



AperTO - Archivio Istituzionale Open Access dell'Università di Torino

Determination of photostability, biocompatibility and efficiency as photo-Fenton auxiliaries of three different types of soluble bio-based substances (SBO)

This is the author's manuscript

Original Citation:

Availability:

This version is available http://hdl.handle.net/2318/1557827 since 2017-05-27T14:20:20Z

Published version:

DOI:10.1016/j.cattod.2014.10.015

Terms of use:

Open Access

Anyone can freely access the full text of works made available as "Open Access". Works made available under a Creative Commons license can be used according to the terms and conditions of said license. Use of all other works requires consent of the right holder (author or publisher) if not exempted from copyright protection by the applicable law.

(Article begins on next page)



UNIVERSITÀ DEGLI STUDI DI TORINO

1	
2	
3	
4	
5	This is an author version of the contribution published on:
6	Questa è la versione dell'autore dell'opera:
7	[Catalysis Today 252 (2015) 177–183
8	10.1016/j.cattod.2014.10.015.]
9	The definitive version is available at:
10	La versione definitiva è disponibile alla URL:
11	http://www.sciencedirect.com/science/article/pii/S092058611400693
12	

- 13
- 14

Determination of photostability, biocompatibility and 15 efficiency as photo-Fenton auxiliaries of three different types 16 of soluble bio-based substances (SBO) 17 Juan Gomis^{,a}, Mayara G. Gonçalves^b, Rosa F. Vercher^a, María C. Sabater^c, María A. 18 Castillo^c, Alessandra Bianco Prevot^d, Ana M. Amat^a, Antonio Argues^{a*}. 19 20 21 ^a Grupo de Procesos de Oxidación Avanzada. Universitat Politècnica de Valéncia, Plaza 22 Ferrándiz y Carbonell (s/n), E-03801, Alcoy, Spain, ^b U. Tecnológica Federal do Paraná, Rua Deputado Heitor Alencar Furtado, 4900, 23 24 Curitiba, Brasil. ^c Dpto. Biotecnología. Univ. Politècnica de València. Camino de Vera, s/ n. 46022 -25 Valencia, Spain 26 ^d Università. di Torino, Dipartimento di Chimica Via Giuria 7, Torino, Italy 27 28 29 *Corresponding author 30 E-mail address: aarques@txp.upv.es (A. Arques) Phone number: ++34 966528417 31 32 Fax: ++34 966528438 33 34 **Graphical abstract**



39 Soluble bio-based substances (SBO) have been employed to drive a mild photo-Fenton

40 SBO are non-toxic, show poor degradability and are relatively photostable

41 The SBO composition is related with their performance in photochemical processes

42

43 Keywords: SBO, photostability, photo-Fenton, biocompatibility

44

45 Abstract

46 The aim of this work is to determine the photostability biocompatibility and efficiency 47 of water soluble bio-based substances (SBO) in photo-oxidative processes for 48 wastewater treatment. Three batches of SBO, isolated from different sources, have been investigated. Differences in the functional groups present in these substances can 49 50 explain major trends in their physical/chemical properties. Bioassays have proven those 51 materials to be non-toxic but to show poor biodegradability. Their ability to enhance a 52 photo-Fenton process at milder pH (5.2) has been investigated using a mixture of 53 emerging compounds in wastewaters. All the tested SBO were able to remove all 54 pollutants in less than one hour irradiation, and the best results were obtained with those

substances showing higher hydrophilic/hydrophobic ratio. Moreover, although SBOs
themselves undergo a slight oxidation, no relevant negative effect has been observed for
their use in wastewater treatment.

58

59 **1. Introduction**

60

Generation of increasing amount of wastes from human activities has become a serious environmental concern. Developing technologies and processes able to minimize their production should be a priority; however, this is not always possible and alternative approaches are required to deal with this problem, such valorization through their re-use in other processes. In this context, much effort has been devoted in recent years to obtain valuable products from urban bio-wastes (UBW) (see [1] and references therein cited).

68

69 UBW have become a sustainable source of valuable materials, such as water soluble 70 bio-organic substances (SBO). They have been isolated from organic wastes submitted 71 to anaerobic and/or aerobic treatment, yielding biomasses successively hydrolyzed at alkaline pH, the obtained solutions submitted to ultrafiltration, then the retentate is dried 72 73 and soluble bio-based products obtained as potassium salts [2]. Depending on the origin 74 and treatment of the sourcing UBW, different batches of SBO can be obtained, showing 75 a wide range of chemical composition and properties [3]. In general, SBO are 76 constituted by a mixture of macromolecules which average molecular weight ranges from 67 to 463 kg mol⁻¹. These supramolecular assemblies contain long aliphatic 77 78 chains, aromatic rings and several functional groups as carboxyl, primary and 79 substituted amine and amide, carbonyl, hydroxyl, phenol, ether or esther. Previous papers studied the potential applications of SBO as surfactants [4], in materials
chemistry [5], in soil-washing treatments [6], in agriculture [7] or animal husbandry [8].

SBO chemical composition shows high similarity with that of dissolved organic matter 83 (DOM, e.g. humic or fulvic acids), which is known to generate, upon solar irradiation, 84 highly reactive species able to oxidize pollutants, thus contributing to the abiotic self-85 remediation of natural ecosystems [9-11]. For this reason, SBO might also be employed 86 87 in solar photochemical processes for wastewater treatment. Some works have been 88 published reporting on the use SBO as photosensitizer for the degradation of aromatic 89 sulphonic acids [12], phenols [13] or dyes [14, 15]. In addition, SBO might be used to 90 drive a photo-Fenton process under milder conditions. Photo-Fenton is based on the 91 ability of iron salts to decompose hydrogen peroxide into highly reactive species, and 92 the process is greatly enhanced upon irradiation [16]. One major drawback is the highly 93 acidic pH which is required to avoid the formation of photochemically inactive iron 94 oxides and hydroxides; in this context the ability of SBO to complex metal cations, such 95 as iron, is useful for the development of photo-Fenton at circumneutral media [17].

96

97 However, some important issues still remain unexplored. For instance, SBO could only 98 be employed for wastewater detoxification if they show a good biocompatibility, 99 namely low toxicity and high biodegradability. Furthermore, SBO are mainly formed of 100 organic molecules, and can in turn be attacked by reactive species generated in the 101 photo-oxidative processes, with possible changes in their biological or chemical 102 properties. Any modification in SBO composition might result in different performance 103 of these materials and hence, comparison of different batches of SBO is meaningful. 104 With this background the aim of this paper is to gain further insight into the

105 performance, photo-stability and biocompatibility of three types of SBO obtained from 106 different sources. A battery of bioassays have been employed for toxicity and 107 biodegradability determination; SBO performance before and after irradiation has been 108 tested in the degradation of a mixture of emerging pollutants (EPs), namely 109 acetaminophen, caffeine, amoxicillin, clofibric acid, carbamazepine and acetamiprid. 110 These chemicals have been chosen because they are examples of xenobiotics commonly 111 employed in the literature and they have previously employed as target pollutants in 112 works involving SBO in the photo-Fenton process [17]. 113

114 **2. Experimental**

115

116 2.1. Materials

117

The following types of SBO were employed: a) FORSUD, isolated from the urban waste organic humid fraction (UWOHF) obtained from separate source collection, mixed with the digestate from an anaerobic reactor fed with UWOHF; b) CVT230, obtained from home gardening and park trimming residues (GR) piles aerated for 230 days; c) CVDFT110, isolated from a mixture 35/55/10 (w/w/w) of FORSUD, GR and urban sewage sludge mix, aerated for 110 days.

124

The SBO isolation was performed in a pilot plant at the Studio Chiono & Associati in Rivarolo Canavese, Italy [2]. Briefly, it consisted in an electrically heated and mechanically stirred 500 L reactor, a 102 cm long x 10.1 cm diameter polysulfone ultrafiltration (UF) membrane with 5 kDa molecular weight cut-off, and a forced ventilation drying oven. A stirred stainless steel reactor was loaded with 300 L of aqueous solution of KOH and 75 kg of solid biomass at 65 °C and a pH value of ca. 13. After 4 hours, the reactor turned off automatically and the mixture was settled down overnight. The supernatant was pumped to a centrifuge (9000 rpm) to remove the small insoluble fraction (mainly silicoaluminates) that was still present after the treatment. The recovered liquid phase was flown at 40 L/h through an UF membrane operating with tangential flow at 7 bar inlet and 4.5 bar outlet pressure to yield a retentate containing 5-10 % of dry matter, which was finally dried at 60 °C. Solid SBO products were obtained in 15-30 % w/w yield, relatively to the starting UBW dry matter.

138

139 Acetaminophen, caffeine, amoxicillin, clofibric acid, carbamazepine and acetamiprid, 140 employed as target pollutants, were purchased from Sigma-Aldrich. Hydrogen peroxide 141 (30% v/v) and ferrous sulphate were supplied by Panreac. Water employed in all the 142 experiments was Milli-Q grade.

143

The bioluminescent bacteria *Vibrio fischeri* (strain NRRL-B-11177) were purchased by
Macherey-Nagel GmbH & Co. (Düren, Germany). *P. subcapitata*, and ephippia
(dormant eggs) of crustacean *Daphnia magna* were supplied by ECOtest S.L. (Valencia,
Spain)

148

149 2.2. Reactions

150

Experiments devoted to check the SBO photo-stability were performed in cylindrical Pyrex vessel (55 mm i.d.). A solar simulator (Sun 2000, ABET Technologies) equipped with a 550 W Xenon Short Arc Lamp was used as irradiation source (see [18] for the spectrum, which closely matches the solar one). For each experiment, the reactor was loaded with 250 mL of solution containing the SBO (100 mg/L). The pH was adjusted

to the desired value by adding diluted sulphuric acid. Eventually, the stoichiometric amount of hydrogen peroxide required to oxidize completely each SBO was added, and irradiation was kept until the solution was free of H_2O_2 . The amount of H_2O_2 was obtained from the initial chemical oxygen demand (COD) of the SBO, as COD indicates the amount of O_2 required to oxidize completely the sample.

The same experimental device was employed to check the performance of SBO in a mild photo-Fenton process. In this case, the mixture of all six EPs was employed at an initial concentration of each pollutant of 5 mg/L. All three types of SBO (10 mg/L) were used in parallel experiments. The initial concentration of iron(II) was 5 mg/L (added as sulfate salt) and half the stoichiometric amount of hydrogen peroxide to mineralize the pollutants present in the solution was added. The pH was adjusted to 5.2.

168

169 2.3. Chemical analysis

170

171 The EPs concentration was determined by chromatography (Perkin Elmer model Flexar 172 UPLC FX-10) equipped with a UV-vis detector. A DB-C18 Brownlee Analytical 173 column was used and the eluent was a mixture of acetonitrile (A) and a 0.1% formic 174 acid aqueous solution (B); its composition was changed in a linear gradient, from 3% A 175 to 70% A in 8 min with a flow rate of 0.3 mL/min. Wavelengths employed for detection 176 were: 205 nm (acetaminophen, amoxicillin, caffeine and carbamazepine), 225 nm 177 (clofibric acid) and 245 nm (acetamiprid). Samples were filtered through polypropylene 178 filters (VWR, 0.45 µm) before analysis.

Dissolved organic carbon (DOC) was determined with a Shimadzu model TOC-V CSH apparatus. Chemical oxygen demand (COD) was determined according to the dichromate method [18]: sample digestions were performed at 148°C in a Thermoreaktor TR300 (Merck) and a Spectroquant NOVA 60 (Merck) was used for the photometric determination. The surface tension of samples was determined by a Krüss K-9 tensiometer.

186

187 2.4. Bioassays

188

In order to check the biodegradability of all three substances, biological oxygen demand (BOD₅) assays were carried out according to the standard manometric method (OECD 301 B, CO₂ evolution test), using an OxiTop[©] (WTW) to seal the bottle and determine the pressure inside [18]. The BOD₅ values were determined for SBO concentrations of 100 mg/L and 1000 mg/L.

194

195 Toxicity assays based on V. fischeri bacteria were performed based on ISO 11348-196 3:2007 standardized test; algae growth inhibition assay was performed according to an adaptation of ISO 8692:2004 test, using the chlorophyceae algae Pseudokirchneriella 197 198 subcapitata; bioassays based on the inhibition of the mobility of D. magna were performed according to the standard ISO 6341:199632 procedure. A detailed 199 200 description of the experimental procedure followed in all three cases can be found as 201 Suplementary Data (S1). They were chosen as examples of bioassays commonly employed to detect toxicity in water involving different micro-organisms, namely 202 bacteria, algae and crustaceans, in order to better assess the toxicity of these substances 203 204 in water.

206 **3. Results**

207

- 208 3.1. Chemical and physical characterization of the SBO
- 209

A chemical characterization of FORSUD, CVT230 and CVDFT110 is summarized in 210 211 Table 1. Data show that FORSUD has the highest organic content, as indicated by the 212 percentages of volatile solids and carbon; on the other hand, scarce differences can be 213 found between CVT230 and CVDFT110 regarding this parameter. It is interesting to 214 notice that a significant amount of metals can be found in all three materials; this can be 215 due either to metal already present in the starting materials or added in the isolation 216 process (e.g. potassium). The presence of metals, in particular iron, might be of interest, 217 as the ability of complexed iron to drive a photo-Fenton process at mild conditions has 218 been recently described [17]; on the other hand, heavy metals, although found at very 219 low amounts, might be a concern, and hence, bioassays to determine SBO 220 biocompatibility are required.

221

The gross amount of ions released into the aqueous medium might be important for the performance of a photo-oxidative reaction and can be related to the conductivity of solutions of these substances. Although values are not very different, higher conductivity was measured for FORSUD (29 mS/cm for 1 g/L concentration, 9 mS/cm for 50 mg/L) while lower values were obtained for the other two studied materials (8-12 mS/cm for 1 g/L and 0.3-0.5 mS/cm for 50 mg/L).

228

In addition, the relative amount of some functional groups present in all three SBOs is shown in Table 2. In the case of FORSUD, aliphatic carbon is the predominating

species (43 %), and although this is also true for the other SBOs, the proportion is lower
(31-37%); on the other hand aromatic carbon is more abundant in CVDFT110 and
CVT230. Consequently, the aliphatic/aromatic ratio obtained was clearly higher for
FORSUD than for the other two SBOs. It is interesting to note that, again in this case,
similar composition was obtained for CFDFT110 and CVT230.

236

237 Another useful gross parameter employed in characterization of natural organic matter 238 is the E_2/E_3 ratio [19], where E_2 and E_3 are the absorbances at 250 nm and 365 nm 239 respectively; this ratio has been related with the average molecular weight of the 240 macromolecules, as well as its aromaticity (the higher ratio, the lower aromaticity and 241 molecular weight). The application to the E2/E3 ratio to SBO gave results consistent 242 with theier composition: higher values for E_2/E_3 were indeed obtained for FORSUD 243 (3.83) that, as stated above, shows the minor proportion of aromatic groups; lower 244 values were obtained for the other two SBO (ca. 2.3)

245

246 The structure of the SBO can be also related their surfactant activity. In order to gain 247 further insight into this behavior, surface tension of water was measured in the presence 248 of four different concentrations of these materials. Results given in Figure 1 shows that 249 higher decreases in surface tension were observed for FORSUD, as 1 g/L of this 250 material was able to decrease the surface tension to 46 mN/m (the value for distilled 251 water is ca. 73 mN/m); although the other SBO also behave as surfactants, significantly 252 lower decreases in surface tension were measured. This different behavior can be 253 explained by considering the hydrophilic/hydrophobic ratio of these materials; this 254 parameter was calculated as a ratio between the carbon atoms present at hydrophobic 255 functional groups (e.g. aliphatic or aromatic) and those present in hydrophilic moieties,

such as carboxylic, hydroxylic, phenolic or amines. By comparing data of Table 2 and
Figure 1, it can be observed that the percentage of aliphatic carbon follows the same
trend as the decrease in the surface tension: FORSUD >> CVT230 > CVDFT110.

259

260 3.2. Biocompatibility of the SBO

261

262 BOD₅ was used to check the biodegradability of these materials in two different 263 concentrations, 100 mg/L and 1000 mg/L. Table 3 shows that low BOD₅ values were 264 obtained in all cases. As this parameter is highly influenced by the amount of organics 265 present in the solution, the BOD/COD ratio has been employed as a better indicator for 266 biodegradability [20]. This ratio was systematically below 0.1 (Table 3), what can be 267 associated with a poor biodegradability of SBOs. This could be attributed to the fact that 268 these materials have been submitted, for a long time, to the action of microorganisms 269 and hence, only the fraction refractory to biological degradation remains. Other organic 270 materials which are the result of the action of microbiological activity for long periods 271 of time also show poor biodegradability; examples of such substances are humic and 272 fulvic substances [21, 22] or landfill leachates [23].

273

However, low biodegradability should not be a major concern if it is not associated to
toxicity, as concentrations employed in wastewater treatment are systematically well
below 100 mg/L [15, 17, 24]. Results shown in Figure 2 indicate that the toxicity of
these materials is rather low in the range of concentrations tested (0-50 mg/L), as
toxicity values were lower than 30% in all the bioassays employed (based on *V. fischeri*, *P. subcapitata* and *D. magna*), and even they were below 5% in most of the assays. This

implies that SBO can be classified as non-toxic materials showing very poorbiodegradability.

282

283 3.3. SBO as auxiliaries in photochemical methods for wastewater treatment

284

285 The role of SBO in photochemical methods for wastewater treatment has been investigated in a series of previous papers, using CVT230. Briefly, results show that the 286 287 use of this material as photocatalyst under UV-visible light irradiation is not attractive, 288 as the screen effect of the strongly coloured SBO solutions compensate their capability 289 to generate reactive species [15, 17]. On the contrary, the use of SBO as iron 290 complexing agent to implement photo-Fenton process at mild conditions seems more 291 promising, as the reaction could be implemented at pH = 5 without a significant loss of 292 efficiency [24].

293

294 However, only results using CVT230 are available and the different composition of the 295 other SBO might influence their performances. Hence, photodegradation of a mixture of 296 six emerging pollutants (acetaminophen, caffeine, acetemiprid, clofibric acid, 297 carbamazepine and amoxicillin) was studied with all three substances, FORSUD, 298 CVT230 and CVDFT110. As happened with CVT230 [17], none of the SBO was able 299 to enhance the photolysis of the studied pollutants under visible light irradiation (data 300 not shown). On the other hand, Figure 3 shows the all three material were able to drive a 301 photo-Fenton process at mild conditions (pH = 5.2), as most pollutants were removed in 302 less than 30 min of irradiation and only acetamiprid required longer treatment times to 303 be completely eliminated (30-60 min). However some slight differences could be found 304 in the efficiency of all three materials, which followed the trend CVDFT110 \geq CVT230

> FORSUD. This could be attributed to differences in the functional groups present in
all three materials: CVDFT110 and CVT230 show relatively similar composition, with
lower amounts of aliphatic carbon, which is expected to be photochemically non active
and unable to complex iron; these functional groups are present in more extent in
FORSUD, showing the worst performance.

310

The toxicity of the sample was followed along the process according to the inhibition of 311 312 the luminescence of V. fischeri. It can be observed in Figure 4, that although a very 313 slight increase in the toxicity was observed, most probably due to the release of toxic 314 by-products at the early stages of the photochemical process, as reported before 315 elsewhere [26']; however, these values were very low (systematically below 15%) and then decreased to reach final values below 5%. As the concentration of pharmaceuticals 316 317 in the environment is significantly lower than that employed in this work, and that SBO 318 amount should not be above 20 mg/L, this might indicate that the risk associated with 319 the release of toxic chemical during the process is acceptable, although further research 320 on this issue is still necessary.

321

323

It is interesting to study the photostability of the SBOs as in the wastewater treatment processes they are also submitted to chemical photo-oxidation; furthermore, this information might be valuable for studies on the behavior of anthropogenic organic matter in the environment. For this purpose, they were irradiated (100 mg/L) for 24 hours with a solar simulator; DOC, COD and BOD₅ were determined at selected sampling times. Figure 4 shows that there was a slow decrease in DOC and COD values

^{322 3.4} Photostability of SBOs

330 for CVT230, (ca. 25 % after 1 day of irradiation) indicating that some oxidation of SBO 331 occurred; this can be more easily appreciated by the chemical oxidation state (COS), 332 which is a parameter which relates the COD at the sampling with the initial DOC and a 333 COS value of +4 is assigned for CO₂ (the most oxidized species of carbon) and -4 for 334 CH₄ (the most reduced) [25]. During the process, the COS was increased from -0.9 to +0.2, and it should be associated with an oxidation of the organics present in the 335 sample; a similar behavior was observed for CVDFT110 and FORSUD (data not 336 337 shown). On the other hand, average oxidation state, which only refers to the organics 338 remaining in solution (CO₂ is excluded from calculations by employing the DOC at the 339 sampling time) did not suffer significant changes oscillating between -1 and -0.6 340 throughout the process; hence, no dramatic changes in the composition of the remaining 341 organics might be expectable. Regarding to biocompatibility, the BOD₅/COD ratio was 342 calculated after 24 hours of irradiation and the obtained values were 0.24 for 343 CVDFT110, 0.08 for CVT230 and 0.14 for FORSUD; this ratio was in all cases slightly 344 above that calculated for the same SBO before irradiation; finally, toxicity values did 345 not vary significantly throughout the process (data not shown). A possible explanation 346 to the above described trends is that the SBO molecules are partly oxidized into smaller 347 and more hydrophilic ones which can be more easily metabolized by micro-organisms. 348 A decrease of the molecular weight of macromolecules under irradiation is a well-349 known behaviour for macromolecules, for instance it has been observed for humic and 350 fulvic substances when they were irradiated [26].

351

The same procedure was followed in the presence of hydrogen peroxide, as it has been described that the amounts of iron present in the SBO composition can drive a photo-Fenton like process. For this purpose, the stoichiometric amount of peroxide required to 355 oxidize completely each SBO was added and it was irradiated until the solution was free 356 of H₂O₂. Table 4 shows variations in significant parameters. In all cases, similar trends 357 were obtained, only diverging in the numerical values, which should not be 358 overemphasized: a) there was a moderate mineralization of the samples, in the range 20-359 30%, b) COD also decreased with a concomitant increase in COS, what indicates that 360 stronger oxidation than for photolysis was obtained, c) AOS also increased, although variation was lower than for COS what might indicate that severe modification of 361 362 organics remaining in solution should be ruled out, d) there was some increase in 363 biodegradability as shown by higher BOD₅ values, and more clearly by the BOD/COD 364 ratio, although these changes were not very acute, e) surface tension increased to reach 365 values close to that of distilled water, f) the E_2/E_3 ratio significantly increased. All these 366 data are again in agreement with a cleavage of the starting SBO to generate smaller, 367 more oxidized and more hydrophilic substancess. In fact, similar trends have been 368 observed for landfill leachates [27].

369

It was also interesting to determine the effect of mild photo-Fenton on SBOs oxidation For this purpose the three different SBOs (100 mg/L) were submitted to a photo-Fenton at pH = 5.2 with the stoichiometric amount of hydrogen peroxide and 5 mg/L of iron(II). Data obtained after 4 hours of irradiation showed very similar trends to that irradiated for one day, with DOC and COD decreases systematically above 50% (in some cases up to 75%) and the E_2/E_3 was between 4 and 5 for CVT230 and CVDFT110 while it reached 15 in the case of FORSUD.

377

Finally, the effect of oxidation on SBO performance was also analyzed. For thispurpose, samples containing 100 mg/L of each SBO were irradiated in the presence of

380 the stoichiometric amount of hydrogen peroxide; once they were free of hydrogen 381 peroxide, 25 mL of this solution was diluted to 250 mL (to reach an initial SBO 382 concentration of 10 mg/L); the emerging pollutants (acetaminophen, caffeine, 383 acetamiprid, clofibric acid, carbamazepine and amoxicillin) were added (5 mg/L of 384 each), as well as hydrogen peroxide (half the stoichiometric amount required to oxidize 385 the pollutants) and iron (5 mg/L); finally the pH adjusted to 5.2 and the mixture was 386 irradiated. A plot of the relative concentration of each pollutant vs. time can be found in 387 Figure 5. When comparing with Figure 2, it can be observed that all three photo-388 oxidized materials were still able to drive a mild photo-Fenton. Nonetheless differences 389 among them can be appreciated: while photo-treated FORSUD showed better 390 performance than the original material, the reverse was true for the others. This might 391 be due to differences in the functional groups present in those substances: oxidation of 392 FORSUD, where aliphatic carbon mas predomination, might result in an increase in the 393 number of hydrophylic groups able to complex iron; on the other hand, in CVDFT110 394 and CVT230, oxidation might result in a stronger degradation of the molecules and 395 some loss of photochemical activity.

396

397 4. Conclusions

The obtained results underline the importance of SBO characterization and of establishing structure-properties relationship in order to optimize the choice of the product to be used depending on the specific application. With particular attention to mild photo-Fenton processes, this work shows the importance of carefully considering the effect of irradiation not only on the target pollutant to be degraded but also on the SBO used to promote the pollutant degradation, in order to choose the most efficient for longer times.

- 406 Despite their low biodegradability, SBO do not show relevant toxicity, thus encouraging407 further investigation on their use in mild photo-Fenton processes.
- 408

409 Acknowledgements

410 The authors want to thank the financial support of the European Union (PIRSES-GA-

411 2010-269128, EnvironBOS) and Spanish Ministerio de Educación y Ciencia

- 412 (CTQ2012-38754-C03-02). Juan Gomis would like to thank UPV for his FPI grant
- 413 (2010-07).
- 414

415 **References**

- 416 [1] A. Demirbas, Energy Convers. Manage. 50 (2009) 2782-2801.
- 417 [2] E. Montoneri, D. Mainero, V. Boffa, D.G. Perrone, C. Montoneri, C. Int. J. Global
- 418 Environ. Issues 11 (2011) 170-196.
- 419 [3] E. Montoneri, V. Boffa, P. Savarino, D.G. Perrone, M. Ghezzo, C. Montoneri, M.
- 420 Mendichi, Waste Manage. 31 (2011) 10-17.
- 421 [4] P. Quagliotto, E. Montoneri, F. Tambone, F. Adani, R. Goberto, G. Viscardi,
- 422 Environ. Sci. Technol. 40 (2006) 1686-1692.
- 423 [5] V. Boffa, D.G. Perrone, E. Montoneri, G. Magnacca, L. Bertinetti, L. Garlasco, R.
- 424 Mendichi, ChemSusChem 3 (2010) 445-452.
- 425 [6] E. Montoneri, L. Tomasso, N. Colajanni, I. Zelano, F. Alberi, G. Cossa, R. Barberis,
- 426 Int. J. Environ. Sci. Technol. 11 (2014) 251-262.
- 427 [7] O. Sortino, M. Dipasquale, E. Montoneri, L. Tomasso, P. Avetta, A. Bianco Prevot,
- 428 Agron. Sustain. Dev. 33 (2013) 433-441.

- 429 [8] C. Montoneri, E. Montoneri, L. Tomasso, A. Piva, J. Agriculture Sci. 13 (2013) 31430 44.
- 431 [9] C. Tixier, H.P. Singer, S. Oellers, S.R. Müller, Environ. Sci. Technol. 37 (2003)
 432 1061-1068.
- 433 [10] J. Peuravuori, K. Pihlaja, Anal. Bioanal. Chem. 394 (2009) 1621-1636.
- 434 [11] J.J. Guerard, Y.P. Chin, H. Mash, C.M. Hadad, Environ. Sci. Technol. 43 (2009)
 435 8587-8592.
- 436 [12] P. Avetta, A. Bianco Prevot, D. Fabbri, E. Montoneri, L. Tomasso, Chem. Eng. J.
 437 197 (2012) 193-198.
- 438 [13] A. Bianco Prevot, P. Avetta, D. Fabbri, E. Laurenti, T. Marchis, D.G. Perrone,
- 439 Montoneri, V. Boffa, ChemSusChem, 4 (2011) 85-90.
- 440 [14] A. Bianco Prevot, D. Fabbri, E. Pramauro, C. Baiocchi, C. Medana, E. Montoneri,
- 441 V. Boffa, J. Photochem. Photobiol. A: Chem. 209 (2010) 224-231.
- 442 [15] J. Gomis, R.F. Vercher, A.M. Amat, D.O. Mártire, M.C. González, A. Bianco
- 443 Prevot, E. Montoneri, A. Arques, L. Carlos, Catal. Today 209 (2013) 176-180.
- 444 [16] J.J. Pignatello, E. Oliveros, A. MacKay, Crit. Rev. Environ. Sci. Technol. 36445 (2006) 1-84.
- 446 [17] J. Gomis, A. Bianco Prevot, E. Montoneri, M.C. González, A.M. Amat, D.O.
- 447 Mártire, A. Arques, L. Carlos, Chem. Eng. J. 235 (2014) 236-243.
- 448 [18] L. Carlos, D.O. Martire, M.C. Gonzalez, J. Gomis, A. Bernabeu, A.M. Amat, A.
- 449 Water Res. 46 (2012) 4732-4740.
- 450 [18] APHA, AWWA, WPCF, Standard Methods for the examination of water and
- 451 wastewater. 20th edition. American Public Health Association, Washington DC, 1998.
- 452 [19] J. Peuravuori, K. Pihlaja, Anal. Chim. Acta 337 (1997) 133-149.
- 453 [20] W. Guhl, J. Steber, Chemosphere 63 (2006) 9-16.

- 454 [21] R. G. Quails, B. L. Raines, Soil Sci. Soc. Am. J. 56 (1992) 578-586.
- 455 [22] Xu, X.Y., Zeng, G.M., Peng, Y.R., Zeng, Z., Chem. Eng. J. 200-202 (2012) 25-31.
- 456 [23] J. Lopes de Morais, P. Peralta Zamora, J. Hazard. Mater. B 123 (2005) 181–186.
- 457 [24] J. Gomis, L. Carlos, A. Bianco Prevot, A.C.S.C. Teixeira, M. Mora, A.M. Amat, R.
- 458 Vicente, A. Arques, Catal Today, 2014, in press.
- 459 [25] A.M. Amat, A. Arques, F. Galindo, M.A. Miranda, L. Santos-Juanes, R.F. Vercher,
- 460 R. Vicente, Appl. Catal. B: Environ. 73 (2007) 51-57.
- 461 [26'] A. Bernabeu, S. Palacios, R.Vicente, R. Vercher, S.Malato, A. Arques, A.M.
- 462 Amat, Chem Eng. J. 198-199 (2012) 65-72.
- 463 [26] L. Lou, H. Xie, Chemosphere 65 (2006) 2333-2342
- 464 [27] R.C. Martins, D.V. Lopes, M.J. Quina, R.M. Quinta-Ferreira, Chem. Eng. J. 192
- **465** (2012) 219-225.

466	
467	

468 Table 1. Elemental composition of the three batches of SBOs employed in this work.

	FORSUD	CVDF	CVT230
Volatile solids (%, w/w)	84.6	72.7	72.1
Carbon (%, w/w)	45.1	35.5	38.2
Nitrogen (%, w/w)	7.8	4.3	4.0
Si (%, w/w)	0.36	0.92	2.55
Fe (%, w/w)	0.16	0.53	0.77
Al (%, w/w)	0.78	0.44	0.49
Mg (%, w/w)	0.18	0.49	1.13
Ca (%, w/w)	1.32	2.59	6.07
K (%, w/w)	9.2	5.4	3.6
Na (%, w/w)	0.39	0.15	0.16
Cu (mg/kg)	100	216	202
Ni (mg/kg)	27	71	92
Zn (mg/kg)	185	353	256
Cr (mg/kg)	11	30	19
Pb (mg/kg)	44	75	85
<mark>Hg (mg/kg)</mark>	0.2	0.4	0.2

	FORSUD	CVDFT110	CVT230
Aliphatic carbon (%)	43	31	37
Amine (%)	10	8	7
Methoxy (%)	4	-	-
Alkoxy (%)	10	20	14
Anomeric carbon (%)	3	7	4
Aromatic (%)	10	16	13
Phenolic carbon (%)	2	6	5
Phenoxy (%)	1	2	2
Carboxylic (%)	7	9	12
Amide (%)	9	1	1
Carbonilic (%)	1	-	5
Lipophilic/hydrophilic ratio	9.3	5.3	3.6
Aliphatic/aromatic ratio	3.3	1.3	1.8
E_2/E_3 ratio	3.83	2.31	2.38

471 Table 2. Main functional groups present in the three batches of SBOs that are studies

474	Table 3:	BOD ₅	values	and	BOD/COD	ratio	obtained	for	all	three	different	types	of
	14010 01	DOD	(araco	un a	D0D/00D	14410	ootamea	101	wi i		GILLOLOLIC	c) p c b	U 1

	100	mg/l	1 g/l			
	BOD (mg/l)	BOD/COD	BOD (mg/l)	BOD/COD		
CVDFT 110	6	0.06	20	0.02		
CVT 230	4	0.04	30	0.03		
FORSUD	10	0.09	80	0.08		

475 SBOs at two different concentrations (COD values are shown in Table 4)

- 478 Table 4: Irradiation of all three SBOs in the presence of hydrogen peroxide: main
- 479 parameters before and after irradiation. The initial SBO concentration was 100 mg/L, the

480 concentration of H₂O₂ was 215 mg/L (the stoichiometric amount required to oxidize

481 completely the SBO) and irradiation time was 6 hours.

482

	CVDFT 110		CVT 230)	FORSUD		
	Before	Irradiated	Before	Irradiated	Before	Irradiated	
DOC (mg/l)	30.2	23.9	28.2	23.5	35.0	25.0	
COD (mg/l)	101	76	95	63	111	67	
BOD ₅	6	15	4	9	10	15	
AOS	-1,01	-0.76	-1.04	0.02	-0.77	0.08	
COS	-1,01	0.24	-1.04	0.65	-0.77	1.12	
BOD/COD	0.06	0.20	0.04	0.14	0.09	0.22	
Surface tension	73	73	69	73	60	66	
E_2/E_3 ratio	2.31	3.55	2.38	3.89	3.83	11.1	



487 Figure 1: Surface tension of solutions containing different concentrations of three
488 batches of SBOs: CVDFT 110 (black bars) CVT 230 (grey bars), FORSUD (white
489 bars).



494 Figure 2: Toxicity of three different batches of SBOs (CVDFT 110 (♦) CVT 230 (■),
495 FORSUD (▲)) measured according to bioassays based on *V. fischeri* (above), *P.*496 *subcapitata* (middle) and *D. magna* (below).







Figure 3. Photo-Fenton process at pH = 5.2 in the presence of three types of SBOs: 505 CVT 230 (above), CVDFT 110 (middle) and FORSUD (below). Plot of the relative 506 concentration of pollutant vs time: amoxycillin (□), acetaminophen (*), acetemiprid (\blacklozenge), caffeine (\times), clofibric acid (\blacklozenge) and carbamazepine (\blacktriangle) 507





509 Figure 4: Inhibition of the luminescence of V. fischeri bacteria obtained during a photo-

510 Fenton process at pH = 5.2 of the mixture of six pollutants in the presence of three types

511 of SBOs: CVT 230 (♦), CVDFT 110 (■) and FORSUD (▲).





515 Figure 5: Changes in COD (\blacksquare), DOC (\blacklozenge), AOS (\blacktriangle) and COS (\times) vs. irradiation time

- 516 for CVT 230.





522 Figure 6. Photo-Fenton process at neutral pH in the presence of three types of irradiated 523 SBOs: CVT 230 (above), CVDFT 110 (middle) and FORSUD (below). Plot of the relative concentration of pollutant vs time: amoxycillin (\Box), acetaminophen (*), 524 acetemiprid (\blacklozenge), caffeine (\times), clofibric acid (\bigcirc) and carbamazepine (\blacktriangle) 525