

AperTO - Archivio Istituzionale Open Access dell'Università di Torino

Anthracene and phenanthrene tropospheric oxidation promoted by the nitrate radical in the gas-phase. Theoretical modelistic study

This is the author's manuscript

Original Citation:

Availability:

This version is available <http://hdl.handle.net/2318/1652434> since 2017-11-21T12:18:20Z

Published version:

DOI:10.1016/j.atmosenv.2017.08.011

Terms of use:

Open Access

Anyone can freely access the full text of works made available as "Open Access". Works made available under a Creative Commons license can be used according to the terms and conditions of said license. Use of all other works requires consent of the right holder (author or publisher) if not exempted from copyright protection by the applicable law.

(Article begins on next page)



UNIVERSITÀ DEGLI STUDI DI TORINO

This is an author version of the contribution published on:

Questa è la versione dell'autore dell'opera:

[Atmospheric Environment 167, 181, 2017, 10.1016/j.atmosenv.2017.08.011]

The definitive version is available at:

La versione definitiva è disponibile alla URL:

[<http://www.sciencedirect.com/science/article/pii/S1352231017305162>]

1 Anthracene and phenanthrene tropospheric oxidation promoted by 2 the nitrate radical in the gas-phase. Theoretical modelistic study. 3

4 Andrea Maranzana, Giovanni Ghigo, and Glauco Tonachini*

5 *Dipartimento di Chimica, Università di Torino, Corso Massimo D'Azeglio 48, I-10125 Torino, Italy*

6

7 **Abstract.** Polycyclic aromatic hydrocarbons and their oxidized derivatives are ubiquitous
8 environmental pollutants that are toxic to different degrees. The NO₃ radical is known to be an
9 important actor in fostering nighttime atmospheric chemistry; hence it could elicit some nocturnal
10 PAH-loss processes. In this study the gas phase pathways open to the initial anthracene and
11 phenanthrene NO₃ π-radical adducts in the presence of O₂, NO and NO₂ are examined by Density
12 Functional Theory to ascertain the mechanistic features of their NO₃-initiated oxidative degradation.
13 Unimolecular steps involving the initial adducts (ring closures or fragmentation) present rather high
14 free energy barriers and seem unlikely. Regarding bimolecular reactions, any radical present in the
15 tropospheric environment can give an intrinsically fast radical coupling with the initial adducts, thus
16 producing bifunctional closed shell species. The addition of ground state dioxygen is a relatively fast
17 step that is reversible in some cases, less so in others. It entails subsequent reaction steps, which
18 involve the peroxy radical intermediate and addition and loss of small species as NO, NO₃, NO₂ and
19 O₂. These steps can either trigger a β-fragmentation with formation of a closed shell dialdehyde or
20 formation of a closed shell nitroxy ketone, bifunctional species that appear to be the most likely
21 products when typical tropospheric concentrations are taken into account.

22

23

24 **Keywords:** 1) anthracene; 2) phenanthrene; 3) nitrate radical; 4) NO₃ 5) tropospheric oxidation;
25 6) gas-phase 7) mechanism; 8) DFT

26 * corresponding author. Address: Dipartimento di Chimica, Università di Torino, Corso Massimo D'Azeglio 48,
27 I-10125 Torino, Italy.

28 **E-mail addresses:** Andrea Maranzana: andrea.maranzana@unito.it - Giovanni Ghigo: giovanni.ghigo@unito.it
29 Glauco Tonachini: glauco.tonachini@unito.it / phone: ++39-011-6707648 / fax: ++39-011-2367648

30 **Proposed running title:** PAH oxidation promoted by NO₃

31

32

1. Introduction.

Polycyclic aromatic hydrocarbons (PAHs) and their oxidized derivatives are ubiquitous environmental pollutants present in the troposphere, which are toxic to different extents (Finlayson-Pitts and Pitts, 2000; Keyte et al., 2013). They are emitted during incomplete combustion of fossil fuels (Lim et al., 1999) and biomass (Jenkins et al., 1996; Mandalakis et al., 2005; Samburova et al., 2016). PAHs can undergo oxidation and functionalization in general, both during combustion and their subsequent transport in the troposphere. (Calvert et al. 2002; Kameda 2011) In particular, anthracene and phenanthrene belong to the list of the 16 priority PAHs indicated by the US Environmental Protection Agency. This list has played an important role in monitoring programs (Lammel, 2015), though reservations about its future use has been recently put forth (Andersson and Achten, 2015).

Homogeneous (gas phase) and heterogeneous (particle phase) experimental and theoretical studies have been published on the reactions of PAHs with ozone (Finlayson-Pitts, 1997; Perraudin et al., 2007, 2006; Zhang et al., 2013) or with the hydroxyl radical (Biermann et al., 1985; Brubaker and Hites, 1998; Dang et al., 2015a, 2015b, 2014; Giordana et al., 2008; Helmig and Harger, 1994; Sasaki et al., 1997; Zhao et al., 2016). Some publications have also dealt with the action of the nitrate radical (Atkinson et al., 1990; Dang et al., 2015c; Qu et al., 2008; Sasaki et al., 1997; Zhang P., 2014; Chapleski et al., 2016). The reaction with the NO_3 radical, known to play in general a significant role in nighttime tropospheric chemistry (Stone et al., 2014), may have some importance in contributing to nocturnal PAH loss pathways. Current knowledge of the NO_3 -initiated gas phase atmospheric oxidation of different PAHs, based on experimental results, is mostly limited to cases in which a high concentration of NO_2 is also present. For example, in their studies on naphthalene and fluoranthene, Sasaki et al. found nitro-derivatives as important products (Arey et al., 1989; Sasaki et al., 1997). In their reacting system they had not only NO_3 , but also substantial concentrations of NO_2 , both coming from N_2O_5 (NO_2 was also added with concentration in the order of 10^{13} molecules cm^{-3}). Whereas a variety of experimental studies dealing with naphthalene and its derivatives + NO_3 reacting system is available (see Keyte et al. 2013, pp 9349-9353), much less has been done for anthracene and phenanthrene, considering both the gas phase and the particle phase (Arey et al., 1989; Jariyasopit et al., 2014b; Kwok et al., 1994).

Some of the oxidized PAH derivatives are of concern for human health, (Asare et al., 2009, 2008; Durant et al., 1996; Landvik et al., 2007; Talaska et al., 1996) and in addition a large share of oxidation products are not identified. Therefore it is worthwhile to further investigate the mechanistic details of these processes. Theoretical investigations as the present one can be complementary to experimental studies, and the comparison be informative. Within it, we explore the gas phase

reaction mechanisms of NO_3 -triggered oxidative degradation of anthracene and phenanthrene. Therefore, we will consider how ground electronic state O_2 , NO , and NO_2 will react with PAH-derived intermediate radicals. We model these reactions in a semi-quantitative way to ascertain their mechanistic features (reaction pathways and related free energy barriers) and also with the purpose of suggesting, on that basis, likely closed shell products. Accordingly, we deliberately do not extend the investigation to all possible reaction pathways when differing for regio- and diastereochemistry, as is done with the goal of obtaining *quantitative* predictions.

Under typical atmospheric conditions, anthracene and phenanthrene are partitioned between gas and particle phases (see Figures 10.2 and 10.3 in Finlayson-Pitts and Pitts, 2000) for which oxidative degradation might present mechanistic differences. In the present study we begin by focusing only on the gas phase reaction pathways, for which the syn additions are explored in a more complete and detailed way (syn additions are expected to be more pertinent to the adsorbed system, in which one face of the PAH system is engaged with the underlying surface).

To our knowledge experimental results are scarce for phenanthrene (Kwok et al., 1994) and anthracene (Arey et al., 1989; Jariyasopit et al., 2014b; Ringuet et al. 2012), with more data available for naphthalene. Some publications deal with our title PAHs, but are either carried out in the presence of high NO_2 concentration, or deal with heterogeneous chemistry of these and other PAHs. (Mak et al 2007; Zhang et al. 2011; Liu et al. 2012; Cochran et al. 2016) Some theoretical papers published in the last few years also bear similarities to the present study. In particular, a study on hydroxycyclohexadienyl peroxy radical (from benzene) plus NO , (Yan et al., 2013) those on benzene (Qu et al., 2006a) or naphthalene (Qu et al., 2006b) plus NO_3 , one on 9-chloroanthracene plus NO_3 (Dang et al., 2015c), and one on the role of water in the formation of nitro-PAHs (Zhang Q. et al., 2014).

2. Theoretical Method.

All stationary points on the energy hypersurface, *i.e.* minima and first order saddle points, corresponding to transition structures (TS), were determined by gradient procedures (Pople et al., 1992; Schlegel, 1982a, 1982b, 1981; Schlegel et al., 1984) within the Density Functional Theory (DFT), (Parr and Yang, 1989) and making use of the M06-2X (Zhao and Truhlar, 2008, 2007) functional. Reactants, transition structures, intermediates, and products have been optimized with the 6-31G(d,p) basis set. Single-point energies of various species were calculated with Dunning's correlation-consistent polarized valence triple- ζ basis set cc-pVTZ (Kendall et al., 1992). On the basis of previous studies, (Trogolo et al., 2014) this computational approach is expected to perform acceptably. The nature of the critical points (and the thermochemistry) was assessed by vibrational

analysis: the 6-31G(d,p) thermochemical corrections gave estimates of the zero point vibrational energy (ZPE), by which the cc-pVTZ relative energies were corrected [here $\Delta E_{\text{ZPE}} = \Delta E_{\text{cc-pVTZ}} + \Delta \text{ZPE}_{6-31\text{G(d,p)}}$], and similarly enthalpy (ΔH) and Gibbs free energy (ΔG) differences were obtained. The thermochemistry was assessed in all cases at $T = 298.15$ K and ΔG values at this temperature are reported in [Scheme 1](#) (in kcal mol⁻¹). Geometry optimization and thermochemistry calculations were carried out by using the GAUSSIAN 09 system of programs (Frisch et al., 2009).

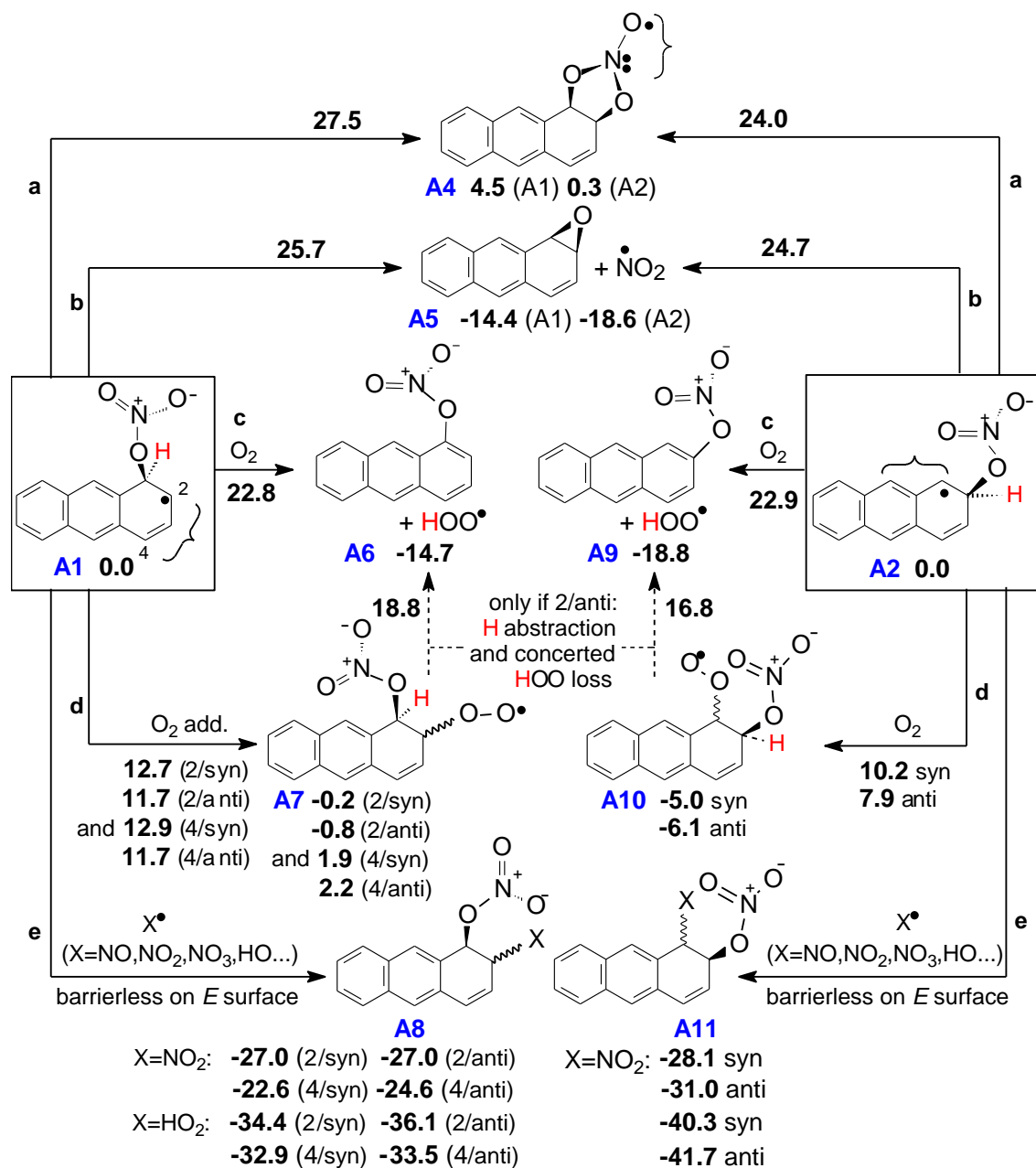
For singlet diradicaloid structures (such as those corresponding to homolytic dissociations), the “automatic” restricted closed shell singlet solution yields zero spin densities. Therefore, the wavefunction stability was checked and obtained by relaxing the wavefunction in the orbital rotations space. The desired spin densities, as expected for diradicaloid structures, were obtained in correspondence of the attained (‘singlet’ unrestricted) wavefunction stability. Because the resultant spin-mixing (the unrestricted wavefunction is not an eigenfunction of \hat{S}^2) gives a better description of the electron distribution but alters the energy, the energy values were refined by Yamaguchi’s formula to get rid of the contribution of the closest spin contaminant of the same parity (Baker et al., 1993; Goldstein et al., 1996; Wittbrodt and Schlegel, 1996; Yamaguchi et al., 1988; Yamanaka et al., 1994).

3. Results and Discussion.

3.1 Anthracene. NO₃ adds to anthracene, without any barrier on the potential energy surface, to form three anthracenyl nitrate radicals. These are the boxed structures **A1** (addition to position 1) and **A2** (position 2) in [Scheme 1](#), and **A3** (position 9) in [Scheme 3](#). Adduct **A1**, at -12.7 kcal mol⁻¹ below the two reactants, is more stable than **A2** by 4.2 kcal mol⁻¹; adduct **A3** is the stablest, being located at -20.2 kcal mol⁻¹, i.e. 7.5 below **A1**. The energetics (ΔG at 298.15 K) for each reaction pathway is reported relative to **A1** in [Scheme 1](#), left, relative to **A2** in [Scheme 1](#), right, and relative to **A3** in [Scheme 3](#). The intermediates **A4** and **A5**, shared by both pathways, are associated with two free energy values, with the relevant zero is indicated in parentheses. [Scheme 2](#) displays further steps departing from the peroxy structures **A7** and **A10** in [Scheme 1](#).

First we will describe the different reaction steps possibly undergone by the initial adducts **A1**, **A2**, and **A3**. Then, we will assess their relative importance. We have taken into account possible unimolecular reaction steps departing from the initial adducts, as well as dioxygen intervention (O₂ is the reactive species present with the highest density, close to 5×10^{18} molecules cm⁻³). Also taken into account are NO and NO₃ reactions with relevant intermediate radicals ([Scheme 2](#)). Section A in the [Supplementary Material](#) displays a larger number of less likely pathways. NO_x species (collective for NO + NO₂) are present in the troposphere at concentrations ranging from 10⁸–10⁹ (unpolluted

- 1 situation) to 10^{12} (polluted) molecules cm^{-3} . The role of NO_x species can be particularly interesting
- 2 in the formation of the concerning PAH nitro derivatives, which have drawn



3

4

5 **Scheme 1.** Anthracene. Initial steps departing from the NO_3 adduct in position 1, **A1** (left, **a**), and in

6 position 2, **A2** (right, **b**). They also involve positions 2 and 1, respectively. Gibbs free energy

7 differences at $T = 298.15$ K, with respect to **A1** and with respect to **A2**; **A2** is $4.2 \text{ kcal mol}^{-1}$ above **A1**.

8 In **A1**, **A2**, and **A4** the brace indicates delocalization.

9 attention in the years (Fiedler and Mücke, 1991; Finlayson-Pitts and Pitts, 2000; Ghigo et al., 2006;

10 Gibson, 1982; Jariyasopit et al., 2014a, 2014b; Keyte et al., 2013; Pitts et al., 1978; Zimmermann et

11 al., 2013). It needs to be taken into account that NO_2 concentration can be high in some laboratory

12 experiments (Arey et al., 1989).

1 **3.1.1 Nitroxy in position 1: initial steps.** Of the initial steps possibly involving **A1** ([Scheme 1](#), left),
 2 two are unimolecular with formation of one extra ring (**A1-A4** and **A1-A5**). Other two (**A1-A6** and
 3 **A1-A7**) imply either H abstraction by O₂, or its addition to the π -system. Still another step similarly
 4 sees NO₂ addition; also other free radicals would similarly add, as NO, HO, or NO₃ itself. In more
 5 detail, (a) **A1** could undergo a 5-membered ring formation (c-CCONO) to its radical isomer **A4** with a
 6 barrier $\Delta G^\ddagger = 27.5$ kcal mol⁻¹. This cyclic intermediate is less stable than **A1** by 4.5 kcal mol⁻¹. (b) As
 7 an alternative, **A1** could lose NO₂, and concertedly close an epoxide ring to give **A5**, with $\Delta G^\ddagger = 25.7$
 8 kcal mol⁻¹. This step is exoergic by $\Delta G = -14.4$ kcal mol⁻¹. Regarding the bimolecular reactions, (c) O₂
 9 could abstract the hydrogen geminal to the nitroxy group in **A1** with consequent formation of 1-
 10 nitroxy-anthracene **A6** (an aryl nitrate) and the hydroperoxyl radical with $\Delta G^\ddagger = 22.8$ kcal mol⁻¹ and a
 11 G gain of 14.7 kcal mol⁻¹ for the step. (d) in **A1**, O₂ could also add to the π -system, namely to the
 12 allylic positions 2 or 4. The related spin densities ($\Delta\rho = \alpha$ electron density - β electron density) are
 13 0.55 on C2, and 0.64 on C4. This addition could proceed through four attacks differing in regio- and
 14 diastereochemistry with respect to the nitroxy group (2 or 4, and syn or anti with respect to the
 15 approximate molecular plane). Only one attack will be fully examined as far as the subsequent steps
 16 are considered: it is the addition to position 2/syn, which presents a barrier $\Delta G^\ddagger = 12.7$ kcal mol⁻¹ to
 17 give the peroxy radical **A7** 2/syn at -0.2 kcal mol⁻¹. By considering that the peroxy radical
 18 intermediates are located at approximately the same energy as the initial NO₃ adduct **A1**, the
 19 dioxygen addition step (if considered alone) can be regarded as quite reversible.
 20 (e) Also a variety of radicals X can similarly add with the same regio- and diastereochemistry, via a
 21 very easy step as regards energy. For simple radical couplings, fairly high rate constants of the order
 22 of 10⁻¹¹ – 10⁻¹² molec⁻¹ cm³ s⁻¹ have been assessed, depending on the reaction type.¹ For instance
 23 NO₂ adds in position 2 without any E barrier, and a nitroxy nitro closed shell adduct forms, located
 24 at -22.6 kcal mol⁻¹ (nitro in 4/syn) and -24.6 (4/anti), or -27.0 kcal mol⁻¹ (nitro in 2, syn and anti both).
 25 A possible HNO₃ loss from the anti nitroxy nitro **A8** intermediate (compare Ghigo et al. 2006) could
 26 produce 2-nitro anthracene, but this step results too demanding as a gas phase process (see Scheme
 27 S1 in the [Supplementary Material](#)), though in conditions different from the gas phase we cannot
 28 exclude such a step. This is consistent with the available experimental results (Kwok et al., 1994; see,
 29 in particular, p 791).

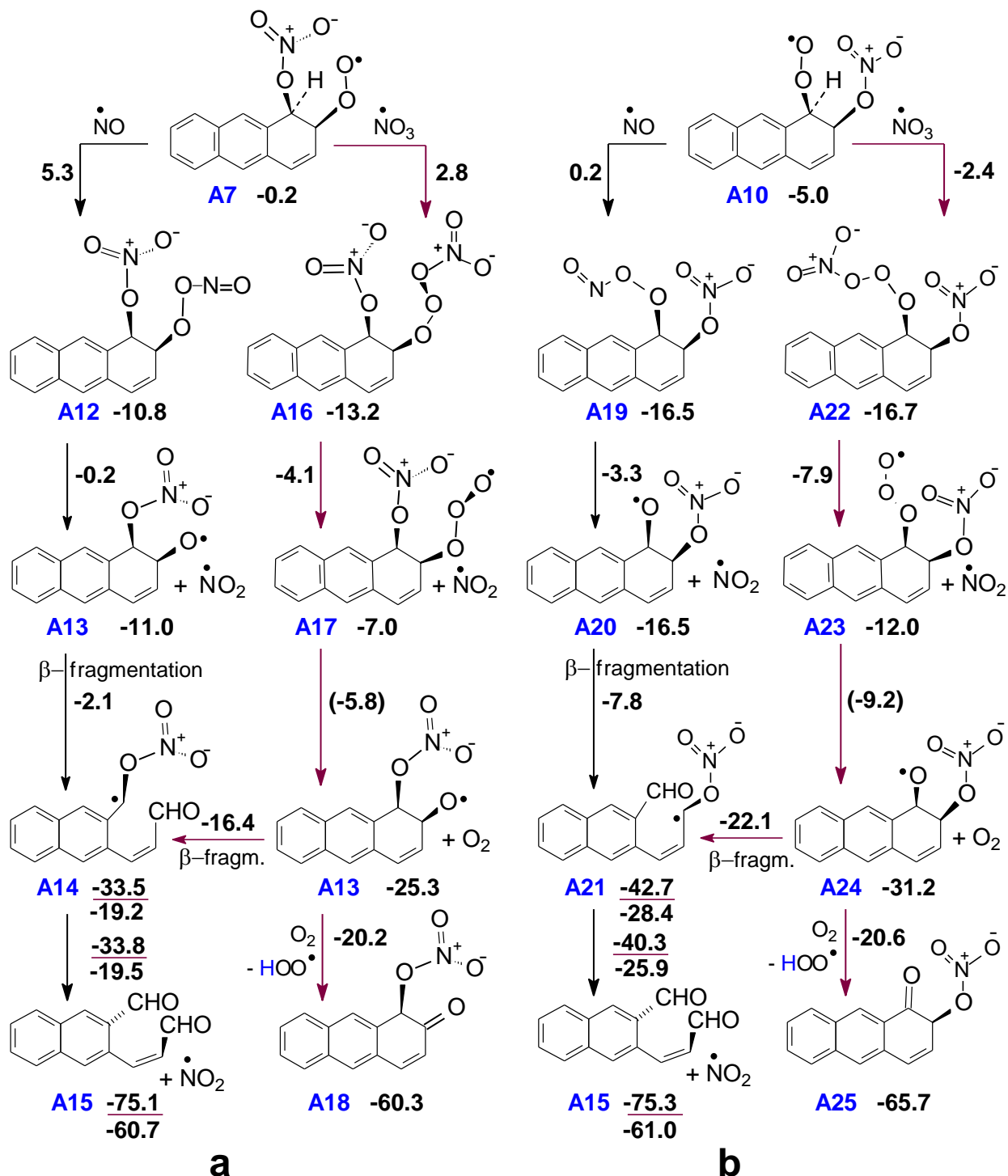
¹ For the radical coupling *phenyl + phenyl* → *biphenyl*, for instance, $k = 1.9 \times 10^{-11}$ molec⁻¹ cm³ s⁻¹ has been assessed at $T = 300$ K (Park and Lin, 1997). For *methyl + phenyl* → *toluene*, $k = 2.1 \times 10^{-11}$ molec⁻¹ cm³ s⁻¹ at $T = 300$ K (Tokmakov et al., 1999). But lower values are found in other cases, as in the presence of delocalized π radicals. For *cyclohexadienyl + NO* at 296 K the values are: $k = 6\text{--}12$ (P-dependent, in the range 20–80 Torr of He) $\times 10^{-13}$ molec⁻¹ cm³ s⁻¹ (Nakashima et al., 2008).

Values for the hydroperoxyl radical are also reported. Similar considerations apply to **A2** (see below). The extent to which bifunctional derivatives formation can compete, depends on the relative concentrations of O_2 and X , and on the accuracy of the relevant rate constants, hence it is rather hard to assess. Regarding **A6** formation, an alternative pathway (O_2 addition to give **A7**, H abstraction from the position geminal to the nitroxy group operated by $-O^*$, hydroperoxyl radical loss) would give the same aromatic nitrate (Scheme 1, pathway d). The barrier for intramolecular H abstraction is, for the syn isomer, much higher than for the step **A1-A6** (see Supplementary Material, section A). If the peroxy group is anti to the nitroxy, i.e. close to the hydrogen to be abstracted, it comes out by contrast to be lower: $19.6 \text{ kcal mol}^{-1}$ with respect to the preceding minimum (compare $22.8 \text{ kcal mol}^{-1}$ for the intermolecular process). In this case, hydroperoxyl loss is concerted with the H abstraction.

3.1.2 Nitroxy in position 2: initial steps. Similarly to **A1**, the two intermediates **A4** and **A5** can be obtained from **A2**, through ring closure steps (Scheme 1b). In **A2** the largest spin density is on C1, $\Delta\rho = 0.64$. It is somewhat higher than that associated to C2 in **A1**, 0.55: an allyl-like system similar to that found in **A1** is present, but with larger spin density on one carbon, C1. Accordingly, the ring closures entail comparable, but slightly smaller, barrier heights than for **A1**: $24.0 \text{ kcal mol}^{-1}$ to get **A4**, and $24.7 \text{ kcal mol}^{-1}$ to get **A5** from **A2** (which is $4.2 \text{ kcal mol}^{-1}$ above **A1**).² Steps c, d, and e follow again, and present barriers not too dissimilar from those of Scheme 1a, somewhat lower in the case of dioxygen addition. Also the peroxy radicals **A10** are more stable by ca. 5 kcal mol^{-1} than those labeled **A7**, with the two substituents in ortho. Dioxygen addition is in this case significantly less reversible (in itself) than for **A7**. The step free energy differences are generally lower by ca. 4 kcal mol^{-1} , which can be traced back to the $4.2 \text{ kcal mol}^{-1}$ energy difference of **A2** vs **A1** (while the step products are quite alike).

3.1.3 Further steps from the peroxy radicals. By considering further different radical additions to the peroxy intermediate **A7**, as those by NO, NO_2 , NO_3 , or HO, one can assess if its terminal oxygen can be successively abstracted in a viable way. Starting from the peroxy radical intermediate **A7** located at $-0.2 \text{ kcal mol}^{-1}$, some steps could be initiated by NO (Scheme 2a). NO could in principle abstract the terminal oxygen by an effective collision, to form the oxyl radical intermediate **A13** + NO_2 , at $-11.0 \text{ kcal mol}^{-1}$, but we could not find such a transition structure.

² In **A2**, the spin density $\Delta\rho$ is somewhat more localized with respect to **A1**. However, some delocalization occurs also in **A2**, involving a system C1(0.64)-C_{tert}-C9(0.33) ($\Delta\rho$ values here in parentheses). This situation bears some resemblance to an allylic system, but is less symmetric than in **A1**. All other $|\Delta\rho|$ values are lower than 0.09.

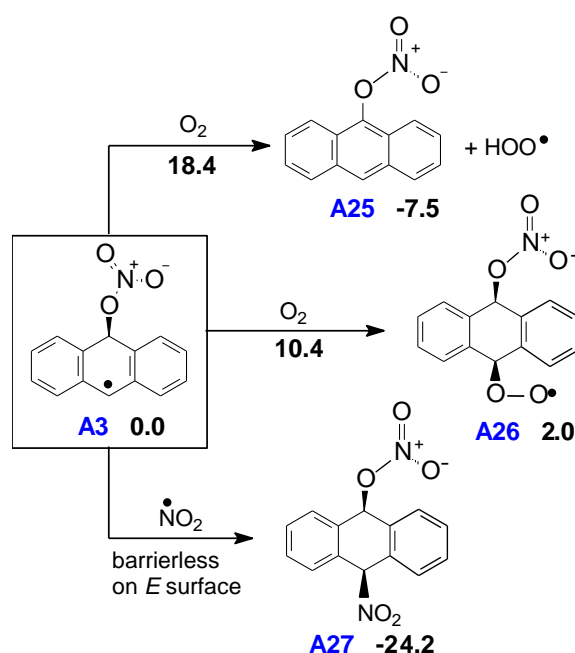


Scheme 2. Anthracene. Reaction steps of the peroxy radicals **A7** and **A10** with NO (black arrows) and NO₃ (red arrows; red-underlined figures pertain to this pathway when double values appear). See text and Scheme 1. Value in parentheses do not include the Yamaguchi treatment.³ Gibbs free energy differences at T=298.15 K, with respect to the initial adduct **A1**.

³ For both anthracene and phenanthrene, in the fragmentations of the trioxyl intermediates, **A17**, **A23**, **A31** (Scheme 4), and **P12** (Scheme 6), respectively, the earliness of the transition structures (in a geometrical sense: R_{O-O}=1.72-1.74 Å), causes the quartet, which is the next spin contaminant involved in the process of finding a stable wavefunction, to be quite high in energy. The ensuing Yamaguchi correction (Yamaguchi et al., 1988) comes out to be large enough to produce “negative barriers”. If we can cast some doubts on the real existence of a trioxyl intermediate, we can deem (in any case) this fragmentation a very easy step.

Alternatively, NO could first add to **A7** (radical coupling) to form the peroxyxynitrite **A12**. The addition step presents a low free energy barrier, 5.3 kcal mol⁻¹, and **A12** is located 10.6 kcal mol⁻¹ lower than **A7**. If (Scheme 2a, first column) a subsequent NO₂ loss from **A12** were feasible, it would then give **A13**, and by β-fragmentation **A14**, bearing an aldehydic group. **A14** is a radical stabilized by delocalization and α-effect. Loss of NO₂ would finally produce the closed shell dialdehyde **A15**.

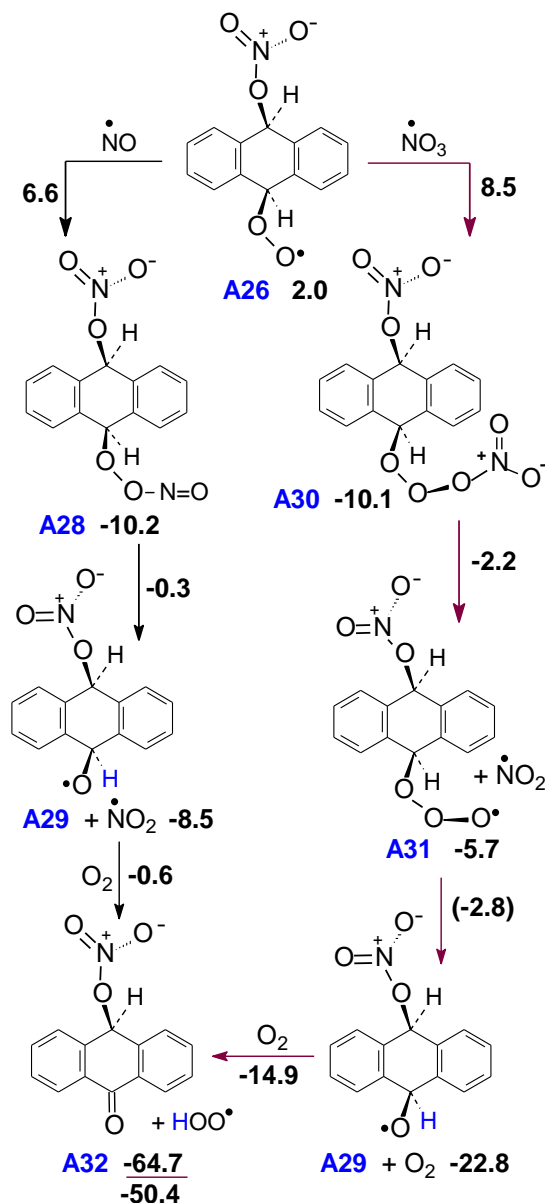
The pathway **A7-A13** shown in Scheme 2a (second column) resulted to be energetically the easiest. It leads to the dialdehyde **A15**, as well as to the closed shell nitroxyketone **A18** (red arrows): these can be considered as end products for the reacting system considered here. Other less fortunate attempts are documented in the Supplementary Material, section A. The nitrate-mediated pathway **A10-A20** opens the way to the formation of the same dialdehyde **A15** and a nitroxyketone **A25** (Scheme 2b).



Scheme 3. Anthracene. Initial steps involving positions 9 and 10. Gibbs free energy differences at T = 298.15 K, with respect to the initial adduct **A3**.

3.1.4 Nitroxy in position 9. In the case of the nitrate adduct having the –ONO₂ group in position 9, **A3**, which is 7.5 kcal mol⁻¹ below **A1**, three of the five cases of Scheme 1 can be considered (Scheme 3) because (a) an –O–N–O– bridge between two carbons forms only through a high barrier (Supplementary Material, Section A), and (b) an epoxide ring formation would cancel the aromaticity of one benzo group. Step **A3-A25** corresponds to pathway c in Scheme 1 and leads to a structural isomer of **A6**. Step **A3-A26** corresponds to pathway d. Step **A3-A27** is, as for pathway e, a coupling of two radicals (here exemplified by the formation of 10-nitro-9,10-dihydroanthr-9-yl nitrate. Further viable pathways from **A3** were not found. Steps relying on a β-fragmentation in an oxyl intermediate

1 similar to **A13** are to be ruled out because it would originate a phenyl radical, which proved to be
 2 too high in energy. From **A26**, further steps could lead to a very stable nitroxyketone. They are
 3 similar to those discussed in connection with Scheme 2 and are displayed in Scheme 4.

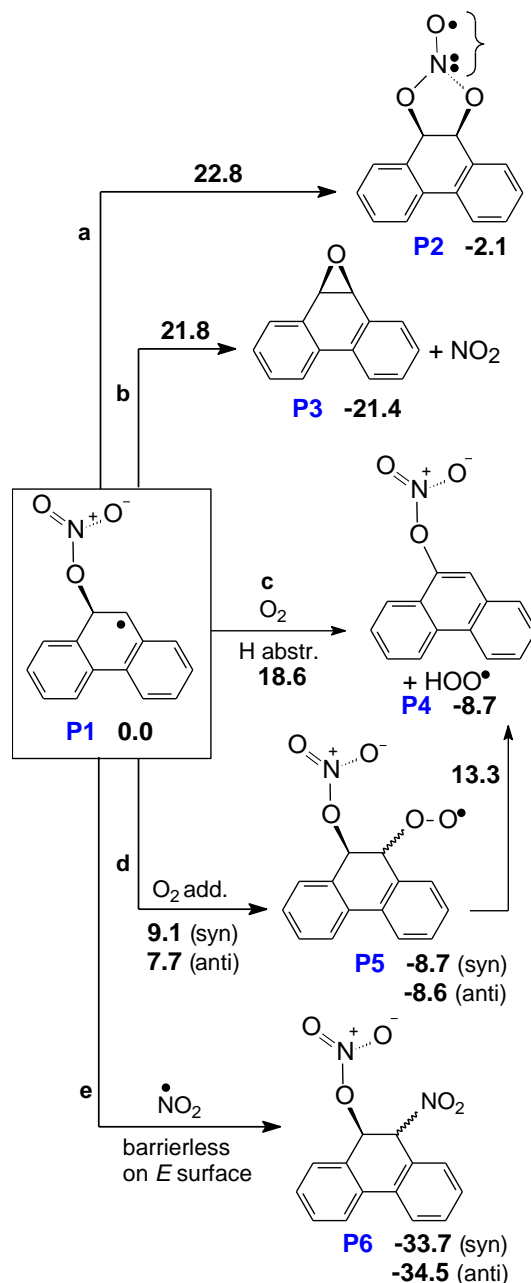


Scheme 4. Anthracene. Steps following initial O_2 addition to C10 in 9-nitroxyanthracene. Additions of NO (black arrows) and NO_3 (red arrows; red-underlined figures pertain to this pathway when double values appear). Value in parentheses do not include the Yamaguchi treatment.³ Gibbs free energy differences at $T = 298.15$ K, with respect to the initial adduct **A3**.

11 By comparing the two reaction channels, we see that NO addition to **A26** is less demanding than
 12 NO_3 addition, while the adducts **A28** and **A30** have similar stability. The oxyl radical **A29** that forms
 13 through further NO_2 or O_2 losses can undergo H abstraction by O_2 to give 9-nitroxy-anthr-10-one.
 14 This time a β -fragmentation similar to that seen in Scheme 2 cannot take place, because it implies
 15 formation of a rather unstable phenyl radical.

3.2 Phenanthrene. This PAH presents its equivalent positions 9 and 10 as the most reactive. NO_3 adds to these positions again without any energy barrier,¹ to form the 9-phenanthrenylnitrate radical **P1** (boxed structure, taken as our ΔG reference in [Scheme 5](#) and [Scheme 6](#)). **P1** could undergo a 5-membered ring formation (c-CCONO) to its radical isomer **P2**, step (a), with a barrier $\Delta G^\ddagger = 22.8 \text{ kcal mol}^{-1}$. This substantial barrier should be overcome, to take a pathway that, through a ring opening, leads to a dialdehyde (biphenyl-2,2'-dicarbaldehyde). Details are provided in the [Supplementary Material](#), section A. (b) As an alternative, **P1** could lose NO_2 , and concertedly close an epoxide ring to give **P3**, with $\Delta G^\ddagger = 21.8 \text{ kcal mol}^{-1}$. Both these pathways do not appear to be very likely, since they correspond to quite slow reaction steps, though less difficult than the analogous steps seen for anthracene. (c) Coming to bimolecular processes, O_2 could abstract the hydrogen geminal to the nitroxy group “on the fly”, with $\Delta G^\ddagger = 18.6 \text{ kcal mol}^{-1}$ and formation of 9-nitroxyphenanthrene, **P4**, an aryl nitrate. (d) Otherwise, O_2 could first add to the adjacent position 10. This step cannot be seen as reversible, in itself, as **A1-A7**, and is more similar in this respect to **A2-A10**, even with a higher barrier for the backwards step.

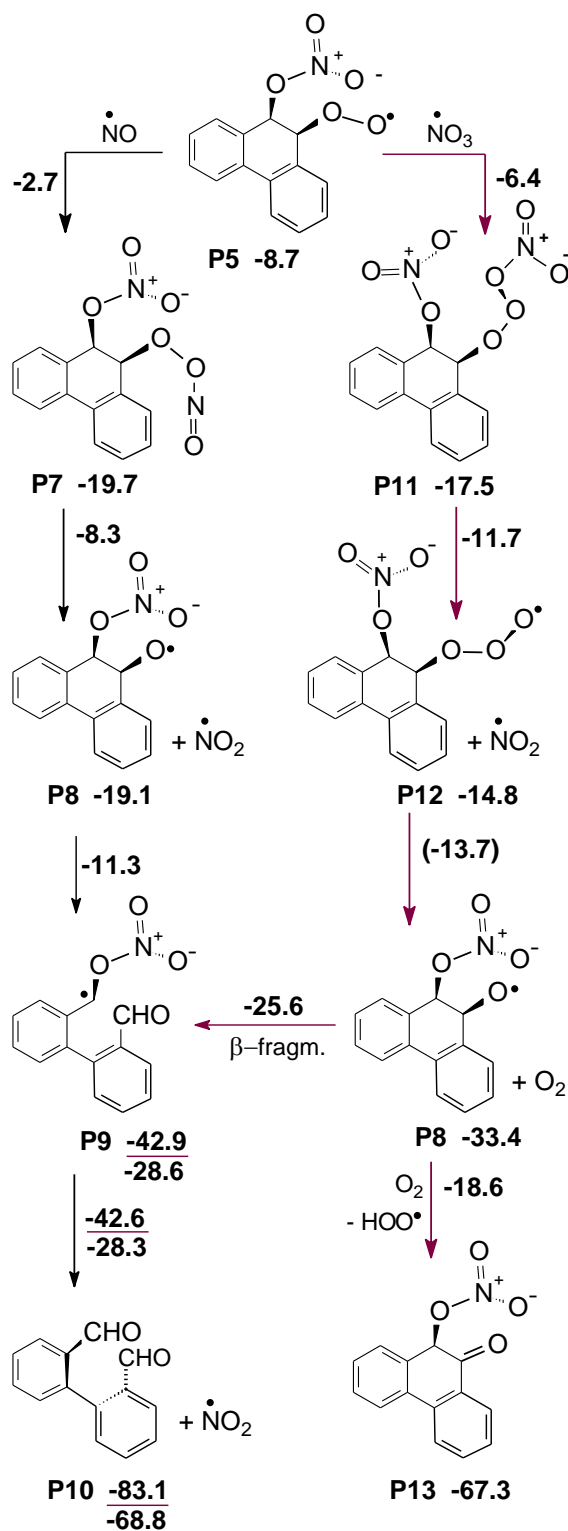
O_2 addition, as already discussed, can occur with syn or anti diastereochemistry. In the case of anti addition, the terminal peroxy $-\text{O}^\bullet$ can abstract the hydrogen geminal to the nitroxy group and then a hydroperoxyl radical can leave. If H abstraction were impossible because of a syn arrangement, it could also occur as NO-mediated process (details in the [Supplementary Material](#), section A). The barrier to the peroxy intermediate **P5** is $3.0 \text{ kcal mol}^{-1}$ high (2.2 if O_2 adds anti), and **P5** lies at -8.7 (syn) or -8.6 (anti) kcal mol^{-1} . Formation of 9-nitroxy-phenanthrene **P4** would ensue from both c and d pathways but with a second sizable barrier of ca. 22 kcal mol^{-1} with respect to **P5** along pathway d. Other possible reaction steps from **P5** are discussed below ([Scheme 6](#)). (e) Any radical tropospheric species (as NO_3 , NO_2 , NO ,...) could also add with the same diastereochemistry, and closed shell products would form. Here this termination step is exemplified in [Scheme 5](#) by NO_2 addition, which takes place via a barrierless step as regards energy.



1

2 **Scheme 5.** Phenanthrene. Initial steps involving positions 9 and 10. Gibbs free energy differences at
 3 $T = 298.15 \text{ K}$, with respect to the initial adduct **P1**. In **P2**, the brace indicates delocalization. Pathway e:
 4 NO_2 taken as an example of any termination step occurring via radical coupling.

5 Similarly to what described for anthracene, starting from the peroxy radical intermediate **P5**, at
 6 $-8.7 \text{ kcal mol}^{-1}$, further steps, initiated by NO , can in principle form an oxyl radical intermediate **P8** +
 7 NO_2 at $-19.1 \text{ kcal mol}^{-1}$. However, we found no trace of a TS for O abstraction by NO “on the fly”.
 8 Instead, the 9-nitroso-10-peroxynitrite **P7** first forms upon radical coupling (Scheme 6, left), with a
 9 barrier of $6.0 \text{ kcal mol}^{-1}$; **P7** is located at $-19.7 \text{ kcal mol}^{-1}$. Upon NO_2 loss, it would give the oxyl
 10 intermediate **P8**. If **P8** forms this way, further steps through a stabilized radical **P11** can finally give
 11 the closed shell dialdehyde, **P10**. **P8** could also form through the NO_3 -addition pathway shown on
 12 the right. The dialdehyde **P12**, and to a slightly lesser extent the nitroso ketone **P13**, appear to be
 13 good candidates for being major products of the NO_3 -mediated oxidation.



1

2 **Scheme 6.** Phenanthrene. Additions of NO (left, black arrows) and NO₃ (right, red arrows; red-
 3 underlined figures pertain to this pathway when double values appear). Value in parentheses do not
 4 include the Yamaguchi treatment.³ Gibbs free energy differences at T = 298.15 K, with respect to the
 5 initial adduct **P1**.

6 The reaction of phenanthrene with HO and NO₃ radicals was studied experimentally (Kwok et
 7 al., 1994). In that study a kinetic constant of 1.2 × 10⁻¹³ molecules⁻¹ cm³ s⁻¹ for NO₃ (1.3 × 10⁻¹¹ for HO)
 8 was determined, concluding that these reactions should dominate phenanthrene atmospheric loss
 9 processes. Their experiments were carried out in the presence of a varying initial concentration of

NO₂, namely 0, 4.4×10^{13} and 2.4×10^{14} molecules cm⁻³. A non-zero intercept in a rate plot vs NO₂, concentration suggested that the phenanthrene-nitrate radical reaction must also involve a pathway independent of NO₂ concentration. This pathway could actually be a set of pathways of different importance, as shown in [Schemes 1, 3, 5](#).

The results of our present study can be compared with those from the recent theoretical study on phenanthrene plus HO, O₂, and NO_x (Zhao et al., 2016). In those studies, the principal products were indicated to be oxidized ring retaining products, as well as products from ring opening, namely, phenantrones, phenanthrols, quinones and dialdehydes. The same research group had previously investigated the reaction of the nitrate radical with 9-chloro-anthracene in the presence of O₂ and NO_x (Dang et al., 2015c), and the same kind of reactions involving naphthalene too (Qu et al., 2006c). They found that, following barrierless nitrate addition to the PAH π -system, the main products should be: 9-chloroanthryl nitrates, diones, epoxides, and dialdehydes.

4. Conclusions

In this study, we have investigated the NO₃-initiated gas phase oxidative degradation of anthracene and phenanthrene. A variety of reaction steps have been considered that contemplate the intervention of O₂, NO, NO₂, and also NO₃ (for a second time), and also unimolecular steps. Some reaction steps, considered less likely, are collected in the [Supplementary Material](#). These results indicate that one NO₃-initiated pathway, which is also O₂ and NO₃-mediated, leads to the formation of a dialdehyde and a nitroxy-ketone. The same kind of pathway is effective for both anthracene ([Schemes 2 and 4](#)) and phenanthrene ([Scheme 6](#)). The key species is an early intermediate, a nitroxy peroxy radical that forms after the initial NO₃ attack to the PAH upon dioxygen addition. It further evolves to carbonyl products, through free energy barriers that are significantly lower than those related to other possible intermediates or closed shell products. Some bifunctional PAH derivatives can also form through couplings of the initial nitroxy radical adducts with free radicals X present in the troposphere. The example of the formation of a nitroxy nitro product is shown (pathway e in [Schemes 1, 3, 5](#)). HNO₃ loss from the closed shell anti nitroxy nitro intermediate could allow the formation of a nitroaromatic. However, it but results very difficult in the gas phase.

Acknowledgments. This work was supported by the Università di Torino (Local Funding 2016). It was conducted in the frame of EC FP6 NoE ACCENT and ACCENT–PLUS projects (Atmospheric Composition Change, the European NeTWork of Excellence).

Supplementary Material for this article is available: it includes the geometries and energetics of all optimized structures, plus the illustration of other reaction channels explored but not reported in the main text because of their more demanding energetics.

References

- Andersson, J.T., Achten, C., 2015. Time to Say Goodbye to the 16 EPA PAHs? Toward an Up-to-Date Use of PACs for Environmental Purposes. *Polycycl. Aromat. Compd.* 35, 330–354. doi:10.1080/10406638.2014.991042
- Arey, J., Zielinska, B., Atkinson, R., Aschmann, S.M., 1989. Nitroarene products from the gas-phase reactions of volatile polycyclic aromatic hydrocarbons with the OH radical and N₂O₅. *Int. J. Chem. Kinet.* 21, 775–799. doi:10.1002/kin.550210906
- Asare, N., Landvik, N.E., Lagadic-Gossmann, D., Rissel, M., Tekpli, X., Ask, K., Låg, M., Holme, J.A., 2008. 1-Nitropyrene (1-NP) induces apoptosis and apparently a non-apoptotic programmed cell death (paraptosis) in Hepa1c1c7 cells. *Toxicol. Appl. Pharmacol.* 230, 175–186. doi:10.1016/j.taap.2008.02.015
- Asare, N., Tekpli, X., Rissel, M., Solhaug, A., Landvik, N., Lecureur, V., Podechard, N., Brunborg, G., Lag, M., Lagadic-Gossmann, D., Holme, J.A., 2009. Signalling pathways involved in 1-nitropyrene (1-NP)-induced and 3-nitrofluoranthene (3-NF)-induced cell death in Hepa1c1c7 cells. *Mutagenesis* 24, 481–493. doi:10.1093/mutage/gep032
- Atkinson, R., 1991. Kinetics and Mechanisms of the Gas-Phase Reactions of the NO₃ Radical with Organic Compounds. *J. Phys. Chem. Ref. Data* 20, 459. doi:10.1063/1.555887
- Atkinson, R., Arey, J., Zielinska, B., Aschmann, S.M., 1990. Kinetics and nitro-products of the gas-phase OH and NO₃ radical-initiated reactions of naphthalene-d₈, Fluoranthene-d₁₀, and pyrene. *Int. J. Chem. Kinet.* 22, 999–1014. doi:10.1002/kin.550220910
- Baker, J., Scheiner, A., Andzelm, J., 1993. Spin contamination in density functional theory. *Chem. Phys. Lett.* 216, 380–388. doi:10.1016/0009-2614(93)90113-f
- Biermann, H.W., Leod, H. Mac, Atkinson, R., Winer, A.M., Pitts, J.N., 1985. Kinetics of the gas-phase reactions of the hydroxyl radical with naphthalene, phenanthrene, and anthracene. *Environ. Sci. Technol.* 19, 244–248. doi:10.1021/es00133a004
- Brubaker, W.W., Hites, R.A., 1998. OH Reaction Kinetics of Polycyclic Aromatic Hydrocarbons and Polychlorinated Dibenzo- p -dioxins and Dibenzofurans. *J. Phys. Chem. A* 102, 915–921. doi:10.1021/jp9721199
- Calvert, J. G., Atkinson, R., Becker, K. H., Kamens, R. M., Seinfeld, J. H., Wallington, T. J., Yarwood, G., 2002. The mechanism of atmospheric oxidation of aromatic hydrocarbons. Oxford University Press. Ch 3: Reactions of mono- and polycyclic aromatic compounds with NO₃ radicals, pp 135–166. ISBN: 9780195146288
- Chapleski C. Jr., Zhang, Y., Troya, D., Morris, J. R., 2016. Heterogeneous chemistry and reaction dynamics of the atmospheric oxidants, O₃, NO₃, and OH, on organic surfaces. *Chem. Soc. Rev.*, 45, 3731–3746.
- Cochran, R. E., Jeong, H., Haddadi, S., Fisseha Derseh, R., Gowan, A., Beránek, J., Kubátová, A., 2016. Identification of products formed during the heterogeneous nitration and ozonation of polycyclic aromatic hydrocarbons. *Atmos. Environ.* 128, 92–103.
- Dang, J., Shi, X., Hu, J., Chen, J., Zhang, Q., Wang, W., 2015a. Mechanistic and kinetic studies on OH-initiated atmospheric oxidation degradation of benzo[a]pyrene in the presence of O₂ and NO_x. *Chemosphere* 119, 387–393. doi:10.1016/j.chemosphere.2014.07.001

- 1 Dang, J., Shi, X., Zhang, Q., Hu, J., Chen, J., Wang, W., 2014. Mechanistic and kinetic studies on the
2 OH-initiated atmospheric oxidation of fluoranthene. *Sci. Total Environ.* 490, 639–646.
3 doi:10.1016/j.scitotenv.2014.04.134
- 4 Dang, J., Shi, X., Zhang, Q., Hu, J., Wang, W., 2015b. Mechanism and kinetic properties for the OH-
5 initiated atmospheric oxidation degradation of 9,10-Dichlorophenanthrene. *Sci. Total Environ.*
6 505, 787–794. doi:10.1016/j.scitotenv.2014.10.081
- 7 Dang, J., Shi, X., Zhang, Q., Hu, J., Wang, W., 2015c. Insights into the mechanism and kinetics of the
8 gas-phase atmospheric reaction of 9-chloroanthracene with NO₃ radical in the presence of NO_x.
9 *RSC Adv.* 5, 84066–84075. doi:10.1039/c5ra11918a
- 10 Durant, J.L., Busby, W.F., Lafleur, A.L., Penman, B.W., Crespi, C.L., 1996. Human cell mutagenicity of
11 oxygenated, nitrated and unsubstituted polycyclic aromatic hydrocarbons associated with
12 urban aerosols. *Mutat. Res. Toxicol.* 371, 123–157. doi:10.1016/s0165-1218(96)90103-2
- 13 Fiedler, H., Mücke, W., 1991. Nitro Derivatives of Polycyclic Aromatic Hydrocarbons (NO₂-PAH), in:
14 Anthropogenic Compounds. Springer Berlin Heidelberg, Berlin, Heidelberg, pp. 97–137.
15 doi:10.1007/978-3-540-46757-1_2
- 16 Finlayson-Pitts, B.J., 1997. Tropospheric Air Pollution: Ozone, Airborne Toxics, Polycyclic Aromatic
17 Hydrocarbons, and Particles. *Science* 276, 1045–1051. doi:10.1126/science.276.5315.1045
- 18 Finlayson-Pitts, B.J., Pitts, J.N., 2000. Chemistry of the Upper and Lower Atmosphere, in: Press, A.
19 (Ed.), Chemistry of the Upper and Lower Atmosphere. Elsevier, pp. 436–546.
- 20 Frisch, M.J., Trucks, G.W., Schlegel, H.B., Scuseria, G.E., Robb, M.A., Cheeseman, J.R., Scalmani, G.,
21 Barone, V., Mennucci, B., Petersson, G.A., Nakatsuji, H., Caricato, M., Li, X., Hratchian, H.P.,
22 Izmaylov, A.F., Bloino, J., Zheng, G., Sonnenberg, J.L., Hada, M., Ehara, M., Toyota, K., Fukuda,
23 R., Hasegawa, J., Ishida, M., Nakajima, T., Honda, Y., Kitao, O., Nakai, H., Vreven, T., Montgomery
24 Jr., J.A., Peralta, J.E., Ogliaro, F., Bearpark, M., Heyd, J.J., Brothers, E., Kudin, K.N., Staroverov,
25 V.N., Kobayashi, R., Normand, J., Raghavachari, K., Rendell, A., Burant, J.C., Iyengar, S.S., Tomasi,
26 J., Cossi, M., Rega, N., Millam, J.M., Klene, M., Knox, J.E., Cross, J.B., Bakken, V., Adamo, C.,
27 Jaramillo, J., Gomperts, R., Stratmann, R.E., Yazyev, O., Austin, A.J., Cammi, R., Pomelli, C.,
28 Ochterski, J.W., Martin, R.L., Morokuma, K., Zakrzewski, V.G., Voth, G.A., Salvador, P.,
29 Dannenberg, J.J., Dapprich, S., Daniels, A.D., Farkas, Ö., Foresman, J.B., Ortiz, J. V, Cioslowski, J.,
30 Fox, D.J., 2009. Gaussian 09 Revision A.02.
- 31 Ghigo, G., Causà, M., Maranzana, A., Tonachini, G., 2006. Aromatic Hydrocarbon Nitration under
32 Tropospheric and Combustion Conditions. A Theoretical Mechanistic Study. *J. Phys. Chem. A*
33 110, 13270–13282. doi:10.1021/jp064459c
- 34 Gibson, T.L., 1982. Nitro derivatives of polynuclear aromatic hydrocarbons in airborne and source
35 particulate matter. *Atmos. Environ.* 16, 2037–2040. doi:10.1016/0004-6981(82)90490-5
- 36 Giordana, A., Maranzana, A., Ghigo, G., Causà, M., Tonachini, G., 2008. Soot Platelets and PAHs with
37 an Odd Number of Unsaturated Carbon Atoms and π Electrons: Theoretical Study of Their Spin
38 Properties and Interaction with Ozone. *J. Phys. Chem. A* 112, 973–982. doi:10.1021/jp075487d
- 39 Goldstein, E., Beno, B., Houk, K.N., 1996. Density Functional Theory Prediction of the Relative
40 Energies and Isotope Effects for the Concerted and Stepwise Mechanisms of the Diels-Alder
41 Reaction of Butadiene and Ethylene. *J. Am. Chem. Soc.* 118, 6036–6043. doi:10.1021/ja9601494
- 42 Helmig, D., Harger, W.P., 1994. OH radical-initiated gas-phase reaction products of phenanthrene.
43 *Sci. Total Environ.* 148, 11–21. doi:10.1016/0048-9697(94)90368-9
- 44 Jariyasopit, N., McIntosh, M., Zimmermann, K., Arey, J., Atkinson, R., Cheong, P.H.-Y., Carter, R.G.,
45 Yu, T.-W., Dashwood, R.H., Simonich, S.L.M., 2014a. Novel Nitro-PAH Formation from
46 Heterogeneous Reactions of PAHs with NO₂, NO₃/N₂O₅, and OH Radicals: Prediction,
47 Laboratory Studies, and Mutagenicity. *Environ. Sci. Technol.* 48, 412–419. doi:10.1021/
48 es4043808
- 49 Jariyasopit, N., Zimmermann, K., Schrlau, J., Arey, J., Atkinson, R., Yu, T.-W., Dashwood, R.H., Tao, S.,
50 Simonich, S.L.M., 2014b. Heterogeneous Reactions of Particulate Matter-Bound PAHs and
51 NPAHs with NO₃/N₂O₅, OH Radicals, and O₃ under Simulated Long-Range Atmospheric

- 1 Transport Conditions: Reactivity and Mutagenicity. *Environ. Sci. Technol.* 48, 10155–10164.
2 doi:10.1021/es5015407
- 3 Jenkins, B.M., Jones, A.D., Turn, S.Q., Williams, R.B., 1996. Emission Factors for Polycyclic Aromatic
4 Hydrocarbons from Biomass Burning. *Environ. Sci. Technol.* 30, 2462–2469.
5 doi:10.1021/es950699m
- 6 Kameda, T., 2011. Atmospheric Chemistry of Polycyclic Aromatic Hydrocarbons and Related
7 Compounds. *J. Health Sci. (Minireview)* , 57, 504–511.
- 8 Kendall, R.A., Dunning, T.H., Harrison, R.J., 1992. Electron affinities of the first-row atoms revisited.
9 Systematic basis sets and wave functions. *J. Chem. Phys.* 96, 6796–6806. doi:10.1063/1.462569
- 10 Keyte, I.J., Harrison, R.M., Lammel, G., 2013. Chemical reactivity and long-range transport potential of
11 polycyclic aromatic hydrocarbons. A review. *Chem. Soc. Rev.* 42, 9333. doi:10.1039/c3cs60147a
- 12 Kwok, E.S.C., Harger, W.P., Arey, J., Atkinson, R., 1994. Reactions of Gas-Phase Phenanthrene under
13 Simulated Atmospheric Conditions. *Environ. Sci. Technol.* 28, 521–527. doi:10.1021/
14 es00052a027
- 15 Lammel, G., 2015. Polycyclic Aromatic Compounds in the Atmosphere – A Review Identifying
16 Research Needs. *Polycycl. Aromat. Compd.* 35, 316–329. doi:10.1080/10406638.2014.931870
- 17 Landvik, N., Gorria, M., Arlt, V., Asare, N., Solhaug, A., Lagadicgossmann, D., Holme, J., 2007. Effects
18 of nitrated-polycyclic aromatic hydrocarbons and diesel exhaust particle extracts on cell
19 signalling related to apoptosis: Possible implications for their mutagenic and carcinogenic
20 effects. *Toxicology* 231, 159–174. doi:10.1016/j.tox.2006.12.009
- 21 Lim, L.H., Harrison, R.M., Harrad, S., 1999. The Contribution of Traffic to Atmospheric Concentrations
22 of Polycyclic Aromatic Hydrocarbons. *Environ. Sci. Technol.* 33, 3538–3542.
23 doi:10.1021/es990392d
- 24 Liu, G., Zhang, P., Yang, B., Wang, Y., Shu, J., 2012. Kinetic Studies of Heterogeneous Reactions of
25 Polycyclic Aromatic Hydrocarbon Aerosols with NO₃ Radicals. *Environ. Sci. Technol.* 46,
26 7575–7580. dx.doi.org/10.1021/es301403d
- 27 Mak, J., Gross, S., Bertram A. K., 2007. Uptake of NO₃ on soot and pyrene surfaces. *Geophys Res.*
28 *Lett.*, 34, L10804, doi:10.1029/2007GL029756.
- 29 Mandalakis, M., Gustafsson, Å., Alsberg, T., Egeback, A.-L., Reddy, C.M., Xu, L., Klanova, J.,
30 Holoubek, I., Stephanou, E.G., 2005. Contribution of Biomass Burning to Atmospheric Polycyclic
31 Aromatic Hydrocarbons at Three European Background Sites. *Environ. Sci. Technol.* 39, 2976–
32 2982. doi:10.1021/es048184v
- 33 Nakashima, Y., Xing, J.-H., Inomata, S., Imamura, T., 2008. Rate coefficients for the reactions of
34 cyclohexadienyl (c-C₆H₇) radicals with O₂ and NO at room temperature. *Chem. Phys. Lett.* 450,
35 248–252. doi:10.1016/j.cplett.2007.11.044
- 36 Park, J., Lin, M.C., 1997. Kinetics for the Recombination of Phenyl Radicals. *J. Phys. Chem. A* 101, 14–
37 18. doi:10.1021/jp961569i
- 38 Parr, R.G., Yang, W., 1989. Density-Functional Theory of Atoms and Molecules, Density-Functional
39 Theory of Atoms and Molecules. Oxford University Press.
- 40 Perraudin, E., Budzinski, H., Villenave, E., 2007. Identification and quantification of ozonation
41 products of anthracene and phenanthrene adsorbed on silica particles. *Atmos. Environ.* 41,
42 6005–6017. doi:10.1016/j.atmosenv.2007.03.010
- 43 Perraudin, E., Budzinski, H., Villenave, E., 2006. Kinetic Study of the Reactions of Ozone with
44 Polycyclic Aromatic Hydrocarbons Adsorbed on Atmospheric Model Particles. *J. Atmos. Chem.*
45 56, 57–82. doi:10.1007/s10874-006-9042-x
- 46 Pitts, J., Cauwenberghe, K. Van, Grosjean, D., Schmid, J., Fitz, D., Belser, W., Knudson, G., Hynds, P.,
47 1978. Atmospheric reactions of polycyclic aromatic hydrocarbons: facile formation of
48 mutagenic nitro derivatives. *Science (80-)*. 202, 515–519. doi:10.1126/science.705341
- 49 Pople, J.A., Gill, P.M.W.W., Johnson, B.G., 1992. Kohn-Sham density-functional theory within a finite
50 basis set. *Chem. Phys. Lett.* 199, 557–560. doi:10.1016/0009-2614(92)85009-y
- 51 Ringuet, J., Albinet, A., Leoz-Garziandia, E., Budzinski, H., Villenave, E., 2012. Reactivity of polycyclic

- 1 aromatic compounds (PAHs, NPAHs and OPAHs) adsorbed on natural aerosol particles exposed
- 2 to atmospheric oxidants. *Atmos. Environ.* 61, 15-22.
- 3 Qu, X., Zhang, Q., Wang, W., 2008. Theoretical study on NO₃-initiated oxidation of acenaphthene in
- 4 the atmosphere. *Can. J. Chem.* 86, 129–137. doi:10.1139/v07-137
- 5 Qu, X., Zhang, Q., Wang, W., 2006a. Degradation mechanism of benzene by NO₃ radicals in the
- 6 atmosphere: A DFT study. *Chem. Phys. Lett.* 426, 13–19. doi:10.1016/j.cplett.2006.05.070
- 7 Qu, X., Zhang, Q., Wang, W., 2006b. Theoretical study on mechanism for NO₃-initiated atmospheric
- 8 oxidation of naphthalene. *Chem. Phys. Lett.* 432, 40–49. doi:10.1016/j.cplett.2006.10.041
- 9 Qu, X., Zhang, Q., Wang, W., 2006c. Theoretical study on mechanism for NO₃-initiated atmospheric
- 10 oxidation of naphthalene. *Chem. Phys. Lett.* 432, 40–49. doi:10.1016/j.cplett.2006.10.041
- 11 Samburova, V., Connolly, J., Gyawali, M., Yatavelli, R.L.N., Watts, A.C., Chakrabarty, R.K., Zielinska,
- 12 B., Moosmüller, H., Khlystov, A., 2016. Polycyclic aromatic hydrocarbons in biomass-burning
- 13 emissions and their contribution to light absorption and aerosol toxicity. *Sci. Total Environ.* 568,
- 14 391–401. doi:10.1016/j.scitotenv.2016.06.026
- 15 Sasaki, J., Aschmann, S.M., Kwok, E.S.C., Atkinson, R., Arey, J., 1997. Products of the Gas-Phase OH
- 16 and NO₃ Radical-Initiated Reactions of Naphthalene. *Environ. Sci. Technol.* 31, 3173–3179.
- 17 doi:10.1021/es9701523
- 18 Schlegel, H.B., 1982a. An efficient algorithm for calculating a b i n i t i o energy gradients using s , p
- 19 Cartesian Gaussians. *J. Chem. Phys.* 77, 3676–3681. doi:10.1063/1.444270
- 20 Schlegel, H.B., 1982b. Optimization of equilibrium geometries and transition structures. *J. Comput.*
- 21 *Chem.* 3, 214–218. doi:10.1002/jcc.540030212
- 22 Schlegel, H.B., 1981. Ab Initio Energy Derivatives Calculated Analytically, in: Csizmadia, I.G., Daudel,
- 23 R. (Eds.), *Computational Theoretical Organic Chemistry*. Springer Netherlands, Dordrecht, pp.
- 24 129–159. doi:10.1007/978-94-009-8472-1
- 25 Schlegel, H.B., Binkley, J.S., Pople, J.A., 1984. First and second derivatives of two electron integrals
- 26 over Cartesian Gaussians using Rys polynomials. *J. Chem. Phys.* 80, 1976–1981.
- 27 doi:10.1063/1.446960
- 28 Stone, D., Evans, M.J., Walker, H., Ingham, T., Vaughan, S., Ouyang, B., Kennedy, O.J., McLeod, M.W.,
- 29 Jones, R.L., Hopkins, J., Punjabi, S., Lidster, R., Hamilton, J.F., Lee, J.D., Lewis, A.C., Carpenter,
- 30 L.J., Forster, G., Oram, D.E., Reeves, C.E., Bauguitte, S., Morgan, W., Coe, H., Aruffo, E., Dari-
- 31 Salisburgo, C., Giammaria, F., Di Carlo, P., Heard, D.E., 2014. Radical chemistry at night:
- 32 comparisons between observed and modelled HO_x, NO₃ and N₂O₅ during the RONOCO project.
- 33 *Atmos. Chem. Phys.* 14, 1299–1321. doi:10.5194/acp-14-1299-2014
- 34 Talaska, G., Underwood, P., Maier, A., Lewtas, J., Rothman, N., Jaeger, M., 1996. Polycyclic Aromatic
- 35 Hydrocarbons (PAHs), Nitro-PAHs and Related Environmental Compounds: Biological Markers
- 36 of Exposure and Effects. *Environ. Health Perspect.* 104, 901. doi:10.2307/3433008
- 37 Tokmakov, I. V., Park, J., Gheys, S., Lin, M.C., 1999. Experimental and Theoretical Studies of the
- 38 Reaction of the Phenyl Radical with Methane. *J. Phys. Chem. A* 103, 3636–3645.
- 39 doi:10.1021/jp983140s
- 40 Trogolo, D., Maranzana, A., Ghigo, G., Tonachini, G., 2014. First Ring Formation by Radical Addition
- 41 of Propargyl to But-1-ene-3-yne in Combustion. Theoretical Study of the C₇H₇ Radical System.
- 42 *J. Phys. Chem. A* 118, 427–440. doi:10.1021/jp4082905
- 43 Wittbrodt, J.M., Schlegel, H.B., 1996. Some reasons not to use spin projected density functional
- 44 theory. *J. Chem. Phys.* 105, 6574. doi:10.1063/1.472497
- 45 Yamaguchi, K., Jensen, F., Dorigo, A., Houk, K.N., 1988. A spin correction procedure for unrestricted
- 46 Hartree-Fock and Møller-Plesset wavefunctions for singlet diradicals and polyradicals. *Chem.*
- 47 *Phys. Lett.* 149, 537–542. doi:10.1016/0009-2614(88)80378-6
- 48 Yamanaka, S., Kawakami, T., Nagao, H., Yamaguchi, K., 1994. Effective exchange integrals for open-
- 49 shell species by density functional methods. *Chem. Phys. Lett.* 231, 25–33. doi:10.1016/0009-
- 50 2614(94)01221-0
- 51 Yan, X., Weng, Y., Li, S., Tao, F.-M., 2013. Troposphere reactions of hydroxycyclohexadienyl peroxy

- 1 radicals with nitric oxide: A DFT study. *Comput. Theor. Chem.* 1018, 6–12.
- 2 doi:10.1016/j.comptc.2013.05.024
- 3 Zhang, Y., Shu, J., Zhang, Y., Yang, B., 2013. Homogeneous and heterogeneous reactions of
- 4 anthracene with selected atmospheric oxidants. *J. Environ. Sci.* 25, 1817–1823.
- 5 doi:10.1016/s1001-0742(12)60233-2
- 6 Zhang, P., Wanqi Sun, W., Li, N., Wang, Y., Shu, J., Yang, B., Dong, L., 2014. Effects of Humidity and
- 7 [NO₃]/[N₂O₅] Ratio on the Heterogeneous Reaction of Fluoranthene and Pyrene with
- 8 N₂O₅/NO₃/NO₂. *Environ. Sci. Technol.* 2014, 48, 13130–13137 dx.doi.org/10.1021/es504508v
- 9 Zhang, Q., Rui Gao, R., Xu, F., Zhou, Q., Jiang, G., Wang, T., Chen, J., Hu, J., Jiang, W., Wang, W., 2014.
- 10 Role of Water Molecule in the Gas-Phase Formation Process of Nitrated Polycyclic Aromatic
- 11 Hydrocarbons in the Atmosphere: A Computational Study. *Environ. Sci. Technol.* 48, 5051–5057.
- 12 dx.doi.org/10.1021/ es500453g
- 13 Zhang, Y., Yang, B., Gan, J., Liu, C., Shu, X., Shu, J., 2011. Nitration of particle-associated PAHs and
- 14 their derivatives (nitro-, oxy-, and hydroxy-PAHs) with NO₃ radicals. *Atmos. Environ.* 45, 2515 -
- 15 2521.
- 16 Zhao, N., Zhang, Q., Wang, W., 2016. Atmospheric oxidation of phenanthrene initiated by OH radicals
- 17 in the presence of O₂ and NO_x - A theoretical study. *Sci. Total Environ.* 563–564, 1008–1015.
- 18 doi:10.1016/j.scitotenv.2016.01.089
- 19 Zhao, Y., Truhlar, D.G., 2008. Density functionals with broad applicability in chemistry. *Acc. Chem.*
- 20 *Res.* 41, 157–67.
- 21 Zhao, Y., Truhlar, D.G., 2007. The M06 suite of density functionals for main group thermochemistry,
- 22 thermochemical kinetics, noncovalent interactions, excited states, and transition elements: two
- 23 new functionals and systematic testing of four M06-class functionals and 12 other function.
- 24 *Theor. Chem. Acc.* 120, 215–241. doi:10.1007/s00214-007-0310-x
- 25 Zimmermann, K., Jariyasopit, N., Simonich, S.L.M., Tao, S., Atkinson, R., Arey, J., 2013. Formation of
- 26 Nitro-PAHs from the Heterogeneous Reaction of Ambient Particle-Bound PAHs with
- 27 N₂O₅/NO₃/NO₂. *Environ. Sci. Technol.* 130718154506004. doi:10.1021/es401789x
- 28
- 29