the ones of the Black lobe (Fig. 10), as the latter seemed to be more influenced by  
 433 liquid precipitation.

434 Results for  $\delta^{13}\text{C}_{\text{DIC}}$  ranged from  $-14.8$  to  $-10.3\text{‰}$  in *Precipitation*, while it was  $-11.7\text{‰}$  in *Snow* (Tab. S5).  
 435  $\delta^{13}\text{C}_{\text{DIC}}$  at *Up* site ranged from  $-5.6$  to  $-4.5\text{‰}$ , while in rock glacier samples ranged from  $-2.7$  to  $-1.8\text{‰}$  (both  
 436 found at *Black spring*). A significant trend toward a  $\delta^{13}\text{C}_{\text{DIC}}$  enrichment over time was registered at the *Black*  
 437 *spring*, whereas the *White spring* presented only a slight oscillatory behaviour during the investigated period  
 438 (Fig. 9).

439 Data for  $^3\text{H}$  were always above the LOQ (Tab. S5). Tritium content was measured in one July sample for  
 440 *Snow* (6.8 TU), and a single sample collection was performed at *Up* (6.0 TU), *White spring* (4.7 TU), and  
 441 *Precipitation* (6.0 TU), in September 2021. A prolonged monitoring of the *Black spring* revealed a slight  
 442 variation in  $^3\text{H}$  content, ranging between 4.8 (July and September 2021) and 7.0 TU (August 2021).



443  
 444 *Figure 11. Scatterplot between  $\text{SO}_4^{2-}$  and  $\delta^{18}\text{O}$  at sites White spring (W) and Black spring (B), in 2021.*  
 445

## 446 5. Discussion

### 447 5.1 Hydrological dynamics and water sources

#### 448 5.1.1 Rock glacier impact on hydrologic regime



449 The highest discharge rates at all measuring sites were measured at the end of the snowmelt period or right  
450 after the melt out of snow (i.e., July). However, these were not the streamflow peaks, which likely occurred  
451 during the peak snow-melt period in June (cf., Geiger et al., 2014; Krainer and Mostler, 2002), when our  
452 measurement sites were inaccessible. With the progression of the snow-free season, discharge rates  
453 significantly decreased, reaching their lowermost values in September-October, as observed in other high-  
454 elevated and rock-glaciated settings (Harrington et al., 2018; Geiger et al., 2014; Krainer and Mostler, 2002).

455 The relative contribution of the rock glacier increased towards late summer - early autumn, with the highest  
456 influence on the main stream measured in September-October (62-63%), as observed from both hydrological  
457 and chemical measurements. Despite draining only 39% of the watershed area, the Gran Sometta Rock Glacier  
458 contributed 55% of catchment streamflow during the period July – October. Therefore, our findings support  
459 the concept of these landforms as important hydrogeomorphic units which can store and transmit quantities of  
460 groundwater, modulating the catchment hydrologic response and sustaining the baseflow during the late  
461 summer - early autumn (Wagner et al., 2021a; Harrington et al., 2018). Indeed, sediment accumulation and its  
462 drainage processes in rock glaciers, together with the subsurface properties (e.g., porosity, transmissivity, and  
463 storage capacity), can impact the seasonal streamflow (Hayashi, 2020; Harrington et al., 2018). In this sense,  
464 rock glaciers have the net effect of increasing total surface runoff from alpine drainage basins, due to their  
465 capability to act as impervious surfaces (Wagner et al., 2020; Harrington et al., 2018; Geiger et al. 2014),  
466 playing a critical role as shallow groundwater reservoirs in mountain headwaters (Wagner et al., 2021a, 2020;  
467 Hayashi, 2020; Winkler et al., 2018, Geiger et al., 2014).

468 During the period July – October, the Black lobe contributed significantly more (71%) than the White lobe  
469 (29%) to catchment streamflow, with no relevant seasonal variations. It is known that different spatial  
470 distribution, depth, and thickness of permafrost affect hydrology, infiltration, and runoff response in high-  
471 mountain catchments (Rogger et al., 2017; Duguay et al., 2015; Geiger et al., 2014). Thus, it is possible that  
472 the continuous layer of frozen ground (and ice) in the Black lobe resulted in larger hydrological contributions  
473 (and possibly, runoff events) during the investigated snow-free seasons because of the more limited losses to  
474 deep groundwater storage that, vice versa, could be enhanced in the White lobe (characterised by low content  
475 and heterogeneous distribution of permafrost and ice) (cf., Rogger et al., 2017; Evans et al., 2015; Geiger et  
476 al., 2014).

477 The dye tracer tests confirmed the diversity of water circulation in the two rock glacier lobes and therefore  
478 their hydrological behaviours. At the Black lobe, the fluorescein was tracked by both the activated charcoal  
479 bags and the field fluorometer, which showed the first detection of the tracer approx. 32 hours after the  
480 injection, with a recovery rate of 7.2% in the first week after the injection. The remaining percentage of dye  
481 infiltrated from the active layer and likely penetrated in the underlying zone. Thus, different flow paths might  
482 be present in the Black lobe: a shallower flow (on the surface of the permafrost table and above the ground  
483 ice) and intra- and sub-permafrost flows (cf., Jones et al., 2019; Winkler et al., 2016; Tenthorey, 1992). The  
484 tracer mass recovered is in line with what was found in other case studies dealing with rock glaciers (e.g., Mari  
485 et al., 2013).

486 Differently from the Black lobe, at the White lobe, the passage of the tracer (tynopal) was not tracked during  
487 the investigation period. Thus, this lobe could be characterised by the prevalence of a deeper, slow groundwater  
488 flow due to major areas of unfrozen layers responsible for groundwater storage and retarded runoff (Wagner  
489 et al., 2021b). In addition, in accordance with the geophysical data collected by Bearzot et al. (2022), the  
490 injection point of the tynopal was located in a thick unfrozen zone, therefore the dye tracer might have directly  
491 infiltrated to a great depth into the debris body and slowly circulated under the permafrost, thus further  
492 explaining a high residence time (cf., Rogger et al., 2017). Another reason for the non-detection of this tracer  
493 and for the lower discharge contribution may be related to the fracturing of bedrock and the presence of bedrock  
494 depression. Since this area was characterised by a small glacier during the Little Ice Age (Bearzot et al., 2022),  
495 it is also possible that glacier erosion created hummocks and depressions in the bedrock of the watershed  
496 (Harrington et al., 2018). The fractured rocks might allow the tynopal and water for a slow drainage of  
497 groundwater stored in partially filled depressions through bedrock fractures and, as a consequence, potentially  
498 drain part of the tynopal/water outside the investigated watershed. On one hand, this last case would imply that  
499 the difference in the discharge between the two lobes may not be solely due to internal hydrological systems  
500 driven by different cryospheric conditions, although their contrasting hydrochemical patterns indicate a  
501 relevant role of permafrost and ice distribution in the two lobes (details in 5.2.2). On the other hand, this would  
502 imply an underestimation of the potential water resource that could be provided by our rock glacier, thus further  
503 highlighting its relevant role as groundwater reservoir.

504

## 505 **5.1.2 Water sources and estimated contribution from internal ice melt**

506 The distinction of different water sources composing rock glacier outflows is generally based on qualitative  
507 or indirect approaches rather than built on reliable measurements of individual system components combined  
508 with quantitative measurements. Here, different lines of evidence have been used to trying dealing with this  
509 issue.

510 The two lobes showed similar  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$  isotopic compositions, also similar to stream water, thus ruling  
511 out the hypothesis of relevant, different recharge inputs within the investigated systems. The isotopic  
512 composition of rock glacier outflows generally represents the reflection of a mixture of at least three  
513 components with different contributions over the hydrological year. In addition to precipitation and snow  
514 contributions, the third component corresponds to the groundwater baseflow input plus ice melt, whose  
515 isotopic fingerprint is not always easy to assess (Hayashi, 2020; Sileo et al., 2020). Differently, a higher  
516 variability in water stable isotopes was found at the Black lobe, as an indication of a system more prone to a  
517 local recharge influence and/or just likely more connected to the subsurface. The significant shift in the Black  
518 lobe stable isotope values reflected a seasonal pattern in accordance with concentration data interpretation,  
519 evidencing a more hydrologically responsive system (i.e., faster water circulation). A more conservative  
520 flowpath might describe the White lobe behaviour (i.e., slower water circulation). Therefore, stable water  
521 isotopes confirmed the existence of different internal hydrological systems in the two lobes, although pointing  
522 to similar recharge sources (also in the stream), thus suggesting minor ice melt inputs from both lobes,  
523 assuming ice melt having a well-distinguishable isotopic signature (cf., Williams et al., 2006).

524 Considering  $^3\text{H}$ , it represents a valuable qualitative tracer due to spike of atmospheric  $^3\text{H}$  in fallout from  
525 bomb testing in the 1950s (Celle-Jeanton et al., 2002).  $^3\text{H}$  has a half-life of 12.4 years, such that precipitation  
526 before the 1950s will contain less than 0.1 TU, whereas precipitation that fell during the decade of 1950–1960  
527 will now contain approximately 70 TU (Celle-Jeanton et al., 2002). Tritium is suitable for untangling residence  
528 times for young flowpaths (in a geological understanding), being the ideal tracer for approx. 80-year-old water  
529 systems (Kralik, 2015). Tritium data for *Precipitation* and *Snow* reflected the signal of modern precipitation  
530 monitored in central Europe (WISER, <https://nucleus.iaea.org/wiser>). Only a slight depletion in  $^3\text{H}$  content  
531 was observed for stream and rock glacier springs, if compared to precipitation data, suggesting a relatively  
532 modern water (infiltrated after the  $^3\text{H}$  peak). Considering the September 2021 sampling, when the contribution

533 of ice melt should have been at its maximum,  $^3\text{H}$  values in the rock glacier springs and the stream differed only  
534 by a maximum of 1.3 TU, possibly indicating a slight difference in residence time ( $\leq 1\text{--}2$  years, c.f., Munroe,  
535 2018). Thus, a relevant hydrological contribution to the rock glacier discharge from the melting of old  
536 permafrost ice (with age from centuries to millennia?) and/or transitional-layer ice at the top of permafrost  
537 (decades old?) was unlikely. It is also true that the seasonal (annual) active-layer ice (i.e., the superimposed  
538 ice that forms each spring due to the refreezing of part of percolating melt water) might have contributed to  
539 the rock glacier discharge. However, the lack of sensitivity in discretizing such small differences in residence  
540 time represents a limitation of the qualitative method using solely  $^3\text{H}$  data. For further discussion aimed at a  
541 fine interpretation of the ice melt apparent ages (Suckow, 2014),  $^3\text{H}$  data needs to be coupled with information  
542 of the dissolved  $^3\text{He}$  generated from  $^3\text{H}$  decay (Schlosser et al., 1988; Solomon et al., 1995), which is still  
543 poorly documented for alpine catchments (e.g., Munroe, 2018).

544 Based on the rock glacier discharge rates and volumetric variations, the overall ice melt contribution was  
545 estimated to represent less than 3.5% of the rock glacier discharge (3.3% in 2019–2020 and 2.5% in  
546 2020–2021). Estimation of ice melt contribution was not performed for the time-lapse 2018–2019; however,  
547 taking into account that the volumetric variation in 2018–2019 ( $-70\text{ mm yr}^{-1}$ ) fell within the range measured  
548 for 2019–2020 and 2020–2021 ( $-53.9 \div -124.6\text{ mm yr}^{-1}$ ), ice melt contribution in 2018–2019 did not likely  
549 diverge consistently from the presented estimations. Even assuming all ice melt occurring in September–  
550 October, when active layer depth is usually at its maximum, the net contribution of permafrost ice melt in the  
551 baseflow period would be 7% and 10% for 2019–2020 and 2020–2021, respectively. It is worth noting, though,  
552 that these values are surely overestimated since they refer to both ice and sediment losses. In this regard, due  
553 to its advancing movement, the rock glacier front is cut every year and the material removed to allow the ski  
554 slope (in winter) and the service road (in summer) to be opened (Bearzot et al., 2022).

555 In light of these findings, internal ice melt should be considered a minor contributor to the rock glacier  
556 discharge, as already reported in other rock-glaciated settings (Harrington et al., 2018; Krainer et al., 2015).

557

## 558 **5.2 Thermal and chemical influence on stream water**

### 559 **5.2.1 Water temperature**

560 The low water temperature of the Black lobe spring, near 0 °C, suggests that the water flow was in direct  
561 contact with ice or permafrost within the debris deposit (Scapozza et al., 2019; Millar et al., 2013). The rock  
562 glacier discharge lowered the average water temperature of the main stream by ca. 2.2 °C, on average. The  
563 thermal effect of the rock glacier was greatest in August, when the thermal effect of the rock glacier was  
564 strongest (mean: 2.9 °C; daily maximum: 4.3 °C). However, the cooling effect of the rock glacier decreased  
565 during periods with lower air temperature (and diminished solar radiation) and higher influence of snowmelt  
566 (e.g., early July and October), which caused upstream temperature to decrease. Our findings are in agreement  
567 with Harrington et al. (2017), who found a rock glacier spring to exert a dominant control on stream  
568 temperature during the summer months, regardless of the magnitude of the distributed energy fluxes.

569 The thermal influence of the two lobes also acted on the amplitude of the daily thermal variations in the  
570 main stream, greatly reducing them downstream. The thermal variations are generally reduced in the outflows  
571 of rock glaciers, due to the water flowing in direct contact with ice or permafrost body (Colombo et al., 2018c;  
572 Carturan et al., 2016; Millar et al., 2013; Krainer et al., 2007) and the “thermal smoothing effect” of the coarse  
573 debris cover (Harrington et al., 2017; Winkler et al., 2016).

574

### 575 **5.2.2. Water chemistry**

576 The influence of the rock glacier on the main stream was also evident from a purely chemical standpoint.  
577 Indeed, most of the solutes displayed higher concentrations at the *Down* site (after the water input from both  
578 lobes) with respect to the *Up* site. Rock glaciers are known to act as a solute-concentrating mechanism  
579 (Brighenti et al., 2019; Colombo et al., 2018c) increasing, in their waters, the concentrations of several  
580 weathering products (e.g.,  $\text{SO}_4^{2-}$ ,  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ) (Colombo et al., 2018b). This is commonly attributed to the  
581 enhanced water-rock contact, together with thawing of permafrost and the release of solute-enriched icemelt  
582 (Colombo et al., 2018b; Munroe, 2018; Williams et al., 2006).

583 Among the different solutes,  $\text{SO}_4^{2-}$  have been often found to be particularly concentrated in rock glacier  
584 waters (Steingruber et al., 2020; Rogora et al., 2020; Ilyashuk et al., 2018; Williams et al., 2006). In our case,  
585 much higher (up to three times) mean  $\text{SO}_4^{2-}$  concentrations were found at the lobes with respect to *Up* site,  
586 while no  $\text{HCO}_3^-$  enrichment was found in the rock glacier water.  $\delta^{13}\text{C}_{\text{DIC}}$  showed enriched values at the rock  
587 glacier lobes (mean: -2.3‰) with respect to stream water (mean: -5.1‰) (further details about the  $\delta^{13}\text{C}_{\text{DIC}}$

588 dynamics are reported in the Supplement, Text S3). This evidence seems to indicate a similar carbon source,  
589 although with a higher contribution from carbonate weathering by sulphuric acid (e.g., produced by sulphide  
590 oxidation) in the rock glacier (cf., Ulloa-Cedamano et al., 2021; Spence and Telmer, 2005). The S-ratio ( $\text{SO}_4^{2-}$   
591  $/(\text{SO}_4^{2-} + \text{HCO}_3^-)$ , where units of concentrations are expressed in  $\mu\text{eq L}^{-1}$ ; Tranter et al., 1997) and the  $R_{\text{SO}_4}$  (the  
592 molar ratio  $\text{SO}_4^{2-}/(\text{Ca}^{2+} + \text{Mg}^{2+})$ ; Ulloa-Cedamano et al., 2021) were 0.58 at the rock glacier and 0.34 at *Up*,  
593 on average. S-ratio and  $R_{\text{SO}_4}$  equal to 0.5 correspond to predominating carbonate weathering by sulphuric acid,  
594 while carbonate weathering by carbonic acid decreases both indices below 0.5 due to excess of  $\text{HCO}_3^-$ . S-ratio  
595 and  $R_{\text{SO}_4}$  values above 0.5 at the rock glacier indicate a  $\text{SO}_4^{2-}$  excess; a combination of sulphide oxidation and  
596 gypsum dissolution might be responsible for this occurrence (cf., Ulloa-Cedamano et al., 2021; Cooper et al.,  
597 2002). Even though there are no mapped gypsum deposits within the catchment nor were found during targeted  
598 field surveys, they might be present in the sub-surface in association with dolomites. Alternatively, the  
599 weathering of calcium silicates driven by sulphide oxidation might also contribute in causing the  $\text{SO}_4^{2-}$  excess,  
600 although considering the predominant presence of carbonatic rocks with no high silica content and the very  
601 low contribution of  $\text{Na}^+$  and  $\text{K}^+$  to the cations at all sampling sites (always  $\leq 3\%$ , on average), this process  
602 should not be relevant (cf., Steingruber et al., 2020; Spence and Telmer, 2005). However, further analyses are  
603 necessary to distinguish between specific weathering processes and discriminating between different geo-  
604 lithological sources, such as detailed lithological investigations and additional geochemical analyses (e.g.,  
605 sulphur and oxygen isotopes of sulphate as well as calcium and magnesium isotopes; e.g., Steingruber et al.,  
606 2020; Colombo et al., 2018b). Nevertheless, our findings are in agreement with previous studies which  
607 attributed the enhanced release of  $\text{SO}_4^{2-}$  from rock glaciers to different weathering processes (and their  
608 combination) in an environment with high availability of fresh mineral surfaces and moisture together with the  
609 presence of sulphide-bearing minerals (e.g., pyrite) and/or gypsum (Rogora et al., 2020; Steingruber et al.,  
610 2020; Ilyashuk et al., 2018; Williams et al., 2006).

611 Regarding the generally low concentrations of trace elements, they can be explained by the relevant  
612 presence of carbonate rocks and basic pH (cf., Colombo et al., 2019). Previous studies reported high trace  
613 element concentrations in rock glacier meltwaters, although generally associated with acid or ultrabasic  
614 metamorphic lithologies (Colombo et al., 2019; Ilyashuk et al., 2018; 2014; Thies et al., 2013). In our case,

615 only Ba was consistently above LOQ, showing an enrichment in the rock glacier outflows and downstream  
616 along the main stream.

617  $\text{NO}_3^-$  concentration was also higher at the lobes. This could be attributed to the presence of soils and  
618 vegetation in restricted areas along the main stream, with the biological community playing a relevant role in  
619 limiting losses of  $\text{NO}_3^-$  to the stream (cf., Colombo et al., 2019; Balestrini et al., 2013; Kopáček et al., 2004),  
620 and/or to microbial communities adapted to the extreme environment of the rock glacier as potential sources  
621 of nitrate in its outflow (Williams et al., 2007).

622 Evident differences in the seasonal trends of solutes, as a function of the days elapsed after the seasonal  
623 snow cover disappearance, were noted in the two lobes. Seasonal increases in solute concentrations are usually  
624 attributed to the progressive reduction of solute-diluted snowmelt (early summer) (e.g., Engel et al., 2019) and  
625 to increasing contributions of solute-enriched water fluxes when the active layer thickness is at its maximum  
626 (late summer - early autumn; Brighenti et al., 2021b; Colombo et al., 2019; 2018b; Williams et al., 2006).  
627 Considering that the two lobes are adjacent and thus have similar snowmelt duration, one would expect  
628 comparable hydrochemical behaviours. However, this was not the case. Thus, it is possible that different  
629 flowpaths, as deduced by the hydrological analyses, corresponded also to different solute sources into the two  
630 lobes. Indeed, besides the absolute variations in solute concentrations, different proxies indicated possible  
631 seasonal changes in the weathering processes in the Black lobe (which were not observed at the White lobe).  
632 Mean S-ratio and  $R_{\text{SO}_4}$  values increased from June-July (S-ratio: 0.48;  $R_{\text{SO}_4}$ : 0.51) to September-October (S-  
633 ratio: 0.63;  $R_{\text{SO}_4}$ : 0.61). In addition,  $\delta^{13}\text{C}_{\text{DIC}}$  showed a seasonal enrichment with a range from  $-2.7$  to  $-1.8\%$   
634 and  $R_{\text{SO}_4}$  and  $\delta^{13}\text{C}_{\text{DIC}}$  exhibited a significant correlation ( $r = 0.82$ ,  $p < 0.05$ ). Finally, pH seasonally decreased.  
635 These patterns point towards seasonal, relative increasing contributions from weathering processes associated  
636 with sulphide oxidation (Ulloa-Cedamano et al., 2021; Colombo et al., 2019; Williams et al., 2006). This  
637 occurrence might indicate that, in the Black lobe, where permafrost and ice content is higher (with possible  
638 greater moisture content), seasonal active layer thickening, thawing of permafrost, and melting of ice would  
639 progressively open flowpaths in the lobe interior (percolation of near-surface water into rock-ice matrix might  
640 also occur). Then, increasing chemical fluxes would originate from chemical weathering of freshly exposed  
641 mineral surfaces (further enhanced by the decreasing water flow velocity through the rock glacier system over  
642 the course of the summer) and, secondarily, from solute-enriched icemelt (cf., Colombo et al., 2018c; Williams

643 et al., 2006). Differently, in the White lobe, this change in space and time of flowpaths, causing seasonal  
644 changes in chemical weathering dynamics, would be reduced, given the lower permafrost and ice content and  
645 the possible presence of preferential flowpaths thorough unfrozen zones in the lobe interior. Lower shear stress  
646 and friction in the White lobe (characterised by slower displacement rates with respect to the Black lobe) might  
647 also play a role in lowering the production of fresh mineral surfaces inside the lobe.

648

### 649 **5.3 Environmental implications**

650 Due to climatically-driven glacial recession, glacier melt contribution to streamflow will gradually decline  
651 whereas runoff from rock glaciers will become more relevant, due to their water storage capability and longer  
652 response times to climatic changes (Wagner et al., 2021a; Jones et al., 2019). However, based on our  
653 observations and most recent findings, only a minor component of the water storage capability of rock glaciers,  
654 directly contributing to runoff, could be represented by melting ice. This is due to the slow conduction of heat  
655 through the ground, causing internal ice to melt at slow rates (Arenson et al., 2022). In addition, unlike solid  
656 water storage resources, such as glaciers and snow, the replenishable storage volume of rock glaciers is  
657 expected to increase as the degradation of permafrost (and ice melting) proceeds (Hayashi, 2020; Jones et al.,  
658 2019). Thus, rock glaciers will be an important source of water supply for both sensitive ecosystems and human  
659 consumption in the future, even with reduced hydrological fluxes originating from ice melt (Wagner et al.,  
660 2021a; Hayashi, 2020; Winkler et al., 2016).

661 Rock glaciers could play an increasingly relevant role not only by influencing timing and magnitude of  
662 runoff, but also its thermal conditions. Water temperature in mountain streams is highly sensitive to  
663 atmospheric warming (Michel et al., 2022; van Vliet et al., 2013), representing a critical factor for aquatic  
664 ecosystems (Niedrist and Füreder, 2020). The thermal state of streams in mountainous areas is expected to  
665 become even more vulnerable to warming in the near future due to the expected reductions in snow and glacier-  
666 melt inputs (Michel et al., 2020). Rock glaciers could compensate for air temperature warming trends and  
667 lower contributions of meltwater from snow and glaciers in mountain streams, especially during warm  
668 atmospheric periods. This effect could persist even after most of the ice will have melted, since low water  
669 temperatures have been measured in springs emanating from rock glaciers with low (Harrington et al., 2017)



670 to no ice content (Winkler et al., 2016), due to the physical characteristics of the coarse blocky cover  
671 (Harrington et al., 2017; Winkler et al., 2016).

672 The hydrological and thermal role of rock glaciers could be accompanied by a chemical influence. Under  
673 continued atmospheric warming, a sustained glacier recession in combination with a prolonged snow-free  
674 period may further boost the hydrochemical influence of intact rock glacier outflows. This influence will be  
675 more evident during baseflow conditions, in late summer - early autumn, when the relative hydrological  
676 contribution of rock glaciers is higher, with potentially significant effects on surface-water ecology and for  
677 drinking water quality (Brighenti et al., 2019; Colombo et al., 2018b; Ilyashuk et al., 2018; 2014; Thies et al.,  
678 2018).

679

#### 680 **5.4 Research limitations**

681 The main limitations of our approach are related to groundwater circulation and discharge analyses. Indeed,  
682 the dye tracer investigation was performed only for a relatively short amount of time during the snow-free  
683 season, therefore no information on the hydrological behaviour of the lobes was obtained for the snow-covered  
684 season. For this reason, the groundwater circulation investigation left some open considerations. For instance,  
685 the residence time of the dye tracer into the White lobe is unknown since the tynopal was not tracked during  
686 the campaign months in 2021, and further observations were not performed during the following melt  
687 season(s). Thus, the White lobe may contribute more groundwater, with respect to the Black lobe, especially  
688 during the snow-covered season, due to its internal characteristics, hence potentially representing an  
689 underestimated water source when considering the entire hydrological year. In addition, the possible presence  
690 of depressions and fractures in the bedrock would require additional investigations in other downstream springs  
691 and water bodies, potentially revealing different flow paths. Furthermore, although the seasonal variations in  
692 the relative contribution from the rock glacier agree well with the chemical analysis, thus confirming the  
693 goodness of our findings, they were not investigated continuously. Therefore, our understanding of the rock  
694 glacier hydrological behaviour is partial. For instance, the discharge response to precipitation events is  
695 unknown, and this could have been helpful in drawing a clearer picture of the internal hydrological systems of  
696 the two lobes. Finally, in the *Up* catchment, small rock glaciers, protalus ramparts, and ice-cored moraines are  
697 potentially present (as deduced from aerial and satellite images), even though targeted geomorphological

698 surveys have not been performed so far. These landforms share similar sedimentological features with the  
699 analysed rock glaciers; thus, it is possible to assume the existence of non-negligible hydrological contributions  
700 from these deposits, especially during the baseflow period. This occurrence would imply that, in other rock-  
701 glaciated catchments without these landforms, the hydrological relevance of rock glaciers might be higher than  
702 what estimated in this study.

703 Further uncertainties exist on the potential impacts of the rock glacier influence on the thermal and chemical  
704 conditions of the main stream, chiefly involving their propagation downstream, which was not investigated in  
705 our study. For instance, atmospheric conditions are the primary driver of the energy budget in mountain  
706 streams, with a dominant role of air temperature and radiation (Somers et al., 2016; Khamis et al., 2015;  
707 Magnusson et al., 2012). Therefore, a rather rapid equilibration of water temperature downstream might occur.  
708 However, the cooling effect of rock glacier discharge has been found to persist over several hundreds of metres,  
709 highlighting an overarching role of the large (negative) advective heat flux of the cold rock glacier discharge  
710 on stream energy fluxes (Harrington et al., 2017). Similarly, rock glacier discharge has been found to exert a  
711 significant influence on the downstream hydrochemistry of alpine streams, up to 3 km, highlighting its  
712 importance on the alpine river networks (Brighenti et al., 2019).

713

## 714 **6. Conclusions**

715 This study investigated the hydrological, thermal and chemical influence of an intact rock glacier on a high-  
716 elevation stream. Despite the estimated minor contribution from ice melt, the rock glacier springs sourced a  
717 disproportionately large amount of discharge to the main stream during the investigated snow-free seasons,  
718 with increasing contribution towards late summer and early autumn. This occurrence confirms the importance  
719 of rock glaciers as water sources, especially in those periods when the contribution of snowmelt is reduced or  
720 absent. In addition, the cold-water discharge from the rock glacier significantly cooled down the stream water  
721 temperature, with a larger influence during warm atmospheric periods. The chemical characteristics of the  
722 main stream were also impacted by the rock glacier discharge. Indeed, solute-concentrated waters sourcing  
723 from the rock glacier significantly increased the concentrations of most solutes in the stream. Remarkably, the  
724 two lobes forming the rock glacier, characterised by different permafrost and ice content, displayed evident  
725 contrasting hydrological and chemical behaviours. Thus, despite both lobes being intact and active, different

726 internal hydrological systems, likely driven by different cryospheric conditions, seemed to play a relevant role  
727 in shaping their hydrological and chemical response.

728 In the future, the increasing relevance of rock glaciers as water resources could be accompanied by an  
729 increasing chemical influence, with potentially significant effects on surface-water ecology and drinking water  
730 quality. At the same time, rock glaciers could play a role as thermal buffers on mountain streams, considering  
731 air temperature warming trends and lower contributions of water from melting snow and glaciers. Future work  
732 should focus on long-term hydrological and hydrochemical monitoring of rock glacier discharge and affected  
733 surface waters, assessing the downstream propagation of this “rock glacier effect”.

734

### 735 **Acknowledgements**

736 This research was funded by the Italian MIUR project Dipartimenti di Eccellenza (2018-2020) and by the  
737 Interreg V-A IT-CH 2014-2020 project Reservaqua. Nicola Colombo and Michele Freppaz were partially  
738 supported by the project NODES, which has received funding from the MUR-M4C2 1.5 of PNRR with grant  
739 agreement no. ECS00000036. We would like to thank Andrea Benech, Michela Alessio, Michel Isabellon,  
740 Paolo Proment, and Roberto Garzonio for their help in data acquisition and analysis.

741

### 742 **Reference list**

743 Arenson, L.U., Harrington, J.S., Koenig, C.E.M., et al., 2022. Mountain permafrost hydrology – a practical  
744 review following studies from the Andes. *Geosciences*, 2022, 12, 48.  
745 <https://doi.org/10.3390/geosciences12020048>.

746 Balestrini, R., Arese, C., Freppaz, M., et al., 2013. Catchment features controlling nitrogen dynamics in  
747 running waters above the tree line (central Italian Alps). *Hydrol. Earth Syst. Sci.*, 17, 989–  
748 1001. <https://doi.org/10.5194/hess-17-989-2013>.

749 Bearzot, F., Garzonio, R., Di Mauro, B., et al., 2022. Kinematics of an Alpine rock glaciers from multi-  
750 temporal UAV surveys and GNSS data. *Geomorphology*, 2022. Doi: 10.1016/j.geomorph.2022.108116.

751 Brighenti, S., Hotaling, S., Finn, D.S., et al., 2021a. Rock glaciers and related cold rocky landforms:  
752 overlooked climate refugia for mountain biodiversity. *Glob. Change Biol.*, 1-14. Doi: 10.1111/gcb.15510.

753 Brighenti, S., Tolotti, M., Bruno, M.C., et al., 2021b. Contrasting physical and chemical conditions of two  
754 rock glacier springs. *Hydrological Processes*, 35: e14159. <https://doi.org/10.1002/hyp.14159>.

755 Brighenti, S., Tolotti, M., Bruno, M.C., et al., 2019. After the peak water: The increasing influence of rock  
756 glaciers on alpine river system. *Hydrological Processes*, 33(21), 1-20. <https://doi.org/10.1002/hyp.13533>.

757 Carturan, L., Zuecco, G., Seppi, R., et al., 2016. Catchment-scale permafrost mapping using spring water  
758 characteristics. *Permafrost and Periglacial Processes*, 27, 253-270. Doi: 10.1002/ppp.1875.

759 Celle-Jeanton, H., Gourcy, L., Aggarwal, P., 2002. Reconstruction of tritium time series in precipitation. In:  
760 Study of Environmental Change Using Isotope Techniques International Conference, 2001 April 23–27,  
761 Vienna. IAEA in Austria, IAEA-CN-80/11P.

762 Colombo, N., Salerno, F., Martin, M., et al., 2019. Influence of permafrost, rock and ice glaciers on chemistry  
763 of high.elevation ponds (NW Italian Alps). <https://doi.org/10.1016/j.scitotenv.2019.06.233>.

764 Colombo, N., Sambuelli, L., Comina, C., et al., 2018a. Mechanisms linking active rock glaciers and impounded  
765 surface water formation in high-mountain areas. *Earth Surface Processes and Landforms*, 43(2), 417-431.  
766 <https://doi.org/10.1002/esp.4257>.

767 Colombo, N, Salerno, F., Gruber, S., et al., 2018b. Impacts of permafrost degradation on inorganic chemistry  
768 of surface fresh water. *Global and Planetary Change*, 162, 69-83.  
769 <https://doi.org/10.1016/j.gloplacha.2017.11.017>

770 Colombo, N., Gruber, S., Martin, M., et al., 2018c. Rainfall as primary driver of discharge and solute export  
771 from rock glaciers: the Col d’Olen rock glacier in the NW Italian Alps. *Science of Total Environment*, 639,  
772 316-330. <https://doi.org/10.1016/j.scitotenv.2018.05.068>.

773 Cooper, R.J., Wadham, J.L., Tranter, M., et al., 2002. Groundwater hydrochemistry in the active layer of the  
774 proglacial zone, Finsterwalderbreen, Svalbard. *J. Hydrol.*, 269(3-4), 208-223.  
775 [https://doi.org/10.1016/S0022-1694\(02\)00279-2](https://doi.org/10.1016/S0022-1694(02)00279-2).

776 Duguay, M.A., Edmunds, A., Arenson, L.U., et al., 2015. Quantifying the significance of the hydrological  
777 contribution of a rock glacier – A review. GEOQuébec 2015, 68th Can. Geotech. Conf. 7th Can. Permafr.  
778 Conf., CD-Rom, Québec, QC, 20-23 September 2015.

779 Dumont, B., Gascoin, S., Hagolle, O., et al., 2021. Brief communication : evaluation of the snow cover  
780 detection in the Copernicus High Resolution Snow & Ice Monitoring Service. *The Cryosphere*, 15(10),  
781 4975-4980.

782 Engel, M., Penna, D., Bertoldi, G., et al., 2019. Controls on spatial and temporal variability in streamflow and  
783 hydrochemistry in a glacierized catchment. *Hydrology and Earth System Sciences*, 23, 2041–2063.  
784 <https://doi.org/10.5194/hess-23-2041-2019>.

785 Evans, S.G., S. Ge, Liang, S., 2015. Analysis of groundwater flow in mountainous, headwater catchments with  
786 permafrost. *Water Resour. Res.*, 51, 9564-9576. Doi: 10.1002/2015WR017732.

787 Gaspar, E., 1987. *Modern trends in tracer hydrology*. Ed. Emilian Gaspar, CRC Press Boca Raton, Fla.

788 Geiger, S.T., Daniels, J.M., Miller, S.N., et al., 2014. Influence of rock glaciers on stream hydrology in the La  
789 Sal Mountains, Utah. *Arctic, Antarctic, and Alpine Research*, 46(3), 645-658.  
790 <https://dx.doi.org/10.1657/1938-4246-46.3645>.

791 Haeberli, W., Schaub, Y., Huggel, C., 2017. Increasing risks related to landslides from degrading permafrost  
792 into new lakes in deglaciating mountain ranges. *Geomorphology*, 293(B), 405-417.  
793 <https://doi.org/10.1016/j.geomorph.2016.02.009>.

794 Haeberli, W., Hallet, B., Arenson, L., et al., 2006. Permafrost creep and rock glacier dynamics. *Permafrost and*  
795 *Periglacial Processes*, 17(3), 189-214. <https://doi.org/10.1002/ppp.561>.

796 Haag, I., Kern, U., Westrich, B., 2000. Assessing in-stream erosion and contaminant transport using the end-  
797 member mixing analysis (EMMA). *The Role of Erosion and Sediment Transport in Nutrient and*  
798 *Contaminant Transfer*. IAHS Publ. no, 263.

799 Halla, C., Blöthe, J.H., Tapia Baldis, C., et al., 2021. Ice content and interannual water storage changes of an  
800 active rock glacier in the dry Andes of Argentina. *The Cryosphere*, 15, 1187–1213,  
801 <https://doi.org/10.5194/tc-15-1187-2021>.

802 Harrington, J.S., Mozil, A., Hayashi, M., et al., 2018. Groundwater flow and storage processes in an inactive  
803 rock glacier. *Hydrological Processes*, 32(20), 3070-3088. <https://doi.org/10.1002/hyp.13248>.

804 Harrington, J.S., Hayashi, M., Kurylyk, B., 2017. Influence of a rock glacier spring on the stream energy  
805 budget and cold-water refuge in an alpine stream. *Hydrological Processes*, 31, 4719-4733. Doi:  
806 10.1002/hyp.11391.

807 Hayashi, M., 2020. Alpine hydrogeology: the critical role of groundwater in sourcing the headwaters of the  
808 world. *Groundwater*, 58(4), 498-510. <https://doi.org/10.1111/gwat.12965>.

809 Hugonnet, R., McNabb, R., Berthier, E., et al., 2021. Accelerated global glacier mass loss in the early twenty-  
810 first century. *Nature*, 592, 726-732. <https://doi.org/10.1038/s41586-021-03436-z>.

811 Hudson, R., Fraser, J., 2008. Introduction to salt dilution gauging for streamflow measurement part IV: the  
812 mass balance (or dry injection) method. *Streamline Watershed Management Bulletin*, 9(1), 6-12.

813 Huss, M., Bookhagen, B., Huggel, C., et al., 2017. Toward mountains without permafrost snow and ice. *Earth's  
814 Future*, 5(5), 418-435. <https://doi.org/10.1002/2016EF000514>.

815 Ilyashuk, B.P., Ilyashuk, E.A., Psenner, R., et al., 2018. Rock glaciers in crystalline catchments: Hidden  
816 permafrost-related threats to alpine headwater lakes. *Global Change Biology*, 24(4), 1548-1562. Doi:  
817 10.1111/gcb.13985.

818 Ilyashuk, B.P., Ilyashuk, E.A., Psenner, R., et al., 2014. Rock glacier outflows may adversely affect lakes:  
819 lessons from the past and present of two neighboring water bodies in a crystalline-rock watershed.  
820 *Environmental Science & Technology*, 2014, 48, 6192-6200. <https://pubs.acs.org/doi/10.1021/es500180c>.

821 Jin, H., Huang, Y., Bense, V.F., et al., 2022. Permafrost degradation and its hydrogeological impacts. *Water*,  
822 14, 372. <https://doi.org/10.3390/w14030372>.

823 Jones, D.B., Harrison, S., Anderson, K., et al., 2018. Mountain rock glaciers contain globally significant water  
824 stores. *Scientific Reports*, 8, 2834.

825 Jones, D.B., Harrison, S., Anderson, K., et al., 2019. Rock glaciers and mountain hydrology: A review. *Earth-  
826 Science Reviews*, 193, 66-90. <https://doi.org/10.1016/j.earscirev.2019.04.001>.

827 Khamis, K., Brown, L.E., Milner, A.M., et al., 2015. Heat exchange processes and thermal dynamics of a  
828 glacier-fed alpine stream. *Hydrological Processes*, 29, 3306–3317. <https://doi.org/10.1002/hyp.10433>.

829 Kopáček, J., Kana, J., Santruckova, H., et al., 2004. Chemical and biochemical characteristics of alpine soils  
830 in the Tatra Mountains and their correlation with lake water quality. *Water Air Soil Pollut.*, 153, 307–327.  
831 <https://doi.org/10.1023/B:WATE.0000019948.23456.14>.

832 Krainer, K., Bressan, D., Dietre, B., et al., 2015. A 10,300-year-old permafrost core from the active rock glacier  
833 Lazaun, southern Ötztal Alps (South Tyrol, northern Italy). *Quaternary Research*, 83, 324-335.

834 Krainer, K., Mostler, W., Spötl, C., 2007. Discharge from active rock glaciers, Austrian Alps: a stable isotope  
835 approach. *Austrian Journal of Earth Sciences*, 100, 102-112.

836 Krainer, K., Mostler, W., 2002. Hydrology of active rock glaciers: examples from the Austrian Alps. *Arctic,*  
837 *Antarctic, and Alpine Research*, 34(2), 142-149.

838 Kralik, M., 2015. How to estimate mean residence times of groundwater. *Procedia Earth and Planetary Science*,  
839 13, 301-306. <https://doi.org/10.1016/j.proeps.2015.07.070>.

840 Longinelli, A., Selmo, E., 2003. Isotopic composition of precipitation in Italy: a first overall map. *Journal of*  
841 *Hydrology* 270, 75-88. DOI: 10.1016/S0022-1694(02)00281-0.

842 Magnusson, J., Jonas, T., Kirchner, J.W., 2012. Temperature dynamics of a proglacial stream: Identifying  
843 dominant energy balance components and inferring spatially integrated hydraulic geometry. *Water*  
844 *Resources Research*, 48, W06510. <https://doi.org/10.1029/2011WR011378>.

845 Mania, I., Gorra, R., Colombo, N., et al., 2018. Prokaryotic diversity and distribution in different habitats of  
846 an alpine rock glacier-pond system. *Microbial Ecology*, 78(1), 70-84. Doi: 10.1007/s00248-018-1272-3.

847 Mari, S., Scapozza, C., Pera, S., et al., 2013. Prove di multitracciamento di ghiacciai rocciosi e ambienti  
848 periglaciali nel Vallon de Réchy (VS) e nella Valle di Sceru (TI). *Bollettino della Società ticinese di scienze*  
849 *naturali*, 101, 13-20. ISSN 0379-1254.

850 Michel, A., Schaefli, B., Wever, N., et al., 2022. Future water temperature of rivers in Switzerland under  
851 climate change investigated with physics-based models. *Hydrol. Earth Syst. Sci.*, 26, 1063-1087.  
852 <https://doi.org/10.5194/hess-26-1063-2022>.

853 Michel, A., Brauchli, T., Lehning, M., et al., 2020. Stream temperature and discharge evolution in Switzerland  
854 over the last 50 years: annual and seasonal behaviour. *Hydrol. Earth Syst. Sci.*, 24, 115-142.  
855 <https://doi.org/10.5194/hess-24-115-2020>.

856 Millar, C.I., Westfall, R.D., Delany, D.L., 2013. Thermal and hydrologic attributes of rock glaciers and  
857 periglacial talus landforms: Sierra Nevada, California, USA. *Quaternary International*, 1-12.  
858 <https://dx.doi.org/10.1016/j.quaint.2012.07.019>.

859 Munroe, J.S., 2018. Distribution, evidence of internal ice, and possible hydrologic significance of rock glaciers  
860 in the Uinta Mountains, Utah, USA. *Quaternary Research*, 90, 50-65. <https://doi.org/10.1017/qua.2018.24>.

861 Niedrist, G.H., Füreder, L., 2020. Real-time warming of Alpine streams: (re)defining invertebrates'  
862 temperature preferences. *River Res. Applic.*, 2021, 37, 283-293. Doi: 10.1002/rra.3638.

863 Penna, D., van Meerveld, H.J., 2019. Spatial variability in the isotopic composition of water in small  
864 catchments and its effect on hydrograph separation. *Wiley Interdisciplinary Reviews Water*, 6(5).  
865 <https://doi.org/10.1002/wat2.1367>.

866 R Development Core Team, 2022. R: A language and environment for statistical computing. R Foundation for  
867 Statistical Computing, Vienna, Austria. <https://www.R-project.org/>

868 Rogger, M., Chirico, G.B., Hausmann, H., et al., 2017. Impact of mountain permafrost on flow path and runoff  
869 response in a high alpine catchment. *Water Resour. Res.*, 53, 1288-1308. Doi: 10.1002/2016WR019341.

870 Rogora, M., Somaschini, L., Marchetto, A., et al., 2020. Decadal trends in water chemistry of Alpine lakes in  
871 calcareous catchments driven by climate change. *Sci. Total Environ.*, 708, 135180.  
872 <https://doi.org/10.1016/j.scitotenv.2019.135180>.

873 Salerno, F., Rogora, M., Balestrini, R., et al., 2016. Glacier melting increases the solute concentrations of  
874 Himalayan glacial lakes. *Environ. Sci. Technol.*, 50(17), 9150–9160,  
875 <https://doi.org/10.1021/acs.est.6b02735>.

876 Scapozza, C., Deluigi, N., Bulgheroni, M., et al., 2019. Assessing the impact of ground ice degradation on  
877 high mountain lake environments (Lago Nero catchment, Swiss Alps). *Aquatic Sciences*, 82(5), 1-16.  
878 <https://doi.org/10.1007/s00027-019-0675-7>.

879 Seppi, R., Zanoner, T., Carton, A., et al., 2015. Current transition from glacial to periglacial processes in the  
880 Dolomites (South-Eastern Alps). *Geomorphology*, 228, 71-86.  
881 <https://doi.org/10.1016/j.geomorph.2014.08.025>.

882 Schmid, M.-O., Gubler, S., Fiddes, J., et al., 2012. Inferring snowpack ripening and melt-out from distributed  
883 measurements of near-surface ground temperatures. *The Cryosphere*, 6, 1127-1139. Doi: 10.5194/tc-6-  
884 1127-2012.

885 Schlosser, P., Stute, M., Dörr, H., et al., 1988. Tritium/<sup>3</sup>He dating of shallow groundwater. *Earth and Planetary*  
886 *Science Letters* 89, 353–362.

887 Schnegg, P.A., 2002. An inexpensive field fluorometer for hydrological tracer tests with three tracers and  
888 turbidity measurement. *Groundwater and Human Development*.



889 Sileo, N. R., Dapeña, C., Trombotto Liaudat, D., 2020. Isotopic composition and hydrogeochemistry of a  
890 periglacial Andean catchment and its relevance in the knowledge of water resources in mountainous areas.  
891 *Isotopes in Environmental and Health Studies*, 56(5-6), 480-494.  
892 <https://doi.org/10.1080/10256016.2020.1814278>.

893 Somers, L.D., Gordon, R.P., McKenzie, J.M., et al., 2016. Quantifying groundwater-surface water interactions  
894 in a proglacial valley, Cordillera Blanca, Peru. *Hydrological Processes*, 30, 2915–2929.  
895 <https://doi.org/10.1002/hyp.10912>.

896 Spence, J., Telmer, K., 2005. The role of sulfur in chemical weathering and atmospheric CO<sub>2</sub> fluxes: evidence  
897 from major ions,  $\delta^{13}\text{C}_{\text{DIC}}$ , and  $\delta^{34}\text{S}_{\text{SO}_4}$  in rivers of the Canadian Cordillera. *Geochim. Cosmochim. Acta*,  
898 69(23), 5441–5458. <https://doi.org/10.1016/j.gca.2005.07.011>.

899 Steingruber, S.M., Bernasconi, S.M., Valenti, G., 2020. Climate change-induced changes in the chemistry of  
900 a high-Altitude mountain lake in the central Alps. *Aquatic Geochemistry*. [https://doi.org/10.1007/s10498-](https://doi.org/10.1007/s10498-020-09388-6)  
901 020-09388-6.

902 Solomon, D. K., Poreda, R. J., Cook, P. G., et al., 1995. Site Characterization Using <sup>3</sup>H/<sup>3</sup> He Ground-Water  
903 Ages, Cape Cod, MA. *Groundwater*, 33(6), 988-996. <https://doi.org/10.1111/j.1745-6584.1995.tb00044.x>

904 Suckow, A., 2014. The age of groundwater—Definitions, models and why we do not need this term. *Applied*  
905 *Geochemistry*, 50, 222-230. <https://doi.org/10.1016/j.apgeochem.2014.04.016>

906 Tenthorey, G., 1992. Perennial névés and the hydrology of rock glaciers. *Permafrost and Periglacial Processes*,  
907 3, 247-252. [https://1045-6740/92/030247-06\\$08.00](https://1045-6740/92/030247-06$08.00)

908 Thies, H., Nickus, U., Tessadri, R., et al., 2018. Peculiar arsenic, copper, nickel, uranium, and yttrium-rich  
909 stone coatings in a high mountain stream in the Austrian Alps. *Austrian J. Earth Sci.*, 110(2).  
910 <https://doi.org/10.17738/ajes.2017.0012>.

911 Thies, H., Nickus, U., Tolotti, M., et al., 2013. Evidence of rock glacier melt impacts on water chemistry and  
912 diatoms in high mountain streams. *Cold Reg. Sci. Technol.*, 96, 77-85.  
913 <https://doi.org/10.1016/j.coldregions.2013.06.006>.

914 Tranter, M., Sharp, M.J., Brown, G.H., et al., 1997. Variability in the chemical composition of in-situ  
915 subglacial meltwaters. *Hydrologic Processes*, 11, 59–77. [https://doi.org/10.1002/\(SICI\)1099-](https://doi.org/10.1002/(SICI)1099-1085(199701)11:1<59::AID-HYP403>3.0.CO;2-S)  
916 1085(199701)11:1<59::AID-HYP403>3.0.CO;2-S

917 Ulloa-Cedamanos, F., Probst, A., Moussa, I., et al., 2021. Chemical weathering and CO<sub>2</sub> consumption in a  
918 multi-lithological karstic critical zone: Long term hydrochemical trends and isotopic survey. *Chemical*  
919 *Geology*, 585, 120567. <https://doi.org/10.1016/j.chemgeo.2021.120567>

920 van Vliet, M.T.H., Franssen, W.H.P., Yearsley, J.R., et al., 2013. Global river discharge and water temperature  
921 under climate change. *Global Environmental Change*, 23, 450-464.  
922 <https://doi.org/10.1016/j.gloenvcha.2012.11.002>.

923 Viviroli, D., Kummu, M., Meybeck, M., et al., 2020. Increasing dependence of lowland populations on  
924 mountain water resources. *Nature Sustainability*, 3, 917-928. <https://doi.org/10.1038/s41893-020-0559-9>.

925 Wagner, T., Kainz, S., Helfricht, K., et al., 2021a. Assessment of liquid and solid water storage in rock glaciers  
926 versus glacier ice in the Austrian Alps. *Science of the Total Environment*, 800.  
927 <https://doi.org/10.1016/j.scitotenv.2021.149593>.

928 Wagner, T., Kainz, S., Krainer, K., et al., 2021b. Storage-discharge characteristics of an active rock glacier  
929 catchment in the Innere Ölgrube, Austrian Alps. *Hydrological Processes*, 35.  
930 <https://doi.org/10.1002/hyp.14210>.

931 Wagner, T., Brodacz, A., Krainer, K., et al., 2020. Active rock glaciers as shallow groundwater reservoir,  
932 Austrian Alps. *Grundwasser – Zeitschrift der Fachsektion Hydrogeologie*. [https://doi.org/10.1007/s00767-](https://doi.org/10.1007/s00767-020-00455-x)  
933 [020-00455-x](https://doi.org/10.1007/s00767-020-00455-x).

934 Westoby, M.J., Brasington, J., Glasser, N.F., et al., 2012. “Structure-from-Motion” photogrammetry: A low-  
935 cost, effective tool for geoscience applications. *Geomorphology*, 179.  
936 <https://doi.org/10.1016/j.geomorph.2012.08.021>.

937 Williams, M.W., Knauf, M., Cory, R., et al., 2007. Nitrate content and potential microbial signature of rock  
938 glacier outflow, Colorado Front Range. *Earth Surface Processes and Landforms*, 32(7).  
939 <https://doi.org/10.1002/esp.1455>.

940 Williams, M.W., Knauf, M., Caine, N., et al., 2006. Geochemistry and source waters of rock glacier outflow,  
941 Colorado front range. *Permafrost and Periglacial Processes*, 17, 13-33. Doi: 10.1002/ppp.535.

942 Winkler, G., Wagner, T., Krainer, K., et al., 2018. Hydrogeology of rock glaciers – storage capacity and  
943 drainage dynamics – an overview. In: Sychev, V.G., Mueller, L. (Eds.), *Novel Methods and Results of*

- 944 Landscape Research in Europe, Central Asia and Siberia. Volume 2, FGBNU "VNII agrochimii", Moscow,  
945 Russia, pp. 329-334.
- 946 Winkler, G., Wagner, T., Pauritsch, M., et al., 2016. Identification and assessment of groundwater flow and  
947 storage components of the relict Schöneben rock glacier, Niedere Tauern Range, Eastern Alps (Austria).  
948 Hydrogeol J., 24, 937-953. Doi: 10.1007/s10040-015-1348-9.
- 949 Zemp, M., Frey, H., Gärtner-Roer, I., et al., 2015. Hystorically unprecedented global glacier decline in the  
950 early 21st century. Journal of Glaciology, 61(228), 745-762. Doi: 10.3189/2015JoG15J017.