

AperTO - Archivio Istituzionale Open Access dell'Università di Torino

Winter-green host-plants, litter quantity and vegetation structure are key determinants of habitat quality for *Coenonympha oedippus* in Europe

This is a pre print version of the following article:

Original Citation:

Availability:

This version is available <http://hdl.handle.net/2318/1612720> since 2017-05-16T11:06:10Z

Published version:

DOI:10.1007/s10841-014-9736-3

Terms of use:

Open Access

Anyone can freely access the full text of works made available as "Open Access". Works made available under a Creative Commons license can be used according to the terms and conditions of said license. Use of all other works requires consent of the right holder (author or publisher) if not exempted from copyright protection by the applicable law.

(Article begins on next page)

1 **Winter-green host-plants, litter quantity and vegetation structure are key determinants of habitat quality**
2 **for *Coenonympha oedippus* in Europe**

3
4
5
6 4 Tatjana Čelik¹; Markus Bräu², Simona Bonelli³, Cristiana Cerrato³, Branko Vreš¹, Emilio Balletto³, Christian
7 Stettmer⁴, Matthias Dolek⁵

8
9
10
11
12 7 ¹Jovan Hadži Institute of Biology, Scientific Research Centre of the Slovenian Academy of Sciences and Arts,
13 Novi trg 2, P. O. Box 306, 1001 Ljubljana, Slovenia

14
15
16 9 ²Büro für ökologische Gutachten M. Bräu, Amperstr. 13, 80638 München, Germany

17
18 10 ³Department of Life Sciences and Systems Biology, University of Turin, Via Accademia Albertina 13, 10123
19 Turin, Italy

20
21
22 12 ⁴Bayerische Akademie für Naturschutz und Landschaftsplanung (ANL), 83406 Laufen/Salzach, Germany

23
24 13 ⁵Ökologische Forschung und Planung, Geyer und Dolek, Alpenblick 12, 82237 Wörthsee, Germany

25
26
27
28 15 **Corresponding author:**

29
30 16 Dr. Tatjana Čelik; e-mail: tcelik@zrc-sazu.si

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

17 **Abstract**

1
2 18 The knowledge on species' habitat preferences at local scales across its range is an essential condition for
3
4 19 defining the most appropriate habitat management for the conservation of any species. In this study, we
5
6 20 combined field observations from three European countries with breeding experiments under field conditions to
7
8 21 identify oviposition and larval preferences of *Coenonympha oedippus* on micro-scale level across contrasting
9
10 22 habitat types (wet vs. dry). Despite the wide geographical range and the different habitats we found some
11
12 23 common features: (i) vegetation structure of the herb layer is an essential factor for oviposition site electivity and
13
14 24 successful development of premature stages; (ii) high cover of litter and/or dwarf shrubs in the microhabitat
15
16 25 (larval: 45–70% , oviposition: 40–50%) creates a herb layer rich in gaps; at their edges eggs are deposited and
17
18 26 the caterpillars are adequately sun-exposed; (iii) egg-laying females are not selective regarding oviposition
19
20 27 substratum; (iv) oviposition height is adjusted to positions with direct sunlight or warm substratum; (v) the host-
21
22 28 plants coverage in oviposition sites was high: 45 and 50% (wet), 18% or 41% (including potential host-plants)
23
24 29 (dry); (vi) the most important host-plant is *Carex panicea* (wet) and *Carex humilis* (dry), but *Molinia caerulea*
25
26 30 (wet) and *Festuca rupicola* (dry) are also used regularly; (vii) the availability of winter-green host-plants in the
27
28 31 vicinity of hibernated larvae plays a substantial role in their survival. As regular mowing or grazing would
29
30 32 remove the litter and destroy the gaps, the management should be restricted to selective reed cutting or manual
31
32 33 shrub removal. Only selective mowing during winter (December–February) can be recommended for keeping
33
34 34 the habitat open where the reduction of bushes is not sufficient.
35
36
37
38

39 36 **Key words:** False Ringlet, microhabitat choice, oviposition preferences, larval ecology, successional habitats
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

37 Introduction

38

39 For the successful conservation of any species, knowledge of its habitat preferences at local scales across a
40 species' range is required (e.g. Anthes et al. 2008), and an understanding of the perceptual world of the target
41 organism and its interactions with the environment from the functional habitat point of view is a benefit (Dennis
42 2003; Dennis et al. 2006, Van Dyck 2012). The resource-based habitat concept based on species-specific
43 resource distribution and individual movements considers all life stages of an organism and thus offers an
44 advanced approach for successful conservation management of threatened butterfly species.

45 The False Ringlet, *Coenonympha oedippus* (Fabricius, 1787), is one out of twelve endangered European
46 butterfly species (van Swaay et al. 2010), and listed on the Annexes II and IV of the Habitats Directive. Its
47 distribution occupies only 3.92% of the area of Europe (Kudrna et al. 2011). The Climatic Risk Atlas of
48 European Butterflies (Settele et al. 2008) placed the species in a category of "climate change risk (R)" due to
49 more than 50% loss of its current grid cells under at least one of the three simulated scenarios. Previous studies
50 represent the first important part of knowledge for better understanding the species' habitat use within the
51 functional resource-based habitat approach including both, consumables and utilities. They aimed on habitat
52 requirements (Bonelli et al. 2010, Bräu et al. 2010, Čelik & Verovnik 2010, Dušej et al. 2010, Örvössy et al.
53 2010, Sielezniew et al. 2010, Šašić 2010), adult movements and population ecology (Čelik 2003, 2004, Čelik et
54 al. 2009a, Čelik & Verovnik 2010, Örvössy et al. 2010, 2013), together with some initial researches on ecology
55 of early developmental stages (Bonelli et al. 2010, Bräu et al. 2010, Čelik 1997, 2003, Čelik & Verovnik 2010).

56 *C. oedippus* is a satyrine butterfly. Larvae of this subfamily feed on plants from the families Poaceae,
57 Cyperaceae and Juncaceae (Munguira et al. 2009). Such grass-feeding species are generally thought to be less
58 specific in their host-plant and oviposition-site choice (Wiklund 1984, Lindman et. 2013). Namely, the ability of
59 larvae to utilize superabundant (i.e., graminoids) and multiple plant species as host-plants reduces limitations of
60 females in oviposition-site selection (Gripenberg et al. 2010), but can also lead to a loss of benefits gained by
61 female discrimination, as are favourable microhabitats for the premature stages (Zalucki et al. 2002, Eilers et al.
62 2013, Lawson et al. 2014), higher nutritional quality of hosts (Awmack et al. 2002), and reduced parasitism or
63 predation or intra- and interspecific competition (Doak et al. 2006). Nevertheless, it has been discovered recently
64 that grass-feeding species have specific requirements concerning the quality of host plants and microhabitat
65 structures (e.g., Möllenbeck et al. 2009, Beyer & Schultz 2010, Weking et al. 2013) especially in terms of
66 vegetation height and density, amounts of grass-litter and sun exposure.

67 Across its European range *C. oedippus* occurs in two contrasting habitat types regarding soil humidity: the
1 majority of populations lives on semi-open wet grasslands (ordo Molinietalia, ordo Tofieldietalia) (Čelik 1997,
2 68 2003, 2004, Dierks 2006, Bonelli et al. 2010, Bräu et al. 2010, Čelik and Verovnik 2010, Dušej et al. 2010,
3 69 Örvössy et al. 2010, Sielezniew et al. 2010, Šašić 2010), but at the southern range limit also on dry habitats
4 70 (Ruehl 1895, Hafner 1910, Kolar 1919, 1929, Bischof 1968, Habeler 1972). However, populations occurring on
5 71 dry abandoned grasslands (class Festuco-Brometea) are presently known only from Slovenia (Čelik 2003, Čelik
6 72 and Verovnik 2010). Conservation of species using distinct habitat types most likely requires different
7 73 management strategies (e.g. Kalarus et al. 2013). The habitats of the existent and last strong European
8 74 populations of the False Ringlet are at least partially abandoned and if they are mown this happens only
9 75 infrequently/extensively (Bonelli et al. 2010, Bräu et al. 2010, Čelik and Verovnik 2010, Örvössy et al. 2010),
10 76 while regular mowing each summer within flight period leads to drastic declines and the extinction of
11 77 populations (Čelik et al. 2009b, own obs.).
12 78

13 79 As already stressed by Thomas (1993), warm early successional habitats are often crucial for oviposition and
14 80 larval development for Lepidoptera species that reach their northern limit in Central Europe. In contrast, *C.*
15 81 *oedippus* inhabits late successional stages. The species seems to avoid large open ranges and prefers clearing-
16 82 like habitats with interspersed bushes, enclosed by hedges (own obs., Örvössy et al. 2010). Such species of late
17 83 successional habitats have been less studied compared to those of early successional stages, resulting in a higher
18 84 demand of new information.
19 85

20 86 Previous studies on ecology of *C. oedippus* contributed to knowledge on habitat requirements and biology of
21 87 early developmental stages within a single European country and habitat type. In this study, we used field
22 88 observations and breeding experiments under field conditions to identify environmental parameters responsible
23 89 for larval and oviposition microhabitat choice in three habitat types differing in soil humidity (wet vs. dry) and
24 90 soil reaction (wet habitat: alkaline vs. acid). Then we compared three habitat types to determine common key
25 91 features leading to the selection of larval and oviposition sites in *C. oedippus* across a large spatial scale.
26 92

27 93 Therefore, we hypothesized that (1) *C. oedippus*, as a grass-feeding species, is not very selective concerning
28 94 oviposition substratum and host-plant species range within the contrasting habitat types; that (2) vegetation
29 95 structure in larval/egg-laying microhabitats should be a more important factor in microhabitat selection than
30 96 species composition of the vegetation; that (3) if differences in microhabitat structure between habitat types exist
31 97 the required structure of larval/egg-laying microhabitat linking with microclimatic conditions most suitable for
32 98
33 99
34 100
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

96 development of premature stages is achieved by adjusting micro-site selection across the macro-environmental
1 97 gradient, meaning that larval/egg-laying habitat plasticity exists in the study species.

3 98 Based on the results from field observations and breeding experiments the implications for habitat management
4 99 of the endangered *C. oedippus* are discussed.
5
6
7

8 100

9 101

12 102 **Materials and methods**

14 103

16 104 *Study species*

18 105 The False Ringlet is distributed from the Pyrenees in the west to Northeast-China, Korea and Japan in the east.

20 106 At present, it is still very widespread and abundant in SE Transbaikalia and Altai (Gorbunov & Kosterin 2007).

22 107 All populations east of the Ural Mountains are considered as different subspecies. In Europe, mostly isolated

24 108 populations are still present in France, Germany, Liechtenstein, Austria, Italy, Slovenia, Croatia, Hungary,

26 109 Poland, Belarus, Ukraine and Russia. Italy plays a central role in the conservation of *C. oedippus* in Europe as

28 110 the species is still abundant in the north of the country, where over 100 populations are known and only 9 were

30 111 documented to be extinct in recent times, mainly following habitat loss due to human activities (Balletto et al.

32 112 2007, 2014, Bonelli et al. 2010). In Slovenia, the species has a disjunct distribution, with two main centres:

34 113 central (wet habitats) and south-western (dry habitats) part of the country. In central Slovenia only one

36 114 metapopulation (ca. 2000 individuals) and 5 small neighbouring populations (each of them < 500 individuals)

38 115 have been known within the last 15 years. During this period, three of the smaller populations have become

40 116 extinct and total population size decreased by about 80%. Main reasons are regular mowing during or before

42 117 flight season and destruction of the habitat by conversion into arable land. In south-western Slovenia, the species

44 118 is more widespread but population densities are much lower than the density of the strongest wet subpopulation

46 119 estimated 15 years ago (Čelik, own obs.). In Germany only few sites (wet habitats) are documented historically.

48 120 They are all in the southern part of Bavaria mainly along the Isar river valley. At present, one of these sites still

50 121 harbours *C. oedippus*, consisting of three habitat patches in close vicinity to each other. Before its rediscovery in

52 122 1996, the species was thought to be extinct in Germany (Bräu & Schwibinger 2013).

54 123 *C. oedippus* inhabits late successional habitats, which are not dominated by woody plants. According to

56 124 Osthelder (cit. in Kolar 1929), *C. oedippus* may already have immigrated from its core territory in Asia to central

58 125 and western Europe during the last interstadial (Allerød) of the Weichselian glaciation about 10.000 years ago.

59
60
61
62
63
64
65

126 Hence, primary habitats of the species may have been places where woodland could not completely close, but
127 had larger open spaces because of very wet (e.g. in spring fens with continuously very wet soil or periodically
128 flooded areas along rivers and brooks) or dry conditions, possibly combined with grazing of wild animals. This
129 may be the reason why the species is also found, apart from wetlands, in dry abandoned grasslands of the karst
130 regions of northern Italy (only old data exist which are not recently confirmed) and south-western Slovenia, and
131 in folds of dry steppes on southern slopes with narrow strips of mesoxerophilous meadow vegetation and shrub-
132 lands in southern Siberia (Gorbunov & Kosterin 2007). Specific ecological needs of *C. oedippus* together with a
133 very low dispersal capacity (Čelik 2003, Örvössy et al. 2013) could explain the very scattered (presumably
134 relict) distribution of the species in its western range. Many of the remaining habitats have changed due to
135 human activities, such as lowering the water table for intensification of land use, peat ditching, and due to
136 abandonment of extensive grazing.

137 Adults fly in one generation from June to July. Females lay single eggs on different substrata. Larvae are
138 heliophilous and thermophilous: they are feeding during the day and basking in the sun in the warmest part of the
139 day in late autumn and early spring (own obs.). Many food plants (grasses and sedges) have been published (e.g.
140 Chretien 1886, Weidemann 1995, Lhonore 1996, Lhonore and Lagarde 1999, Lafranchis 2000, Tshikolovets
141 2003, Dierks 2006) but most of them refer to breeding experiments. Some of them are very likely to be used in
142 the field as well (e.g. *Deschampsia caespitosa* (L.) P. Beauv.), while others were not confirmed in recent
143 breeding experiments (e.g. *Schoenus nigricans* L., Dierks 2006, own obs.).

144

145 *Study sites*

146 Between 2008 and 2012, we studied larval and egg-laying habitats of *C. oedippus* in three countries (Germany –
147 DE, Italy – IT, Slovenia – SLO) that represent the main three habitat types occupied by this species in Europe:
148 wet grasslands on alkaline soil (DE), wet grasslands on acid soil (IT) and dry grasslands (SLO).

149 In DE, the only remaining population was studied, which inhabits a small and isolated area (ca. 1.1 ha, altitude
150 490 m) in open wetlands partially enclosed by hedges (Bräu & Schwibinger 2013). The wetter parts of the
151 habitat patches can be classified as *Schoenetum ferruginei* Du Rietz 1925, with interspersed clusters of
152 *Cladietum marisci* Allorge 1922. However, the *Schoenetum* is highly dominated by *Carex panicea* L. Outside
153 depressions created by peat ditching in former times, vegetation is transient to *Allio suaveolentis-Molinietum*
154 *caeruleae* Görs in Oberd. ex Oberd. 1983 merging with *Cirsio tuberosi-Molinietum arundinaceae* Oberd. et
155 Philippi ex Görs 1974 in drier parts. The habitat patches have been abandoned for decades. While in some parts

156 of the site bushes (mainly *Frangula alnus* Mill.) are highly abundant, they are rare in areas with high densities of
157 *C. oedippus*.

158 In IT one site within the "Baraggia" Regional Oriented Reserve was studied (WGS 84: 45°31'39.6" N 8°09'17.4"
159 E, altitude 300 m). The Reserve protects fragments of natural areas, surrounded by human-modified habitats,
160 mainly rice fields. It is characterised by scattered woods (*Quercus robur* L., *Betula pendula* Roth, *Carpinus*
161 *betulus* L.), interrupted by large clearings, which are dominated by *Calluna vulgaris* (L.) Hull and *Molinia*
162 *caerulea* (L.) Moench. The study site is composed of three isolated (1600–5500 m apart) patches (1.0, 1.2, 1.7
163 ha). All are *Molinia caerulea* meadows, partially covered by *Calluna vulgaris* and with a shrubby perimeter
164 (*Betula pendula*, *Populus tremula* L., *Frangula alnus*). One patch is regularly mown in its central part once per
165 year; the second is extensively grazed; the third is unmanaged, but regularly used by tourists for recreational
166 activities. All three patches are periodically used for military exercises.

167 In SLO one site on the Kras plateau (WGS 84: 45°52'12.17" N 13°36'27.32" E, altitude 260 m) was investigated.
168 It consists of three adjoining (350–1000 m apart) patches (1.7, 1.5, 0.8 ha) characterised by abandoned,
169 floristically poor, and overgrown submediterranean–illyrian dry grasslands (associations of *Danthonio-*
170 *Scorzoneretum villosae* Ht. & Ht-ić in Ht-ić 1963 and *Carici humilis-Centaureetum rupestris* Ht. & Ht-ić 1934
171 on deeper soils). It is characteristic for these successional stages that sedges and grasses dominate the vegetation,
172 while non-graminoid herbs are less abundant. Characteristic is also a high structural heterogeneity, which is
173 composed of dense swards of *Sesleria autumnalis* (Scop.) F. W. Schultz scattered over a predominant cover of
174 mainly lower and sparse herb layer consisting of other grasses and sedges. Individual islands of shrubs (*Cotinus*
175 *coggyria* Scop., *Prunus mahaleb* L., *Ligustrum vulgare* L.) and young thermophylic low trees (*Fraxinus ornus*
176 L., *Ailanthus altissima* (Mill.) Swingle, *Pinus nigra* Arnold) grow scattered over the entire abandoned grasslands
177 and surround the study patches.

178

179 *Larval microhabitat*

180 To obtain data on host-plants and habitat preferences of larvae after hibernation, field surveys were carried out
181 from April to May (DE: 2008, 2009; IT: 2009, SLO: 2010, 2011). For each detected larva we recorded the host-
182 plant species (if feeding was observed) or plant species with feeding traces near the caterpillar, and the percent
183 cover of main structural parameters within a radius of 50 cm around the larva: bare ground, rocks, mosses, litter,
184 shrubs, herbs and known host-plants. Additional structural parameters were assessed for each country: percent
185 cover of grass-like herbs (GLH, i.e. herbs with plant structure, that do not shade the lower parts of the herb layer;

186 e.g., *Allium* spp., *Anthericum ramosum* L., *Genista sylvestris* Scop.; DE, SLO), percent cover of plants from the
187 families *Poaceae*, *Juncaceae*, *Cyperaceae* (PJC, i.e. plants with erectophile leaf orientation; DE, SLO), percent
188 cover of *Calluna vulgaris* (IT), percent cover of each species from the group PJC (SLO), average vegetation
189 height (i.e. prevailing height of herb layer in cm; SLO) and maximum vegetation height (SLO), the lattermost
190 being the tallest plant in microhabitat. Those plants corresponded mainly to *Chrysopogon gryllus* (L.) Trin.,
191 *Bromopsis erecta* (Huds.) Fourr. s.str., *B. condensata* (Hack.) Holub, *B. transsilvanica* (Steud.) Holub and *Stipa*
192 sp.; they deviated considerably from average vegetation height. In Germany and Italy, the percent cover of main
193 and additional structural parameters were also recorded for random microlocations (May 2009) which were
194 selected by a randomly thrown stick (Anthes et al. 2003), and represented the spectrum of available
195 microlocations within the studied habitat type. In total, we recorded 49 random microhabitats (DE: 39, IT: 10)
196 and 76 larval microhabitats (DE: 31, IT: 34, SLO: 11).

198 *Oviposition microhabitat*

199 We tracked egg-laying females on sunny days from June to July (DE: 2008, 2010 and some additional
200 observations in 2011 and 2012; IT: 2009; SLO: 2010, 2011). Each female was chosen randomly and then
201 followed for a maximum of 10 ovipositions. If no egg-laying occurred within 20 minutes, we selected another
202 female. Females followed for multiple ovipositions clearly switched plant species during consecutive
203 ovipositions, meaning that repeated sampling of the same female did not represent pseudo-replication and not
204 bias the result on oviposition electivity. The same main and additional structural parameters were recorded for
205 oviposition (DE: 76, IT: 101, SLO: 55) and random (DE: 35, IT: 150, SLO: 30) microlocations (DE, SLO: July
206 2010; IT: June 2009) as for the larval microhabitats (see above). Random microlocations again reflect the
207 spectrum of available structures of the site. Additionally, we collected data on the oviposition substratum: plant
208 species, plant part (leaf, stem, bud, other), support freshness (vital, dead), and oviposition height above ground.

210 *Breeding experiments*

211 Breeding of five generations of *C. oedippus* under field conditions (ex-situ) was performed in Germany from
212 2009 to 2014. Several vivaria of different size (minimum 20 x 30 x 30 cm) were planted with sods taken from
213 the habitat, containing mainly *Molinia caerulea* and *Carex panicea*. The top of the vivaria was covered with
214 gauze which was fixed with an elastic rubber band, allowing high air circulation. The vivaria were always kept
215 outdoors, in the same region where the butterfly population lives. They received direct sunlight for several hours

216 a day and were sheltered from rain and snow. Humidity was held at a high level by regular watering.
1
2 217 Temperature and humidity were measured with a digital thermo-hygrometer over a long period under different
3
4 218 weather conditions. Both were similar to the values occasionally measured at the species' habitat. For
5
6 219 oviposition within the vivaria, some female butterflies were taken from the field (after observation of oviposition
7
8 220 in the field, to ensure that they already started ovipositing) and some that emerged and mated in captivity could
9
10 221 be used for maintaining the breeding stock as well. Depending on the year, between 26 and 92 pupae were
11
12 222 achieved.

13
14 223 In Slovenia, only one young caterpillar was transferred from the field (i.e. study site) in October 2011 to a glass
15
16 224 vivarium (50 x 50 x 50 cm) covered with gauze and planted with the sod taken from the location of the
17
18 225 caterpillar. The vivarium was exposed to outdoor temperature and sunlight but sheltered from rain and snow. Its
19
20 226 inside was humidified with water spraying through the mesh every time it rained or snowed out. The caterpillar
21
22 227 was bred until the butterfly hatched (in June 2012).

23
24 228 In both experiments, the behaviour of caterpillars was observed and noted in short intervals.

25
26 229

27
28 230 *Statistical analyses*

29
30 231 Larval/oviposition microhabitat electivity within each habitat type was evaluated by comparison between
31
32 232 larval/oviposition and random microlocations using multiple stepwise forward logistic regression. We fitted a
33
34 233 presence-absence logistic model to our presence-only data (i.e., random microlocations represent pseudo-
35
36 234 absences, cf. Ward et al. 2009) as we anticipate that the probability of selecting the random microhabitat with the
37
38 235 presence of eggs/larvae must be very low given the high density of oviposition substrata (considering non-
39
40 236 selectivity of egg-laying females) and host-plants (considering host-plants growth form and larval polyphagy)
41
42 237 compared with the density of *C. oedippus* in each study site. Further, in the case that some of selected random
43
44 238 microlocations contained the eggs/larvae, the difference between oviposition/larval and random microlocations
45
46 239 were even underestimated (cf. Eilers et al. 2013), meaning that our estimations are conservative. Before
47
48 240 regression analysis, all explanatory variables (i.e. structural parameters of microlocations) were tested for
49
50 241 intercorrelations by calculating Spearman's Rho correlation coefficients. We defined two types of explanatory
51
52 242 variables, i.e. "basic" and "derived" (see Table 2 and 4). Basic variables are main structural parameters of the
53
54 243 microhabitat, while derived variables are structural parameters which represent only a part of the coverage of
55
56 244 corresponding basic variable. Within both types, we distinguished between "composed" and "simple" variables.
57
58 245 Composed variables can be substituted by more simple variables, e.g. Herbs can be substituted by HPs and
59
60
61
62
63
64
65

246 Herbs_without HPs, or PJC, GLH and NGLH. Simple variables can not be replaced by a set of other variables.
1
2 247 In the case of strong correlation (Spearman $Rho \geq |0.9|$) between two variables within pairs “basic vs. basic” or
3
4 248 “basic vs. derived”, the simple variable was maintained for entering in regression analysis. If two derived
5
6 249 variables were strongly correlated, the one which correlated strongly with selected basic variables was excluded
7
8 250 from further regression analysis. The last criterium was considered also in the case of strong correlation between
9
10 251 two simple variables.

11 252 For identifying the differences between three habitat types for each structural parameter separately, the Kruskal-
12
13 253 Wallis Chi test (KW) and Mann-Whitney U tests (MWU) were applied on all possible comparisons as post hoc
14
15 254 procedures with Bonferroni correction.

16
17
18 255 To find out whether the difference in oviposition height between three habitat types depends on height of
19
20 256 prevailing egg-laying support, the Jonckheere–Terpstra test was applied. Four categories of egg-laying supports
21
22 257 were coded according to plant height (Lauber & Wagner 1996): 1 = *Carex humilis* (3–11 cm), 2 = *Calluna*
23
24 258 *vulgaris* (10–50 cm), 3 = *Carex panicea* (20–40 cm), 4 = *Molinia caerulea* agg. (30–250 cm).

25
26 259 To detect the relationships between the type of oviposition support (dependent variable) and most abundant
27
28 260 structural parameter (independent variable) in egg-laying microhabitats, Chi-Square tests using Likelihood ratio
29
30 261 statistic were applied because of small sample sizes, which resulted in expected frequencies lower than 1 in some
31
32 262 cases. For assessing the strength of association between both variables, Cramer's *V* was used. Standardized
33
34 263 residuals were used to define the significant contributors to the overall chi-square value. For the purpose of Chi
35
36 264 square testing, the egg-laying supports and most abundant structural parameters were arranged in the following
37
38 265 categories: DE – *Molinia caerulea*, *Carex panicea*, other herbs, litter, shrub; IT – *M. caerulea*, *Carex* sp.,
39
40 266 *Calluna vulgaris*, other herbs; SLO – *Carex humilis*, *Poaceae*, other plants. As litter was the most abundant
41
42 267 structural parameter in 76% of the Slovenian egg-laying microhabitats, the possible relationships between the
43
44 268 type of oviposition support and the abundance of structural parameters was analyzed using the second most
45
46 269 abundant structure parameter in the microhabitats.

47
48 270 All statistical analyses were performed using SPSS 13.0 (SPSS Inc. 1989–2004).

49
50 271

51
52 272

53 54 273 **Results**

55
56 274

57
58 275 *Larval preferences*
59
60
61
62
63
64
65

276 In our three habitat types, a total of 85 feeding observations were detected on 6 plant substrata: *Carex panicea*,
277 *C. davalliana* Sm., *C. humilis*, *Carex* sp., *Molinia caerulea* and *Festuca rupicola* (Table 1). Sedges represented
278 the majority of the larval diet after winter in all three habitat types (DE: 71%, IT: 94%, SLO: 77%), while
279 observed alternative host-plants were *M. caerulea* in both wet grassland types and *F. rupicola* in dry habitat.
280 Larval microhabitats differed between three habitat types not only in the presence of main structural parameters
281 (Table 2), but also in the coverage of those they had in common: litter (KW $\chi^2 = 22.67$, df=2, $p = 0.000$), herbs
282 (KW $\chi^2 = 11.94$, df=2, $p = 0.003$) and shrubs (KW $\chi^2 = 60.34$, df=2, $p = 0.000$) (Fig. 1a). Wet larval
283 microhabitats were characterized by a higher coverage of litter and by lower abundances of herbs than dry
284 microhabitats. Shrub cover was significantly higher in Italian microhabitats than in the other two countries. In
285 Italy, the majority of shrub cover was formed by *Calluna vulgaris* (Table 2). If we presumed that this perennial
286 dwarf shrub has a similar function as litter for overwintering larvae and add its coverage to litter, the larval
287 microhabitats still significantly differed between Italy and Slovenia, but not between both wet habitat types (Fig.
288 1a). The cover of host-plants (Table 2) was significantly higher in German (median = 30%) than in Slovenian
289 (median = 12%) larval microhabitats (DE vs. SLO, MWU $Z = -4.23$, $p = 0.000$; respective data not available for
290 IT). If we assume that the other grass species present in dry habitats are also used as larval food-plants (PJC in
291 Table 2), the host-plants abundances differ only slightly between both habitat types (MWU $Z = -2.15$, $p =$
292 0.032).
293 In Germany, occupied microhabitats had a significantly higher coverage of litter and of the preferred host-plant
294 *C. panicea*, and lower abundances of shrubs, *M. caerulea* and non grassy-like herbs than available sites (Table
295 2). Due to lower coverage of *M. caerulea* and herbs with planophile leaf orientation, larval microhabitats were
296 characterized by lower abundances of host-plants, graminoids (PJC), all herbs and herbs other than host-plants.
297 Larval preferences for microlocations with high coverage of *C. panicea* revealed that this preferred host-plant
298 has additional characteristics, which enable better survival of caterpillars after hibernation compared to the
299 alternative host *M. caerulea* (see section *Ex-situ breeding observations*). In Italy, preferences of overwintering
300 larvae showed a similar pattern as in Germany: they preferentially occurred on sites with higher coverage of
301 litter or litter+*Calluna vulgaris* and lower abundance of herbs compared to the available microlocations (Table
302 2). Besides, larval microhabitats had lower coverage of bare ground. No information was available to assess the
303 effect of host-plants on larval microhabitat selection in Italy. However, in both wet habitat types, the coverage of
304 litter already explained most of the variation in microhabitat selection by overwintering larvae (Table 3). The
305 likelihood of a microlocation being occupied by larvae after hibernation increased with litter coverage.

307 *Preferences of egg-laying females*

308 During this study we found out that females apply two oviposition modes in the pre-alighting phase of
309 oviposition site selection, (i) “normal mode” – female lands on plants parts in the upper layer of the herb
310 vegetation and remains there for post-alighting phase (egg-laying), and (ii) “dropdown behaviour” – female
311 lands on plants parts in the upper layer of the herb vegetation, then crawls or drops down to the ground, and
312 walks on the ground searching for suitable egg-laying support. The second mode is less frequently used and it
313 was observed only in Germany.

314 The eggs were deposited on plants growing at the edge of gaps in the herb layer. In three habitat types, a total of
315 236 eggs were laid on 27 different egg-laying supports (Table 1), including herbs (71%), shrubs (19%), litter
316 (2%) and undetermined plants (8%). All eggs were deposited singly on leaves (89%), stems (10%) or buds (1%).
317 Three quarters (76%) of the eggs were attached on vital parts of plants, the remaining on dry parts or litter (20%)
318 and undetermined substratum (4%).

319 The above described wide range indicates a rather unselective egg-laying behaviour. Nevertheless, some plant
320 species dominate the plant spectrum used for oviposition. The prevailing supports were *Molinia caerulea* (55%)
321 and *Carex panicea* (28%) in Germany ($\chi^2 = 40.35$, d.f. = 1, $p < 0.001$), *Calluna vulgaris* (41%) and *M. caerulea*
322 (34%) in Italy ($\chi^2 = 34.21$, d.f. = 1, $p < 0.001$), and *Carex humilis* (65%) in Slovenia ($\chi^2 = 5.27$, d.f. = 1, $p <$
323 0.05). The plant species used for oviposition and the most abundant structure parameter in the egg-laying
324 microhabitat show a significant positive association (DE, LR = 40.90, df = 16, $p = 0.001$; Cramer's V = 0.560, p
325 = 0.000; IT, LR = 43.47, df = 15, $p = 0.000$; Cramer's V = 0.406, $p = 0.000$; SLO, LR = 13.03, df = 4, $p = 0.020$;
326 Cramer's V = 0.306, $p = 0.032$). Eggs are simply deposited on the most abundant plant species, which explains
327 the regular use of the non-host *C. vulgaris* in Italy.

328 The vertical positions of eggs on substrata ranged from 0 to 44 cm and oviposition height increased with the
329 height of prevailing egg-laying support (Jonckheere–Terpstra $z = 5.12$, No. of levels = 4, $p = 0.000$). The eggs
330 were deposited significantly higher in the most often applied (“normal”) oviposition mode (N = 67) than in
331 dropdown mode (N = 9) (MWU $Z = -4.37$, $p = 0.000$).

332 Oviposition sites differed between three habitat types not only in vegetation composition (due to different
333 vegetation grassland communities) but also in microhabitat structure considering the presence and abundances of
334 main structural parameters (Table 4). German and Slovenian microhabitats were denoted by a higher cover of
335 litter and a lower cover of shrubs than Italian ones (KW, litter: $\chi^2 = 38.36$, df=2, $p = 0.000$; shrubs: $\chi^2 = 81.15$,

336 df=2, $p = 0.000$; Fig. 1b). The coverage of herbs other than host-plants was significantly higher in dry than in
1
2 337 wet microhabitats (KW $\chi^2 = 146.09$, df=2, $p = 0.000$; Fig. 1b). Consequently, host-plants were more abundant in
3
4 338 wet than in dry oviposition spots (KW $\chi^2 = 65.39$, df=2, $p = 0.000$; Fig. 1b). But in dry habitats several further
5
6 339 grass species occurred (PJC in Table 4); if they are taken into consideration as potential host-plants, the
7
8 340 difference is equalled out (KW $\chi^2 = 5.91$, df=2, $p = 0.052$). Litter is an important structure in oviposition
9
10 341 microhabitats, because it creates a vegetation structure (herb layer) rich in gaps. Eggs are usually deposited at the
11
12 342 edges of these gaps. As it was one of the most variable structural parameter in Italian oviposition microhabitats,
13
14 343 and its median cover was much lower than in the other two countries (Fig. 1b), we presumed that other structures
15
16 344 take over its role when its cover is very low, e.g. perennial *C. vulgaris* according to its significant negative
17
18 345 correlation with litter cover (Spearman rho = -0.399 , $p = 0.000$). Indeed, the sum of litter and *C. vulgaris* in
19
20 346 Italian oviposition microhabitats was only slightly higher than litter cover in Germany and Slovenia (Fig. 1b).
21
22 347 In all three habitat types, the shrub cover was significantly lower in oviposition than in available sites (Table 4).
23
24 348 In Germany, this was the only difference between both types of microlocations. Italian egg-laying microhabitats
25
26 349 had significantly higher abundances of litter and *M. caerulea* than available sites. In Slovenia, oviposition
27
28 350 microhabitats were characterized by a higher coverage of litter and maximum vegetation height than were in
29
30 351 available microlocations, and also by lower abundances of all herbs and herbs other than observed host-plants.
31
32 352 Thus, German oviposition microhabitats were the standard structure of the locality inhabited by *C. oedippus* as
33
34 353 the difference in shrub cover between oviposition and available sites was not enough to improve a constant-only
35
36 354 model. In both other countries, oviposition pattern/site selection was best explained by a combination of litter
37
38 355 cover and abundances of shrubs and non-host plants. In Italy, the likelihood of a spot being accepted for
39
40 356 oviposition increased with the litter coverage and decreased with the cover of shrubs other than *Calluna vulgaris*
41
42 357 and herbs other than host-plants (Table 5). In Slovenia, the presence of high plant stems of *Chrysopogon gryllus*,
43
44 358 *Bromopsis* sp., *Stipa* sp. which deviated from prevailing/average vegetation height positively influenced the
45
46 359 selection of oviposition microhabitat, but high covers of shrubs and other herbs than host plants decreased the
47
48 360 likelihood of a spot to be chosen by an egg-laying female (Table 5). Presence of high grass species indicates
49
50 361 microlocations with higher coverage of *C. humilis* (Spearman rho, *C.h.* vs. *Bromopsis* sp. = 0.268, $p = 0.014$,
51
52 362 *C.h.* vs. *C. gryllus* = 0.249, $p = 0.022$, *C.h.* vs. *Stipa* sp. = 0.269, $p = 0.013$) and litter (Spearman rho, *Bromopsis*
53
54 363 sp. vs. litter = 0.233, $p = 0.041$), and lower abundance of *Sesleria autumnalis* (Spearman rho, *S. a.* vs. *C. humilis*
55
56 364 = -0.363 , $p = 0.001$; *S. a.* vs. *Bromopsis* sp. = -0.368 , $p = 0.001$; *S. a.* vs. *C. gryllus* = -0.386 , $p = 0.000$; *S. a.* vs.
57
58 365 *Stipa* sp. = -0.282 , $p = 0.009$) which creates a very homogeneous and dense sward. Furthermore, coverage of
59
60
61
62
63
64
65

366 herbs with planophile leaf orientation was lower in occupied than in available microlocations although the
1 difference was only close to significance (Table 4: Herbs without PJC).
2
3
4 368 Oviposition microhabitats were characterised by a clearly lower proportion of litter and a higher cover of herbs
5
6 369 than larval spots (Table 2 and 4; MWU: litter, DE (40 vs. 70%): $Z = -6.71$, $p = 0.000$; IT (10 vs. 45%): $Z = -$
7
8 370 5.01 , $p = 0.000$; SLO (40 vs. 45%): $Z = -2.70$, $p = 0.007$; herbs, DE (60 vs. 30%): $Z = -6.43$, $p = 0.000$; IT (50
9
10 371 vs. 30%): $Z = -2.70$, $p = 0.007$; SLO (56 vs. 50%): $Z = -0.98$, $p = 0.327$), which is largely a seasonal effect.

11 372

14 373 *Ex-situ breeding observations*

16 374 Our observations from breeding under field conditions showed that caterpillars concentrate on the sun-facing
17
18 375 side of tufts while feeding or resting, except during search for a place to pupate and during hibernation. From
19
20 376 late October onwards, most caterpillars begin to retreat for hibernation: some overwinter at the base of sedges or
21
22 377 grasses or even on upper parts of their food-plants hidden from the sun, presumably to avoid awakening too early
23
24 378 on sunny winter days.

26 379 During breeding experiments we observed that caterpillars which hibernate in their third instar usually awake in
27
28 380 spring when temperatures rise to about 20°C for several consecutive days. At that time not all host-plants are
29
30 381 available. In Germany, the time lag between larval awakening and the start of *M. caerulea* growth varied from
31
32 382 about 10 to 30 days (Table 6). In contrast, *C. panicea* which remains green during winter is always available for
33
34 383 larvae as food resource. Hibernated caterpillars were observed feeding on it shortly after the start of activity.
35
36 384 Moreover, caterpillars do not feed on *Molinia* leaves shorter than about 5 cm, thus even enlarging the time lag.
37
38 385 In Slovenia, comparing awaking time with observations of host-plants growth status in the field and in ex-situ
39
40 386 showed (Table 6) that besides known host-plants, *C. humilis* and *F. rupicola* which overwintered green, some
41
42 387 other grass species, e.g. *Brachypodium rupestre*, *Sesleria autumnalis*, are also available.

44 388

46 389

49 390 **Discussion**

51 391

54 392 *Habitat requirements of ovipositing females and larvae*

56 393 Egg-laying females of *C. oedippus* are not selective regarding oviposition substratum, e.g. plant species or exact
57
58 394 position on the plant. Eggs were mainly deposited on the most abundant structure parameter (plant/plant group)

59
60
61
62
63
64
65

395 available in the microhabitat. This wide range is also reflected by previously known egg-laying plants from the
1 species' European range (Table 7).
2

3
4 397 Such absence of electivity is in contrast to many butterfly species, who's females carefully choose the
5
6 398 oviposition plant, e.g. *Phengaris (Maculinea)* butterflies (Dolek et al. 1998, Thomas & Elmes 2001, Kassai &
7
8 399 Peregovits 2005), *Lycaena alciphron* (Dolek & Geyer 2001), *Boloria aquilonaris* (Turlure et al. 2013),
9
10 400 *Euphydryas desfontainii* (Pennekamp et al. 2013), *E. maturna* (Dolek et al. 2013), *Colias myrmidone* (Dolek et
11
12 401 al. 2005, Szentirmai et al. 2014). This contrast very likely relates to the wide host-plant range of *C. oedippus*
13
14 402 larvae (i.e. different species from Poaceae and Cyperaceae), the growth form of host-plants (i.e. dense ground-
15
16 403 covering plants growing mostly in tufts) and their relatively high stability in terms of growing period and
17
18 404 abundance (e.g. due to possibility of vegetative reproduction).

19
20 405 Nevertheless, an important factor for suitable reproduction habitats is the availability of host-plants in close
21
22 406 vicinity to the oviposition substratum, i.e within reach of young caterpillars. The host-plants coverage in
23
24 407 oviposition sites was always high: between 45 and 50% in wet, and 18% (only observed host-plant species) or
25
26 408 41% (including potential host-plant species) in dry habitats. The finding that coverage of bare ground was
27
28 409 highest in Italian microlocations, and only there larval microhabitats had significantly lower bare ground cover
29
30 410 than random sites indicates that microlocations with higher amounts of bare ground are less suitable for
31
32 411 overwintering larvae, possibly because bare ground can restrict larval movements and expose them to predation
33
34 412 (e.g., Doak 2000). Alternatively, it could simply reduce the host-plants coverage and its connectivity around the
35
36 413 caterpillar what could increase the time of larval searching for a new suitable host. We indentified six food-plant
37
38 414 species in the field. Feeding on *M. caerulea* in the field was also observed in France (Dierks 2006) and Poland
39
40 415 (Sielezniew et al. 2010). In former publications many additional grass and sedge species have been listed, but
41
42 416 probably all of these refer to breeding experiments. Nevertheless, we are convinced that many more grass and
43
44 417 sedge species are used, if they are present in the habitat.

45
46 418 Non-specificity for oviposition substratum, together with a high proportion of eggs (DE: 55%, IT: 34%)
47
48 419 deposited on less frequently used host-plants (*M. caerulea*, DE: 27%, IT: 6%) or even on non-host plant material
49
50 420 (Table 1) points out to other crucial factors observed by egg-laying females to maximize offspring performance
51
52 421 (e.g., Mayhew 1997, Janz 2002). The models which best explain the oviposition and larval preferences of *C.*
53
54 422 *oedippus* showed that vegetation structure of microlocation is an important parameter in oviposition site
55
56 423 selection. High cover of litter and/or dwarf shrubs as *C. vulgaris* in selected microhabitats (median, larval: 45–
57
58 424 70%, oviposition: 40–50%) creates a herb layer rich in gaps. Egg deposition mostly occurred at the edges of
59
60
61
62
63
64
65

425 these gaps in order to allow helio- and thermophilous larvae to be adequately sun-exposed. Dry plant biomass
1
2 426 provides warmer environments (WallisDeVries & van Swaay 2006) which is of special importance for
3
4 427 overwintering larvae in early spring to enable them to reach optimal body temperature, and may also function as
5
6 428 a microclimatic buffer (e.g., Turlure et al. 2010, Weking et al. 2013). In the field, most caterpillars could be
7
8 429 found on sun exposed edges of tufts what is in harmony with the observations from ex-situ breeding. The lower
9
10 430 abundances of shrubs in egg-laying microlocations than in random sites in all three habitat types also point to the
11
12 431 importance of vegetation architecture with high solar insolation and low shading in oviposition/larval
13
14 432 microhabitat selection.
15
16 433 Structure of host-plant/oviposition substratum and surrounding vegetation directly influences the microclimate,
17
18 434 e.g. humidity, temperature and solar exposure (Beyer & Schultz 2010, O'Connor et al. 2014). Considering that
19
20 435 there was a positive association between the type of oviposition support and the most abundant structural
21
22 436 parameter in egg-laying microhabitat in all three habitat types, the height of prevailing egg-laying support could
23
24 437 be used as an indicator for average vegetation height in egg-laying microhabitat. Thus, positive correlation
25
26 438 between oviposition height and height of prevailing egg-laying support across three habitat types (from min to
27
28 439 max height: *C. humilis* (SLO)–*C. vulgaris* (IT)–*C. panicea* (IT, DE)–*M. caerulea* (IT, DE)) suggests that female
29
30 440 try to oviposit as high as possible on the selected substratum and within the vegetation (cf. Obermaier et al.
31
32 441 2006), meaning that eggs are deposited at positions that receive high solar radiation. The adjusting of oviposition
33
34 442 height to the height of the local radiation surface to maximise heat absorption was also observed in two other
35
36 443 satyrinae species which attach the eggs to a substratum (not simply drop them between the grasses), *Hipparchia*
37
38 444 *fagi* (Möllenbeck et al. 2009) and *Coenonympha tullia* (Weking et al. 2013). However, in normal oviposition
39
40 445 mode, females of *C. oedippus* laid eggs just below the top of the surrounding herb vegetation, but despite a
41
42 446 limited number of field observations, it seems that dropdown mode occurs when air temperature unexpectedly
43
44 447 decreases due to clouds temporarily covering the sun. In such cases, egg-laying females climbed down and
45
46 448 deposited eggs on the surface of the litter cover, probably as dry plant biomass provides warmer environment
47
48 449 than green plants (WallisDeVries & van Swaay 2006). Furthermore, the lower coverage of herbs with planophile
49
50 450 leaf orientation in oviposition (SLO) and larval (DE) than in random sites indicates that shading of the lower
51
52 451 herb layer is not favourable for the development of eggs and caterpillars. It seems that oviposition site selection
53
54 452 in *C. oedippus* is influenced by the thermal requirements of eggs and larvae. This is additionally supported by
55
56 453 the female's avoidance of dense tufts of *Sesleria autumnalis* in dry habitats, which do not offer adequate sun-
57
58 454 exposition. These results are also in correspondence with a previous study on within-patch movements of *C.*
59
60
61
62
63
64
65

1
2 456 *oedippus* adults in Slovenian wet habitat (Čelik et al. 2009a) which showed that spatial and temporal patterns of
3
4 457 female micro-distribution is affected by vegetation height, the homogeneity of host plant stands and the shading
5
6 458 of the the lowest parts of the herb layer.

7
8 459 Preferences of overwintering larvae and egg-laying females for microlocations with high amounts of host-plants
9
10 460 and litter and/or dwarf shrubs, and low amounts of bare ground, shrubs and herbs other than graminoids, together
11
12 461 with adjusting the oviposition height as high as possible within the vegetation across all three habitat types
13
14 462 revealed the high importance of vegetation structure in *C. oedippus* larval/oviposition microhabitat selection. At
15
16 463 first sight, this is a relatively general pattern of habitat use across the macro-environmental gradient. The
17
18 464 utilization of host-plants specific to the habitat type, and differences of preferred microlocations in presence and
19
20 465 relative abundances of structural parameters between the habitat types showed that such patterns are a result of
21
22 466 microhabitat selection conformed to local environmental conditions.

23 24 467 *Winter green food-plants as key factor for larval survival*

25
26 468 The butterfly host-plant synchronisation is a known phenomenon (e.g. review in Munguira et al. 2009), which
27
28 469 directly influences larval growth and survival, and ultimately population fitness. A perfect synchronisation is of
29
30 470 crucial importance for monophagous and oligophagous species overwintering as egg (e.g. de Vries et al. 2011) or
31
32 471 as young caterpillar, as with *C. oedippus* (e.g. Gradl 1946). Based on breeding experiments Gradl (1946) already
33
34 472 pointed out that a temporal mismatch exists between larval awakening and availability of the “prime” host-plant,
35
36 473 *Molinia caerulea*. He reported that caterpillars awoke from hibernation very early due to enduring foehn weather
37
38 474 in spring on March 20th, while their "normal" food resource *Molinia* was not yet available. Our breeding data and
39
40 475 field observations after hibernation showed that caterpillars of *C. oedippus* awake from hibernation at a time
41
42 476 period when only some host-plants are available (Table 6). At that time the larvae are still quite small (about 1.3
43
44 477 cm) and not able to move very far to search for food. Thus, they have to find host-plants in their immediate
45
46 478 vicinity within a few days as their need for food and liquid is urgent after the long period of starvation during
47
48 479 hibernation. Significant differences between larval and random spots in cover of the winter-green *C. panicea*
49
50 480 support the assumption that the likelihood of larval survival is strongly influenced by the availability of this plant
51
52 481 in German wetland habitats. *Molinia* starts to grow relatively late in spring, so in *Molinia*-dominated meadows
53
54 482 *C. panicea* or other winter-green *Carex* species with soft leaves are needed as interim food. For *C. oedippus*
55
56 483 habitats near Bordeaux, Dierks (2006) discussed the role of *Pseudoarrhenaterum longifolium* (Thore) Rouy
57
58 484 which starts to grow earlier than *Molinia* and could serve as food for early awakening caterpillars. It seems that
59
60
61
62
63
64
65

485 synchronisation between host-plants growing and larval diet requirements after hibernation is better harmonized
1
2 486 in dry habitats of *C. oedippus* as the preferred food plant *C. humilis* overwinters green and sprouts already in
3
4 487 early spring (March).

5
6 488 We hypothesize that winter-green host-plants play a substantial role also in the survival of overwintering
7
8 489 caterpillars of several other butterfly species occurring in habitats with highly dominating non-winter green host-
9
10 490 plants; a potential relationship that is so far not adequately taken into account. Namely, field observations on two
11
12 491 satyrinae species, *Coenonympha hero* (Wagner 2010) and *Minois dryas* (Sachteleben & Winterholler 2013),
13
14 492 living in *Molinia*-dominated habitats also showed that winter-green grasses (e.g. *Festuca* spp.) or sedges (*Carex*
15
16 493 spp.) were used as interim food source immediately after larval awakening.

17
18 494

19 20 495 *Late successional habitats*

21
22 496 It is well known and demonstrated in detail for many butterfly species that their caterpillars need early
23
24 497 successional stages, short turf, or otherwise hot and open habitats for their development (e.g. *Pyrgus malvae*
25
26 498 Krämer et al. 2012; *Hesperia comma* Hermann & Steiner 1997; *Parnassius apollo* Geyer & Dolek 1995;
27
28 499 *Lycaena alciphron* Dolek & Geyer 2001; *Scolitantides baton* Konvička et al. 2008; *Phengaris arion* Thomas
29
30 500 1980, Pauler et al. 1995, Fartmann 2005; *Polyommatus bellargus* Thomas 1983; *Chazara briseis* Königsdorfer
31
32 501 1997, Leopold 2001). The present study shows a contrasting habitat choice of *C. oedippus*, a species being
33
34 502 restricted to largely unmanaged grassland with a dominating litter layer, but no substantial growth of woody
35
36 503 plants. Our results on larval and oviposition preferences are in accordance with the findings from a population
37
38 504 study of *C. oedippus* in Hungary (Örvössy et al. 2013) which reveals that large amounts of grass litter and
39
40 505 structured vegetation with tussocks positively affect population size and density. Importance of late successional
41
42 506 habitats, mainly characterised by pronounced litter layer, relatively nutrient-poor conditions, no considerable
43
44 507 growth of shrubs and direct sun exposition, were also demonstrated for some other butterflies, e.g. *Coenonympha*
45
46 508 *hero* (Steiner & Hermann 1999, Dolek 2011, Bräu & Dolek 2013), *Lopinga achine* (Geyer & Dolek 2013) and
47
48 509 *Lycaena helle* (Nunner 1995).

49
50 510

51 52 511 *Implications for conservation*

53
54 512 The results of the present study showed that besides the availability of fresh-green host-plants in the vicinity of
55
56 513 hibernated larvae, mainly the vegetation structure and microclimate of the herb layer are essential factors for
57
58 514 oviposition site selection and successful development of premature stages of *C. oedippus*. The herb layer has to
59
60
61
62
63
64
65

515 be rich in gaps (but not bare ground), usually created by large amount of litter or alternatively by dwarf shrubs.
1
2 516 Such vegetation structure enables egg-laying females to adapt the oviposition height to local radiation surface
3
4 517 with high heat absorption. In a gap-rich herb vegetation, the ectotherm and heliophilous caterpillar dependent on
5
6 518 direct sunlight (i.e. when basking for elevation of body temperature) can select thermally favourable
7
8 519 microclimates by behavioural thermoregulation (c.f. Stevenson 1985, Turlure et al. 2011), i.e. translocation of its
9
10 520 position between the top of herbaceous plants, litter surface (warmer than upper green vegetation during
11
12 521 colder/cloudy days) and the more balanced microclimate inside the litter in terms of temperature and humidity. A
13
14 522 dense litter layer can also slow down the further secondary succession in the habitat as it prevents germination of
15
16 523 groves (Ellenberg 2009, Ruprecht & Szabó 2012), and consequently affects the structure, diversity and dynamics
17
18 524 of grassland plant communities (Ruprecht et al. 2010, Loydi et al. 2013). As litter in dry grasslands persists for
19
20 525 longer periods than litter from wet areas due to higher lignin concentration (Fortunel et al. 2009), the site
21
22 526 management in *C. oedippus* habitats must be diversified and adapted to meet the special needs of this highly
23
24 527 endangered species across its range. Hence, in most of *C. oedippus* habitats some kind of management has
25
26 528 become essential to avoid overgrowth with bushes and trees or, in wet habitats, with reed. Manual removal of
27
28 529 groves should be preferred. Regular mowing or forms of grazing which create a uniform vegetation structure (i.e.
29
30 530 close homogeneous sward) are thought to be detrimental. Furthermore, direct losses of larvae can be caused by
31
32 531 cutting, as caterpillars partially feed on their food-plants until the onset of November and some keep sitting on
33
34 532 upper parts of their food-plants even in winter (observations from breeding). In habitats where the reduction of
35
36 533 bushes is not sufficient, only patchy mowing in winter period (December–February) can be recommended for
37
38 534 keeping the habitat open without harming the population too much. Overgrowing with dense reed can be a
39
40 535 problem in some wet habitats. Mowing experiments since 2009 (Bräu & Völkl unpubl.) have already shown
41
42 536 promising results on reduction of reed if cutting is done during flight period with a cutter bar at a high level of
43
44 537 about 30 cm above ground. However, this kind of mowing should be restricted just to parts of the habitat with
45
46 538 dense reed to avoid emigration of butterflies and should mainly be used for habitat restoration.
47

48 539

50 540 **Acknowledgements**

51 541

52
53
54
55 542 Special thanks to all colleagues, who contributed to this study in providing information on other populations in
56
57 543 Europe and beyond, and to two anonymous referees for suggestions that greatly improved the manuscript. We
58
59 544 are grateful to R. Völkl and T. Hermann for support during field work, and to Giorgio Buffa for the
60
61
62
63
64
65

545 determination of the Italian plant species. We also particularly thank the Bavarian Academy for Nature
1
2 546 Conservation and Landscape Management (ANL) for funding this project. Many thanks also to the Bavarian
3
4 547 State Office for Environment for financing the investigations in all previous years and to the Government of
5
6 548 Upper Bavaria as well as local nature conservation authorities for their support. The Slovenian part of study was
7
8 549 partly funded by the Slovenian Research Agency (P1-0236).

9
10 550

11 551 **References**

12
13 552

14
15
16 553 Anthes N, Fartmann T, Hermann G, Kaule (2003) Combining larval habitat quality and metapopulation structure
17
18 554 – the key for successful management of pre-alpine *Euphydryas aurinia* colonies. *J Insect Conserv* 7: 175–185

19
20 555

21
22 556 Anthes N, Fartmann T, Hermann G (2008) The Duke of Burgundy butterfly and its dukedom: larval niche
23
24 557 variation in *Hamearis lucina* across Central Europe. *J Insect Conserv* 12: 3–14

25
26 558

27
28 559 Awmack CS, Leather SR (2002) Host plant quality and fecundity in herbivorous insects. *Annu Rev Entomol* 47:
29
30 560 817–844

31
32 561

33
34 562 Balletto E, Bonelli S, Cassulo L (2007) *Insecta Lepidoptera Papilionoidea*. In: Ruffo S, Stoch F (eds) *Checklist*
35
36 563 *and Distribution of the Italian Fauna*. 10.000 terrestrial and inland water species, 2nd and revised edition,
37
38 564 Verona, pp 257–261

39
40 565

41
42 566 Balletto E, Bonelli S, Zilli A (2014) *Lepidotteri*. In: Genovesi P, Angelini P, Bianchi E, Dupré E, Ercole S,
43
44 567 Giacanelli V, Ronchi F, Stoch F (2014) *Specie e habitat di interesse comunitario in Italia: distribuzione, stato di*
45
46 568 *conservazione e trend*. ISPRA, Serie Rapporti, 194/2014

47
48 569

49
50 570 Beyer LJ, Schultz CB (2010) Oviposition selection by rare grass skipper *Polites mardon* in montane habitats:
51
52 571 Advancing ecological understanding to develop conservation strategies. *Biol Conserv* 143: 862–872

53
54 572

55
56 573 Bischof A (1968) *Coenonympha oedippus* Fabricius, eine kleine Chorographie (Lepidoptera, Satyridae). *Mitt Ent*
57
58 574 *Ges Basel* 18: 41–63

59
60
61
62
63
64
65

- 1
2 576 Bonelli S, Canterino S, Balletto E (2010) Ecology of *Coenonympha oedippus* (Fabricius, 1787) (Lepidoptera:
3
4 577 Nymphalidae) in Italy. *Oedippus* 26: 25–30
5
6 578
7
8 579 Bräu M, Dolek M (2013) Wald-Wiesenvögelchen *Coenonympha hero* (Linnaeus, 1758). In: Bräu M, Bolz R,
9
10 580 Kolbeck H, Nunner A, Voith J, Wolf W (eds) Tagfalter in Bayern. Eugen Ulmer Verlag, Stuttgart, pp 472–475
11
12 581
13
14 582 Bräu M, Dolek M, Stettmer C (2010) Habitat requirements, larval development and food preferences of the
15
16 583 German population of the False Ringlet *Coenonympha oedippus* (Fabricius, 1787) (Lepidoptera: Nymphalidae) –
17
18 584 Research on the ecological needs to develop management tools. *Oedippus* 26: 41–51
19
20 585
21
22 586 Bräu M, Schwibinger M (2013) Moor-Wiesenvögelchen *Coenonympha oedippus* (Fabricius, 1787). In: Bräu M,
23
24 587 Bolz R, Kolbeck H, Nunner A, Voith J, Wolf W (eds): Tagfalter in Bayern. Eugen Ulmer Verlag, Stuttgart, pp
25
26 588 460–463
27
28 589
29
30 590 Čelik T (1997) Ecological researches of endangered species *Coenonympha oedippus* Fabricius, 1787
31
32 591 (Lepidoptera: Satyridae) on the Ljubljansko barje. M.Sc. Thesis. University of Ljubljana
33
34 592
35
36 593 Čelik T (2003) Population structure, migration and conservation of *Coenonympha oedippus* Fabricius, 1787
37
38 594 (Lepidoptera: Satyridae) in a fragmented landscape. Ph.D. Thesis. University of Ljubljana
39
40 595
41
42 596 Čelik T (2004) Population dynamics of endangered species *Coenonympha oedippus* Fabricius, 1787
43
44 597 (Lepidoptera: Satyridae) on the Ljubljansko barje. *Acta Ent Slo* 12: 99–114
45
46 598
47
48 599 Čelik T, Verovnik R (2010) Distribution, habitat preferences and population ecology of the False Ringlet
49
50 600 *Coenonympha oedippus* (Fabricius, 1787) (Lepidoptera: Nymphalidae) in Slovenia. *Oedippus* 26: 7–15
51
52 601
53
54 602 Čelik T, Vreš B, Seliškar A (2009a) Determinants of within-patch microdistribution and movements of
55
56 603 endangered butterfly *Coenonympha oedippus* (Fabricius, 1787) Nymphalidae: Satyrinae). *Hacquetia* 8/2: 115–
57
58 604 128
59
60
61
62
63
64
65

605

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

606 Čelik T, Vreš B, Seliškar A (2009b) The status of populations and habitats and recommendations for monitoring
607 of threatened species False Ringlet (*Coenonympha oedippus*), Snake'shead fritillary (*Fritillaria meleagris*) and
608 Loesel's twayblade (*Liparis loeselii*) on Ljubljansko barje. Final report (in Slovene with English Summary).
609 Biološki inštitut ZRC SAZU, Ljubljana

610

611 Chrétien MP (1886) Une note sur les premiers états du *Coenonympha oedippus*. In: Bourgeois MJ (ed) Séance
612 du 13 octobre 1886 (Bulletin entomologique). Ann Soc Entomol Fr 6: 638 (157).

613

614 Dennis RLH (2003) Towards a functional resource-based concept for habitat: a butterfly biology viewpoint.
615 Oikos 102: 417–426

616

617 Dennis RLH, Shreeve TG, Van Dyck H (2006) Habitats and resources: the need for a resource-based definition
618 to conserve butterflies. Biodivers Conserv 15: 1943–1966

619

620 De Vries HH, Ens SH, de Graaf G, Teunissen L, van der Velde R, Vogelaar L, Winterink A, Visser ME (2011)
621 Synchronisation of egg hatching of brown hairstreak (*Thecla betulae*) and budburst of blackthorn (*Prunus*
622 *spinosa*) in a warmer future. J Insect Conserv 15: 311–319

623

624 Dierks K (2006) Beobachtungen zur Larvalbiologie von *Coenonympha oedippus* (Fabricius, 1787) im
625 Südwesten Frankreichs (Lepidoptera: Satyridae). Entomologische Zeitschrift 116 (4): 186–188.

626

627 Doak P (2000) Population consequences of restricted dispersal for an insect herbivore in a subdivided habitat.
628 Ecology 81: 1828–1841

629

630 Doak P, Kareiva P, Kingsolver J (2006) Fitness consequences of choosy oviposition for time-limited butterfly.
631 Ecology 87: 395–408

632

633 Dolek M (2011) Wald-Wiesenvögelchen *Coenonympha hero* (Linnaeus, 1758). Merkblatt Artenschutz 37,
634 Bayerisches Landesamt für Umwelt, Augsburg

635

- 1
2 636 Dolek M, Freese A, Geyer A, Stetter H (2005) The decline of *Colias myrmidone* at the western edge of its range
3
4 637 and notes on its habitat requirements. *Biologia* 60: 607–610
5
6 638
7
8 639 Dolek M, Freese-Hager A, Geyer A, Balletto E, Bonelli S (2013) Multiple oviposition and larval feeding
9
10 640 strategies in *Euphydryas maturna* (Linné, 1758) (Nymphalidae) at two disjoint European sites. *J Insect Conserv*
11
12 641 17: 357–366
13
14 642
15
16 643 Dolek M, Geyer A (2001) Der Violette Feuerfalter (*Lycaena alciphron*): Artenhilfsprogramm für einen wenig
17
18 644 bekannten Tagfalter. *Schriftenreihe des Bayerischen LfU* 156: 341–354
19
20 645
21
22 646 Dolek M, Geyer A, Bolz R (1998) Distribution of *Maculinea rebeli* and host plant use along the river Danube. *J*
23
24 647 *Insect Conserv* 2: 85–89
25
26 648
27
28 649 Dušej G, Wemeille E, Carron G, Ziegler H (2010) Concerning the situation of the False Ringlet *Coenonympha*
29
30 650 *oedippus* (Fabricius, 1787) (Lepidoptera: Nymphalidae) in Switzerland. *Oedippus* 26: 38–40
31
32 651
33
34 652 Eilers S, Pettersson LB, Öckinger E (2013) Micro-climate determines oviposition site selection and abundance in
35
36 653 the butterfly *Pyrgus armoricanus* at its northern range margin. *Ecol Entomol* 38: 183–192
37
38 654
39
40 655 Ellenberg H (2009) *Vegetation ecology of Central Europe*. Cambridge University Press, Cambridge
41
42 656
43
44 657 Fartmann T (2005) Quendel-Ameisenbläuling *Glaucopsyche arion* (Linnaeus, 1758). *Naturschutz und*
45
46 658 *Biologische Vielfalt* 20: 175–180
47
48 659
49
50 660 Fortunel C, Garnier E, Joffre R, Kazakou E, Quested H, Grigulis K, Lavorel S, Ansquer P, Castro H, Cruz P
51
52 661 (2009) Leaf traits capture the effects of land use changes and climate on litter decomposability in grasslands
53
54 662 across Europe. *Ecology* 90: 598–611
55
56 663
57
58
59
60
61
62
63
64
65

664 Geyer A, Dolek M (1995) Ökologie und Schutz des Apollofalters (*Parnassius apollo*) in der Frankenalb. Mitt Dt
1
2 665 Gesell Allg Ang Ent 10: 333–336
3
4 666
5
6 667 Geyer A, Dolek M (2013) Gelbringfalter *Lopinga achine* (Scopoli, 1763). In: Bräu M, Bolz R, Kolbeck H,
7
8 668 Nunner A, Voith J, Wolf W (eds) Tagfalter in Bayern. Eugen Ulmer Verlag, Stuttgart, pp 452–455
9
10 669
11
12 670 Gorbunov P, Kosterin O (2007) The butterflies (Hesperioidea in et Papilionidea) of North Asia (assian part of
13
14 671 Russia) in Nature. Vol. 2, Rodina & Fodio, Moscow
15
16 672
17
18 673 Gradl F (1946) *Coenonympha oedippus* F. Z Wien Ent Ges 30: 14–20
19
20 674
21
22 675 Gripenberg S, Mayhew PJ, Parnell M, Roslin T (2010) A meta-analysis of preference–performance relationships
23
24 676 in phytophagous insects. Ecol Lett 13: 383–393
25
26 677
27
28 678 Habeler H (1972) Zur Kenntnis der Lebensräume von *Coenonympha oedippus* F. (Lep. Satyridae). Nachr Bayer
29
30 679 Ent 21/3: 51–54
31
32 680
33
34 681 Hafner J (1910) Makrolepidopteren von Görz und Umgebung. Entomologischen Zeitschrift, Sonder-Abdruck: 1–
35
36 682 40
37
38 683
39
40 684 Hermann G, Steiner R (1997) Eiablage und Larvalhabitat des Komma-Dickkopffalters (*Hesperia comma*
41
42 685 Linnaeus 1758) in Baden-Württemberg (Lepidoptera, Hesperiiidae). Carolea 55: 35–42
43
44 686
45
46 687 Janz N (2002) Evolutionary ecology of oviposition strategies. In: Hilker M, Meiners T (eds) Chemoecology of
47
48 688 insect eggs and egg deposition. Blackwell, Berlin, pp 349–376
49
50 689
51
52 690 Kalarus K, Skórka P, Nowicki P (2013) Resource use in two contrasting habitat types raises different challenges
53
54 691 for the conservation of the dryad butterfly *Minois dryas*. J Insect Conserv 17: 777–786
55
56 692
57
58
59
60
61
62
63
64
65

- 693 Kassai F & Peregovits L (2005) Contrasting egg laying behaviour of the ecotypes of *Maculinea alcon* in
1
2 694 Hungary. In: Settele J, Kühn E, Thomas JA (eds) Studies on the ecology and conservation of butterflies in
3
4 695 Europe. Vol. 2: Species ecology along a European gradient: Maculinea butterflies as a Model. Sofia - Moskau,
5
6 696 Pensoft Publishers, p 73
7
8 697
9
10 698 Kolar H (1919) Über das Vorkommen von *Coenonympha oedippus* F. Z Österr Entomol Ver Wien 4: 96
11
12 699
13
14 700 Kolar H (1929) Verbreitung von *Coenonympha oedippus* F. in Europa. Verh Zoo Bot Ver Wien 78: 105–108
15
16 701
17
18 702 Königsdorfer M (1997) Die Berghexe (*Chazara briseis* L.) in Schwaben und angrenzenden Gebieten. Ber Nat
19
20 703 Ver Schwaben 101: 69–87
21
22 704
23
24 705 Konvička M, Dvořák L, Pavličko A, Fric Z (2008) The Baton blue (*Pseudophilotes baton*) (Lepidoptera,
25
26 706 Lycaenidae) in south-western Bohemia: iron curtain, military ranges and endangered butterfly. Silva Gabreta 14:
27
28 707 187–198
29
30 708
31
32 709 Krämer B, Kämpf I, Enderle J, Poniatowski D, Fartmann T (2012) Microhabitat selection in a grassland
33
34 710 butterfly: a trade-off between microclimate and food availability. J Insect Conserv 16: 857–865
35
36 711
37
38 712 Kudrna O, Harpke A, Lux K, Pennerstorfer J, Schweiger O, Settele J, Wiemers M (2011) Distribution Atlas of
39
40 713 Butterflies in Europe. Gesellschaft für Schmetterlingsschutz e.V., Halle
41
42 714
43
44 715 Lafranchis T (2000) Les Papillons de jour de France, Belgique et Luxembourg et leurs chenilles. Collection
45
46 716 Parthenope, editions Biotope, Meze
47
48 717
49
50 718 Lauber K, Wagner G (1996) *Flora Helvetica*. Verlag Paul Haupt, Bern, Stuttgart, Wien
51
52 719
53
54 720 Lawson CR, Bennie J, Hodgson JA, Thomas CD, Wilson RJ (2014) Topographic microclimates drive
55
56 721 microhabitat associations at the range margin of a butterfly. Ecography 37: 732–740
57
58 722
59
60
61
62
63
64
65

- 723 Leopold P (2001) Schmetterlingszönosen ausgewählter Kalk-Magerrasen im Saale-Unstrut-Gebiet (Sachsen-
1 Anhalt) unter besonderer Berücksichtigung der Habitats des Segelfalters und der Berghexe. Dipl.-Arb. Inst. F.
2
3
4 725 Landschaftsökologie, Univ. Münster
5
6 726
7
8 727 Lhonore J (1996) *Coenonympha oedippus*. In: Helsdingen van PJ, Willemse L, Speight MCD. (eds) Background
9
10 728 information on invertebrates of the Habitats Directive and the Bern Convention. Part I – Crustacea, Coleoptera
11
12 729 and Lepidoptera. Council of Europe, Strasbourg, Nature and environment 79: 98–104
13
14 730
15
16 731 Lhonore J, Lagarde M (1999) Biogeographie, ecologie et protection de *Coenonympha oedippus* (Fab., 1787)
17
18 732 (Lepidoptera: Nymphalidae: Satyrinae). Ann Soc Entomol Fr 35 (suppl.): 299–307
19
20 733
21
22 734 Lindman L, Johansson B, Gotthard K, Tammamaru T (2013) Host plant relationships of an endangered butterfly,
23
24 735 *Lopinga achine* (Lepidoptera: Nymphalidae) in northern Europe. J Insect Conserv 17: 375–383
25
26 736
27
28 737 Loydi A, Eckstein RL, Otte A, Donath T (2013) Effects of litter on seedling establishment in natural and semi-
29
30 738 natural grasslands: a meta-analysis. J Ecol 101: 454–464
31
32 739
33
34 740 Mayhew PJ (1997) Adaptive Patterns of Host-Plant Selection by Phytophagous Insects. Oikos 79: 417–428
35
36 741
37
38 742 Möllenbeck V, Hermann G, Fartmann T (2009) does prescribed burning mean a threat to the rare satyrine
39
40 743 butterfly *Hipparchia fagi*? Larval-habitat preferences give the answer. J Insect Conserv 13: 77–87
41
42 744
43
44 745 Munguira ML, García-Barros E, Cano JM (2009) Butterfly herbivory and larval ecology. In: Settele J, Shreeve
45
46 746 T, Konvička M, Van Dyck H (eds) Ecology of butterflies in Europe. Cambridge University Press, Cambridge, pp
47
48 747 43–54
49
50 748
51
52 749 Nunner A (1995) Zur Autökologie von *Boloria eunomia* (Esper 1799) und *Lycaena helle* (Denis &
53
54 750 Schiffermueller 1775) (Lepidoptera: Rhopalocera) im bayerischen Alpenvorland. Dipl. Thesis, Universität
55
56 751 Tübingen
57
58 752
59
60
61
62
63
64
65

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65
- 753 Obermaier E, Heisswolf A, randlkofer B, Meiners T (2006) Enemies in low places – insects avoid winter
754 mortality and egg parasitism by modulating oviposition height. *B Entomol Res* 96: 337–343
755
756 O'Connor RS, Hails RS, Thomas JA (2014) Accounting for habitat when considering climate: has the niche of
757 the Adonis blue butterfly changed in the UK? *Oecologia* 174: 1463–1472
758
759 Örvössy N, Vozár Á, Kőrösi Á, Batáry P, Peregovits L (2010) Structure and size of a threatened population of
760 the False Ringlet *Coenonympha oedippus* (Fabricius, 1787) (Lepidoptera: Nymphalidae) in Hungary. *Oedippus*
761 26: 31–37
762
763 Örvössy N, Vozár Á, Kőrösi Á, Batáry P, Peregovits L (2013) Potential metapopulation structure and the effects
764 of habitat quality on population size of the endangered False Ringlet butterfly. *J Insect Conserv* 17: 537–547
765
766 Pauler R, Kaule G, Verhaagh M, Settele J (1995) Untersuchungen zur Autökologie des Schwarzgefleckten
767 Ameisenbläulings, *Maculinea arion* (Linnaeus 1758) (Lepidoptera, Lycaenidae) in Südwestdeutschland. *Nachr*
768 *Entomol Ver Apollo* 16: 147–186
769
770 Pennekamp F, Monteiro E, Schmitt T (2013) The larval ecology of the butterfly *Euphydryas desfontainii*
771 (Lepidoptera: Nymphalidae) in SW Portugal: food plant quantity and quality as main predictors of habitat
772 quality. *J Insect Conserv* 17: 195–206
773
774 Ruehl (1895) Die paläarktischen Großschmetterlinge und ihre Naturgeschichte. Lief. 55, Leipzig
775
776 Ruprecht E & Szabo A (2012) Grass litter is a natural seed trap in long-term undisturbed grassland. *J Veg Sci* 23:
777 495–504
778
779 Ruprecht E, Enyedi MZ, Eckstein RL, Donath TW (2010) Restorative removal of plant litter and vegetation 40
780 years after abandonment enhances re-emergence of steppe grassland vegetation. *Biol Conserv* 143: 449–456
781

782 Šašić M (2010) False Ringlet *Coenonympha oedippus* (Fabricius, 1787) (Lepidoptera: Nymphalidae) in Croatia:
1
2 783 current status, population dynamics and conservation management. *Oedippus* 26: 16–19
3
4 784
5
6 785 Sachteleben J & Winterholler M (2013) Blaukernaue *Minois dryas* (Scopoli, 1763). In: Bräu M, Bolz R,
7
8 786 Kolbeck H, Nunner A, Voith J, Wolf W (eds) Tagfalter in Bayern. Eugen Ulmer Verlag, Stuttgart, pp 527–529
9
10 787
11
12 788 Selezniev M, Pałka K, Michalczyk W, Bystrowski C, Hołowiński M, Czerwiński M (2010) False Ringlet
13
14 789 *Coenonympha oedippus* (Fabricius, 1787) (Lepidoptera: Nymphalidae) in Poland: state of knowledge and
15
16 790 conservation prospects. *Oedippus* 26: 20–24
17
18 791
19
20 792 Settele J, Kudrna O, Hharpke A, Kühn I, van Swaay C, Verovnik R, Warren M, Wiemers M, Hanspach J,
21
22 793 Hickler T, Kühn E, van Halder I, Veling K, Vliegenthart A, Wynhoff I, Schweiger O (2008) Climatic Risk Atlas
23
24 794 of European Butterflies. Pensoft, Sofia-Moscow
25
26 795
27
28 796 SPSS Inc. 1989–2004. *SPSS for Windows. Release 13.0* (1 Sep 2004)
29
30 797
31
32 798 Steiner R, Hermann G (1999) Freilandbeobachtungen zu Eiablageverhalten und –habitat des Wald-
33
34 799 Wiesenvögelchen, *Coenonympha hero* (Linnaeus, 1761), an einer Flugstelle in Baden-Württemberg. Nach
35
36 800 Entomol Ver Apollo 20: 111–118
37
38 801
39
40 802 Stevenson RD (1985) The relative importance of behavioral and physiological adjustments controlling
41
42 803 temperature in terrestrial ectotherms. *Am Nat* 126: 362–386
43
44 804
45
46 805 Szentirmai I, Mesterházy A, Varga I, Schubert Z, Sándor LC, Ábrahám L, Körösi Á (2014) Habitat use and
47
48 806 population biology of the Danube Clouded Yellow butterfly *Colias myrmidone* (Lepidoptera: Pieridae) in
49
50 807 Romania. *J Insect Conserv* 18: 417–425
51
52 808
53
54 809 Thomas JA (1980) Why did the Large Blue become extinct in Britain? *Oryx* 15: 243–247
55
56 810
57
58
59
60
61
62
63
64
65

811 Thomas JA (1983) The ecology and conservation of *Lysandra bellargus* (Lepidoptera, Lycaenidae) in Britain. J
1
2 812 Appl Ecol 20: 59–83
3
4 813
5
6 814 Thomas JA (1993) Holocene climate change and warm man-made refugia may explain why a sixth of british
7
8 815 butterflies inhabit unnatural early successional habitats. Ecography 16: 278–284
9
10 816
11
12 817 Thomas JA & Elmes GW (2001) Food-plant niche selection rather than the presence of ant nests explains
13
14 818 oviposition patterns in the myrmecophilous butterfly genus *Maculinea*. P Roy Soc Lond B Bio 268: 471–477
15
16 819
17
18 820 Tshikolovets VV (2003) Butterflies of Eastern Europe, Urals and Caucasus. Brno
19
20 821
21
22 822 Turlure C, Chouff J, Baguette M, van Dyck H (2010) Microclimatic buffering and resource-based habitat in a
23
24 823 glacial relict butterfly: significance for conservation under climate change. Glob Change Biol 16: 1883–1893
25
26 824
27
28 825 Turlure C, Radchuk V, Baguette M, Meijrink M, van den Burg A, WallisDeVries M, van Duinen GJ (2013)
29
30 826 Plant quality and local adaptation undermine relocation in a bog specialist butterfly. Ecology and Evolution 3:
31
32 827 244–254
33
34 828
35
36 829 Turlure C, Radchuk V, Baguette M, van Dyck H, Schtickzelle N (2011) On the significance of structural
37
38 830 vegetation elements for caterpillar thermoregulation in two peat bog butterflies: *Boloria eunomia* and *B.*
39
40 831 *aquilonaris*. J Thermal Biol 36: 173–180
41
42 832
43
44 833 Van Dyck H (2012) Changing organisms in rapidly changing anthropogenic landscapes: the significance of the
45
46 834 “Umwelt”-concept and functional habitat for animal conservation. Evol Appl 5: 144–153
47
48 835
49
50 836 Van Swaay CAM, Cuttelod A, Collins S, Maes D, Munguira López M, Šašić M, Settele J, Verovnik R, Verstrael
51
52 837 T, Warren M, Wiemers M, Wynhoff I (2010) European Red List of Butterflies. Publications Office of the
53
54 838 European Union, Luxembourg
55
56 839
57
58 840 Wagner W (2010) Europäische Schmetterlinge und ihre Ökologie. www.pyrgus.de
59
60
61
62
63
64
65

841

- 1
2 842 WallisDeVries MF, van Swaay CAM (2006) Global warming and excess nitrogen may induce butterfly decline
3
4 843 by microclimate cooling. *Glob Change Biol* 12: 1620–1626
5
6 844
7
8 845 Ward G, Hastie T, Barry S, Elith J, Leathwick JR (2009) Presence-only data and the EM algorithm. *Biometrics*
9
10 846 65: 554–563
11
12 847
13
14 848 Weidemann HJ (1995) *Tagfalter: beobachten, bestimmen*. 2nd Ed., völlig neu bearb. Aufl. Naturbuch Verlag,
15
16 849 Augsburg
17
18 850
19
20 851 Weking S, Hermann G, Fartmann T (2013) Effects of mire type, land use and climate on a strongly declining
21
22 852 wetland butterfly. *J Insect Conserv* 17: 1081–1091
23
24 853
25
26 854 Wiklund C (1984) Egg-laying patterns in butterflies in relation to their phenology and visual apparency and
27
28 855 abundance of their host plants. *Oecologia* 63: 23–29
29
30 856
31
32 857 Zalucki MP, Clarke AR, Malcolm SB (2002) Ecology and behaviour of first instar larval Lepidoptera. *Annu Rev*
33
34 858 *Entomol* 47: 361–393
35
36 859
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

860 **Table captions**

1
2 861

3
4 862 **Table 1** Substrata used by feeding larvae and egg-laying females of *C. oedippus* in Germany (DE), Italy (IT) and
5
6 863 Slovenia (SLO). Given are the numbers of observations, in brackets percentage of all observations in the
7
8 864 respective country

9
10 865

11 866 **Table 2** Univariate comparison between larval microhabitats and available microlocations after winter (in May)
12
13
14 867 of *C. oedippus* in Germany, Italy and Slovenia (note, that for Slovenia no information is disposable for available
15
16 868 sites). Median (Min–Max) values of parameters and significance (p) of comparison using Mann-Whitney test are
17
18 869 presented. B = basic variable, D = derived variable, C = composed variable, S = simple variable (see Material
19
20 870 and methods for explanations)

21
22 871

23
24 872 **Table 3** Analysis of preferences of overwintering larvae of *C. oedippus* based on occupied and available
25
26 873 microlocations using binary stepwise-forward logistic regression

27
28 874

29
30 875 **Table 4** Univariate comparison between oviposition microhabitats and available microlocations of *C. oedippus*
31
32 876 in Germany, Italy and Slovenia. Median (Min–Max) values of parameters and significance (p) of comparison
33
34 877 using Mann-Whitney test are presented. B = basic variable, D = derived variable, C = composed variable, S =
35
36 878 simple variable (see Material and methods for explanations)

37
38 879

39
40 880 **Table 5** Analysis of preferences of egg-laying females of *C. oedippus* based on occupied and available
41
42 881 microlocations using binary stepwise-forward logistic regression

43
44 882

45
46 883 **Table 6** Dates of *C. oedippus* larval awakening and of host-plants availability provided by field observations (on
47
48 884 site) and by breeding under field conditions (ex-situ) in Germany (DE) and Slovenia (SLO)

49
50 885

51
52 886 **Table 7** Egg-laying plants used by *C. oedippus* in Europe known before present study

53
54 887

55
56 888

57
58 889 **Figure caption**

59
60
61
62
63
64
65

890

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

Fig. 1 Coverage of main structural parameters in (a) larval and (b) oviposition microhabitats of *C. oedippus* in Germany (DE), Italy (IT) and Slovenia (SLO). Mann-Whitney tests: **(a)** *litter*, DE > IT: $p = 0.000$, DE > SLO: $p = 0.000$, IT \approx SLO: $p = 0.831$; *herbs*, DE \approx IT: $p = 0.297$, DE < SLO: $p = 0.005$, IT < SLO: $p = 0.001$; *shrubs*, DE < IT: $p = 0.000$, DE < SLO: $p = 0.000$, IT > SLO: $p = 0.000$; *litter+C. vulgaris*, DE \approx IT: $p = 0.724$, IT > SLO: $p = 0.001$. **(b)** *litter*, DE \approx SLO: $p = 0.577$, DE > IT: $p = 0.000$, SLO > IT: $p = 0.000$; *shrubs*, DE \approx SLO: $p = 0.545$; DE < IT: $p = 0.000$, SLO < IT: $p = 0.000$; *Herbs without hostplants*, SLO > DE: $p = 0.000$; SLO > IT: $p = 0.000$; DE > IT: $p = 0.000$; *hostplants*, DE \approx IT: $p = 0.360$; DE > SLO: $p = 0.000$; IT > SLO: $p = 0.000$; *litter+C. vulgaris*, DE \approx IT: $p = 0.053$, IT > SLO: $p = 0.021$

Figure1a
[Click here to download high resolution image](#)

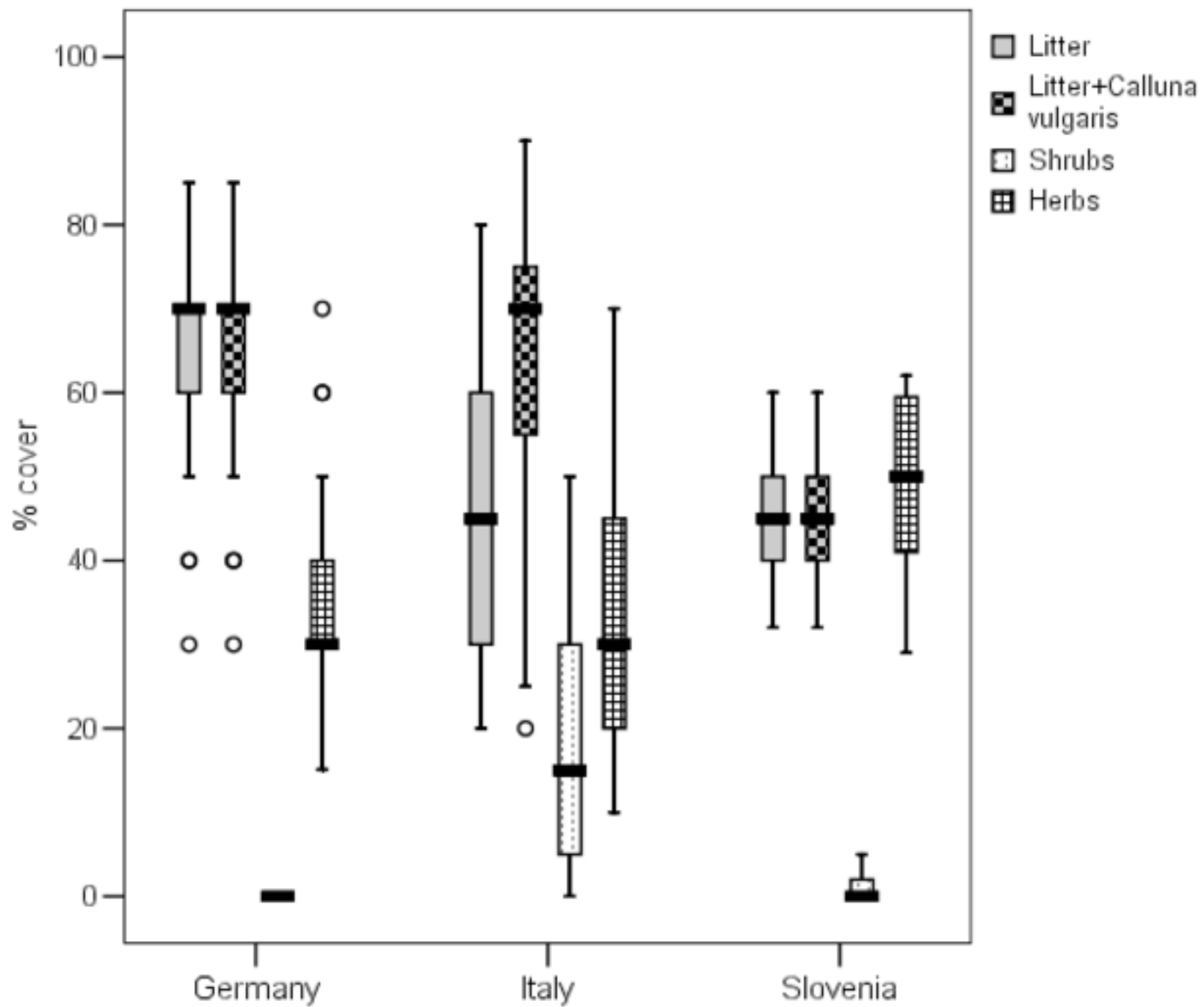


Figure1b
[Click here to download high resolution image](#)

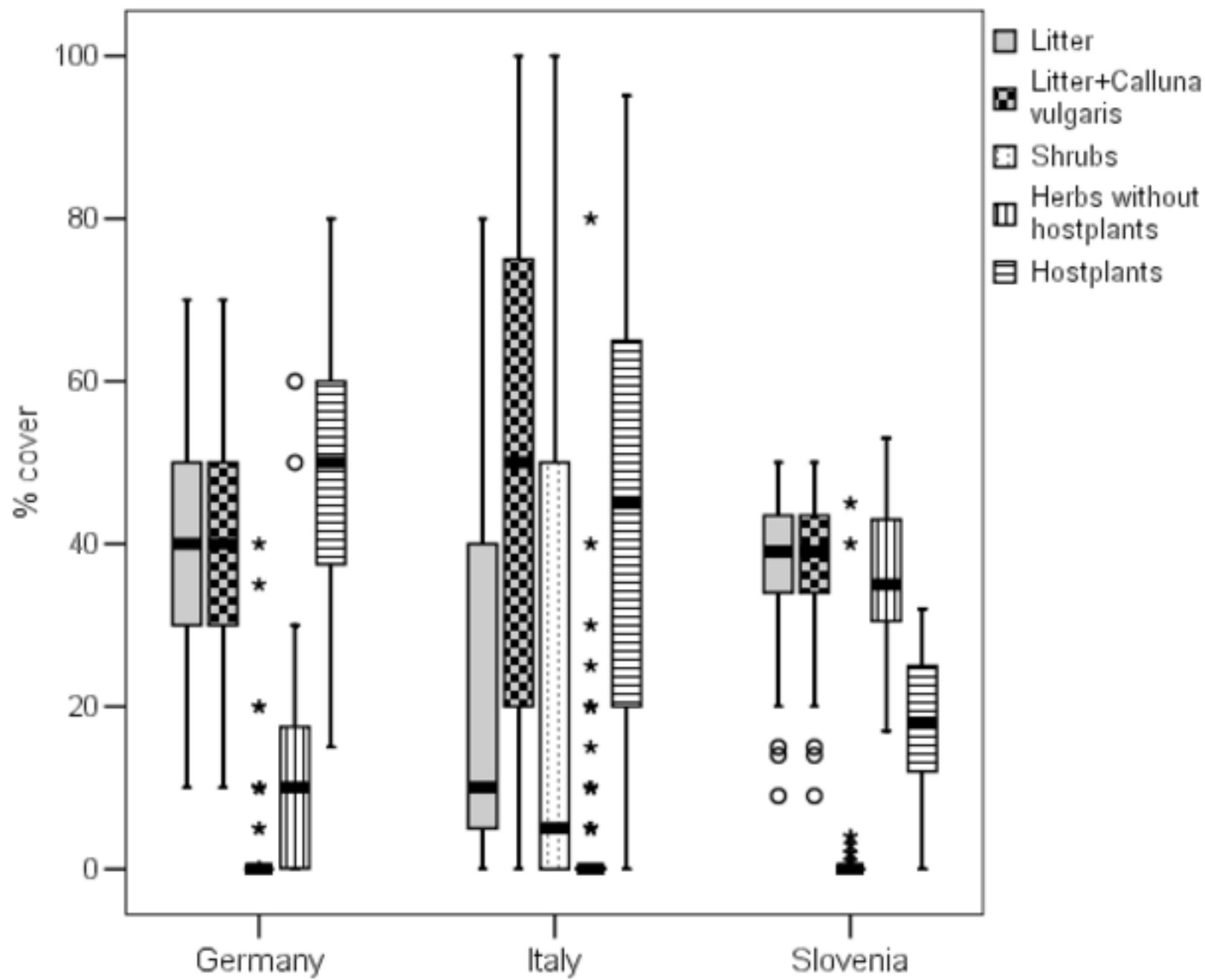


Table 1 Substrata used by feeding larvae and egg-laying females of *C. oedippus* in Germany (DE), Italy (IT) and Slovenia (SLO). Given are the numbers of observations, in brackets percentage of all observations in the respective country

| Substrata | | Larval host plants | | | Oviposition supports | | |
|-------------------------------|---|--------------------|---------|---------|----------------------|---------|---------|
| | | DE | IT | SLO | DE | IT | SLO |
| | | N = 38 | N = 34 | N = 13 | N = 79 | N = 102 | N = 55 |
| <i>Molinia caerulea</i> | H | 10 (26) | 2 (6) | – | 43 (54) | 34 (33) | – |
| <i>Deschampsia caespitosa</i> | H | 0 | – | – | 1 (1) | – | – |
| <i>Festuca rupicola</i> | H | – | – | 3 (23) | – | – | 3 (5) |
| <i>Sesleria autumnalis</i> | H | – | – | 0 | – | – | 2 (4) |
| <i>Bromopsis condensata</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Bromopsis erecta</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Stipa sp.</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Chrysopogon gryllus</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Melica ciliata</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Anthoxanthum odoratum</i> | H | 0 | 0 | 0 | 1 (1) | 0 | 0 |
| <i>Carex panicea</i> | H | 27 (71) | 28 (82) | – | 22 (28) | 9 (9) | – |
| <i>Carex davalliana</i> | H | 1 (3) | – | – | 0 | – | – |
| <i>Carex humilis</i> | H | – | – | 10 (77) | – | – | 36 (65) |
| other <i>Carex</i> species | H | 0 | 4 (12) | 0 | 0 | 0 | 0 |
| <i>Cladium mariscus</i> | H | 0 | – | – | 1 (1) | – | – |
| <i>Eupatorium cannabinum</i> | H | 0 | – | – | 2 (3) | – | – |
| <i>Valeriana dioica</i> | H | 0 | 0 | – | 1 (1) | 0 | – |
| <i>Asperula cynanchica</i> | H | – | – | 0 | – | – | 3 (5) |
| <i>Anthericum ramosum</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Galium purpureum</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Genista sylvestris</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Satureja montana</i> | H | – | – | 0 | – | – | 1 (2) |
| <i>Calluna vulgaris</i> | S | – | 0 | – | – | 42 (41) | – |
| <i>Rhamnus catharticus</i> | S | 0 | – | 0 | 1 (1)* | – | 0 |
| <i>Frangula alnus</i> | S | 0 | 0 | – | 0 | 1(1) | – |
| <i>Ligustrum vulgare</i> | S | – | – | 0 | – | – | 1 (2) |
| litter | L | 0 | 0 | 0 | 3 (4) | 0 | 1 (2) |
| NA | | 0 | 0 | 0 | 4 (5) | 16 (16) | 0 |

H = herb, S = shrub, L = litter

NA = undetermined substratum

0 = substratum present at the study site but not used

– = substratum not present at the study site

* This egg was erroneously given as laid on *Rhamnus frangula* (*Frangula alnus*) in Bräu et al. (2010)

Table 2 Univariate comparison between larval microhabitats and available microlocations after winter (in May) of *C. oedippus* in Germany, Italy and Slovenia (note, that for Slovenia no information is disposable for available sites). Median (Min–Max) values of parameters and significance (*p*) of comparison using Mann-Whitney test are presented. B = basic variable, D = derived variable, C = composed variable, S = simple variable (see Material and methods for explanations)

| Parameter | Type of variable | Germany | | | Italy | | | Slovenia |
|---------------------------------------|------------------|-----------------|--------------------|----------|-----------------|--------------------|----------|-----------------|
| | | Larval (N = 31) | Available (N = 39) | <i>p</i> | Larval (N = 34) | Available (N = 10) | <i>p</i> | Larval (N = 11) |
| Bare ground (%) | B, S | 0.00 (0–20)\$ | 0.00 (0–10)\$ | 0.854 | 0.00 (0–20) | 10.00 (0–50) | 0.001 | 0.00 (0–4) |
| Rocks (%) | B, S | ^a | ^a | | ^a | ^a | | 0.00 (0–20) |
| Mosses (%) | B, S | ^a | ^a | | 0.00 (0–5.0) | 0.00 (0–0) | 0.689 | 0.00 (0–3) |
| Litter (%) | B, S | 70.00 (30–85) | 40.00 (20–70) | 0.000 | 45.00 (20–80) | 20.00 (10–30) | 0.000 | 45.00 (32–60) |
| Shrubs (%) | B, C | 0.00 (0–0) | 0.00 (0–15) | 0.001 | 15.00 (0–50) | 10.00 (0–40) | 0.591 | 0.00 (0–5) |
| Herbs (%) | B, C | 30.00 (15–70) | 55.10 (30–80) | 0.000 | 30.00 (10–70) | 47.50 (30–70) | 0.005 | 50.00 (29–62) |
| HPs (%) | D, C | 30.00 (15–65) | 40.00 (20–70) | 0.000 | n.a. | n.a. | | 12.00 (8–20) |
| Herbs without HPs (%) | D, S | 0.00 (0–20) | 10.00 (0–40) | 0.000 | n.a. | n.a. | | 38.00 (12–51) |
| Herbs without PJC (%) | D, S | 0.00 (0–10) | 5.00 (1–30) | 0.000 | n.a. | n.a. | | 12.00 (3–24) |
| <i>Calluna vulgaris</i> (%) | D, S | ^a | ^a | | 12.50 (0–50) | 10.00 (0–40) | 0.572 | ^a |
| Litter+ <i>C. vulgaris</i> (%) | D, S | ^a | ^a | | 70.00 (20–90) | 25.00 (15–70) | 0.001 | ^a |
| Shrubs without <i>C. vulgaris</i> (%) | D, S | ^a | ^a | | 0.00 (0–15) | 0.00 (0–5) | 0.923 | ^a |
| <i>Molinia caerulea</i> (%) | D, S | 15.00 (0–60) | 30.00 (10–70) | 0.000 | n.a. | n.a. | | ^a |
| <i>Carex panicea</i> (%) | D, S | 10.00 (5–40) | 5.00 (0–40) | 0.022 | n.a. | n.a. | | ^a |
| <i>Carex humilis</i> (%) | D, S | ^a | ^a | | ^a | ^a | | 10.00 (6–15) |
| <i>Festuca rupicola</i> (%) | D, S | ^a | ^a | | ^a | ^a | | 2.00 (0–12) |
| PJC (%) | D, S | 30.00 (15–67) | 50.00 (30–70) | 0.000 | n.a. | n.a. | | 37.00 (24–54) |
| GLH (%) | D, S | 0.00 (0–3) | 0.00 (0–0.1) | 0.604 | n.a. | n.a. | | 2.00 (1–12) |
| NGLH (%) | D, S | 0.00 (0–10) | 5.00 (0–30) | 0.000 | n.a. | n.a. | | 8.00 (1–18) |

HPs = Host plants (Germany: *M. caerulea* + *Carex panicea*; Slovenia: *Carex humilis* + *Festuca rupicola*)

PJC = Poaceace, Juncaceae, Cyperaceae (i.e. plants with erectophile leaf orientation)

GLH = Grass-like herbs (i.e. herbs with plant stature which does not shade the lower parts of herb layer; e.g., *Allium* spp., *Anthericum ramosum*, *Genista sylvestris*)

NGLH = Non grass-like herbs (i.e. herbs with plant stature which shades the lower parts of herb layer; plants with planophile leaf orientation) = cover of Herbs – cover of GLH

^aNot existent in the microhabitat

n.a. = data not available

\$ Bare ground was present only in one larval and one random microhabitat.

Table 3 Analysis of preferences of overwintering larvae of *C. oedippus* based on occupied and available microlocations using binary stepwise-forward logistic regression

| Parameter | (B) | SE (B) | Wald | P | Exp (B) | 95% CI for Exp (B) | |
|---|--------|--------|--------|-------|---------|--------------------|-------|
| | | | | | | Lower | Upper |
| <i>(a) Germany^a</i> | | | | | | | |
| Litter | 0.114 | 0.027 | 18.353 | 0.000 | 1.121 | 1.064 | 1.181 |
| Const. | -6.560 | 1.544 | 18.047 | 0.000 | 0.001 | | |
| Model $\chi^2 = 31.186$, df = 1, P < 0.0001, Nagelkerke R ² = 0.481, correctly classified 81.4% (larvae: 83.9%, available: 79.5%) | | | | | | | |
| <i>(b) Italy^b</i> | | | | | | | |
| Litter | 0.171 | 0.061 | 7.862 | 0.005 | 1.187 | 1.053 | 1.337 |
| Const. | -3.814 | 1.606 | 5.642 | 0.018 | 0.022 | | |
| Model $\chi^2 = 21.011$, df = 1, P < 0.0001, Nagelkerke R ² = 0.577, correctly classified 81.8% (larvae: 88.2%, available: 60.0%) | | | | | | | |

(a) Larvae sites (N = 31), available sites (N = 39); *(b)* larvae sites (N = 34), available sites (N = 10)

^a Variables entered into the regression analysis: coverage (%) of litter, bare ground, shrubs, HPs, herbs without HPs, *Carex panicea*, *Molinia caerulea*, GLH, NGLH; variables excluded from analysis because of multicollinearity: coverage (%) of herbs, PJC, herbs without PJC

^b Variables entered into the regression analysis: coverage (%) of litter, bare ground, mosses, litter+*Calluna vulgaris*, *Calluna vulgaris*, shrubs without *C. vulgaris*; variables excluded from analysis because of multicollinearity: coverage (%) of shrubs, herbs

Table 4 Univariate comparison between oviposition microhabitats and available microlocations of *C. oedippus* in Germany, Italy and Slovenia. Median (Min–Max) values of parameters and significance (*p*) of comparison using Mann-Whitney test are presented. B = basic variable, D = derived variable, C = composed variable, S = simple variable (see Material and methods for explanations)

| Parameter | Type of variable | Germany | | | Italy | | | Slovenia | | |
|---------------------------------------|------------------|----------------------|--------------------|----------|-----------------------|---------------------|----------|----------------------|--------------------|----------|
| | | Oviposition (N = 76) | Available (N = 35) | <i>p</i> | Oviposition (N = 101) | Available (N = 150) | <i>p</i> | Oviposition (N = 55) | Available (N = 30) | <i>p</i> |
| Bare ground (%) | B, S | 0.00 (0–10)\$ | 0.00 (0–0) | | 0.00 (0–75) | 0.00 (0–50) | 0.064 | 0.00 (0–5) | 0.00 (0–3) | 0.478 |
| Rocks (%) | B, S | a | a | | a | a | | 2.00 (0–55) | 1.00 (0–15) | 0.099 |
| Mosses (%) | B, S | a | a | | 0.00 (0–0.1) | 0.00 (0–70) | 0.922 | 0.00 (0–3) | 0.10 (0–5) | 0.196 |
| Litter (%) | B, S | 40.00 (10–70) | 40.00 (10–70) | 0.823 | 10.00 (0–80) | 5.00 (0–70) | 0.000 | 39.00 (9–50) | 33.00 (20–50) | 0.010 |
| Shrubs (%) | B, C | 0.00 (0–40)\$\$ | 0.00 (0–40)\$\$ | 0.012 | 5.00 (0–100) | 25.00 (0–100) | 0.033 | 0.00 (0–45) | 0.00 (0–55) | 0.025 |
| Herbs (%) | B, C | 60.00 (30–90) | 60.00 (30–80) | 0.268 | 50.00 (0–95) | 47.50 (0–100) | 0.506 | 56.00 (25–75) | 60.00 (20–70) | 0.018 |
| HPs | D, C | 50.00 (15–80) | 40 (20–70.0) | 0.397 | 45 (0–95.1) | 40.00 (0–100) | 0.053 | 18 (0–32) | 15 (0–42) | 0.174 |
| Herbs without HPs (%) | D, S | 10.00 (0–60) | 10.00 (0–50) | 0.632 | 0.00 (0–80) | 0.00 (0–95) | 0.841 | 35.00 (17–53) | 43.00 (14–67) | 0.024 |
| Herbs without PJC (%) | D, S | 5.00 (0–30) | 5.00 (0–30) | 0.948 | n.a. | n.a. | | 14.00 (6–33) | 15.50 (3–35) | 0.160 |
| <i>Calluna vulgaris</i> (%) | D, S | a | a | | 5.00 (0–100) | 20.00 (0–100) | 0.327 | a | a | |
| Litter+ <i>C. vulgaris</i> (%) | D, S | a | a | | 50.00 (0–100) | 40.00 (0–100) | 0.307 | a | a | |
| Shrubs without <i>C. vulgaris</i> (%) | D, S | a | a | | 0.00 (0–30) | 0.00 (0–70) | 0.000 | a | a | |
| <i>Molinia caerulea</i> (%) | D, S | 30.00 (0–70) | 40.00 (0–70) | 0.972 | 40.00 (0–95) | 30.00 (0–100) | 0.031 | a | a | |
| <i>Carex panicea</i> (%) | D, S | 10.00 (0–40) | 10.00 (0–60) | 0.277 | 0.00 (0–40) | 0.00 (0–65) | 0.087 | a | a | |
| <i>Carex humilis</i> (%) | D, S | a | a | | a | a | | 15.00 (2–30) | 15.00 (0–40) | 0.230 |
| <i>Festuca rupicola</i> (%) | D, S | a | a | | a | a | | 1.00 (0–13) | 0.00 (0–15) | 0.055 |
| PJC (%) | D, S | 50.00 (25–80) | 40.00 (30–75) | 0.223 | n.a. | n.a. | | 41.00 (13–53) | 41.50 (16–62) | 0.596 |
| GLH (%) | D, S | 0.00 (0–0) | 0.00 (0–0) | 1.000 | n.a. | n.a. | | 2.00 (0–8) | 4.00 (0–15) | 0.039 |
| NGLH (%) | D, S | 5.00 (0–30) | 5.00 (0–30) | 0.625 | n.a. | n.a. | | 10.00 (3–30) | 13.50 (2–30) | 0.733 |
| Average veget. height (cm) | S | n.a. | n.a. | | n.a. | n.a. | | 20.00 (8–35) | 20.00 (10–40) | 0.205 |
| Max. veget. height (cm) | S | n.a. | n.a. | | n.a. | n.a. | | 45.00 (25–65) | 40.00 (30–70) | 0.027 |

HPs = Host plants (Germany, Italy: *M. caerulea* + *Carex panicea*; Slovenia: *Carex humilis* + *Festuca rupicola*)

PJC = Poaceae, Juncaceae, Cyperaceae (i.e. plants with erectophile leaf orientation)

GLH = Grass-like herbs (i.e. herbs with plant stature which does not shade the lower parts of herb layer; e.g., *Allium* spp., *Anthericum ramosum*, *Genista sylvestris*)

NGLH = Non Grass-like herbs (i.e. herbs with plant stature which shades the lower parts of herb layer; plants with planophile leaf orientation) = cover of Herbs – cover of GLH

^aNot existent in the microhabitat

n.a. = data not available

\$ Bare ground was present only in one oviposition microhabitat

\$\$ Mean±SD, oviposition: 2.44±7.14; available: 5.06±9.21

Table 5 Analysis of preferences of egg-laying females of *C. oedippus* based on occupied and available microlocations using binary stepwise-forward logistic regression

| Parameter | (B) | SE (B) | Wald | P | Exp (B) | 95% CI for Exp (B) | |
|---|--------|--------|--------|-------|---------|--------------------|-------|
| | | | | | | Lower | Upper |
| <i>(a) Italy^a</i> | | | | | | | |
| Litter | 0.024 | 0.007 | 10.750 | 0.001 | 1.024 | 1.010 | 1.039 |
| Herbs without HPs | -0.020 | 0.010 | 3.923 | 0.048 | 0.980 | 0.961 | 1.000 |
| Shrubs without <i>C. vulgaris</i> | -0.086 | 0.034 | 6.340 | 0.012 | 0.918 | 0.859 | 0.981 |
| Const. | -0.524 | 0.200 | 6.868 | 0.009 | 0.592 | | |
| Model $\chi^2 = 35.868$, $df = 3$, $P < 0.0001$, Nagelkerke $R^2 = 0.180$, correctly classified 65.7% (eggs: 37.6%, available: 84.7%) | | | | | | | |
| <i>(b) Slovenia^b</i> | | | | | | | |
| Shrubs | -0.119 | 0.034 | 12.117 | 0.000 | 0.888 | 0.830 | 0.949 |
| Herbs without HPs | -0.146 | 0.038 | 14.686 | 0.000 | 0.864 | 0.802 | 0.931 |
| GLH | -0.246 | 0.117 | 4.401 | 0.036 | 0.782 | 0.621 | 0.984 |
| Max. vegetation height | 0.114 | 0.039 | 8.641 | 0.003 | 1.121 | 1.039 | 1.210 |
| Const. | 2.759 | 1.787 | 2.383 | 0.123 | 15.784 | | |
| Model $\chi^2 = 33.948$, $df = 4$, $P < 0.0001$, Nagelkerke $R^2 = 0.456$, correctly classified 78.6% (eggs: 88.9%, available: 60.0%) | | | | | | | |

(a) Oviposition sites ($N = 101$), available sites ($N = 150$); *(b)* oviposition sites ($N = 54$), available sites ($N = 30$)

^a Variables entered into the regression analysis: coverage (%) of litter, bare ground, mosses, herbs, herbs without HPs, *Calluna vulgaris*, shrubs without *C. vulgaris*, *Carex panicea*, *Molinia caerulea*, litter+*Calluna vulgaris*; variables excluded from analysis because of multicollinearity: coverage (%) of shrubs, HPs

^b Variables entered into the regression analysis: coverage (%) of litter, bare ground, mosses, rocks, shrubs, herbs, herbs without HPs, *Carex humilis*, *Festuca rupicola*, PJC, GLH, NGLH, average vegetation height, maximum vegetation height; variables excluded from analysis because of multicollinearity: coverage (%) of HPs (*C. humilis*, *F. rupicola*), herbs without PJC

Table 6 Dates of *C. oedippus* larval awakening and of host-plants availability provided by field observations (on site) and by breeding under field conditions (ex-situ) in Germany (DE) and Slovenia (SLO)

| Year | Locality | Larval awakening | First observation of larval feeding on | | | | Growing of <i>M. caerulea</i> |
|------|---------------|--------------------------|--|------------------------|------------------------|-----------------------|-------------------------------|
| | | | <i>C. panicea</i> | <i>M. caerulea</i> | <i>C. humilis</i> | <i>F. rupicola</i> | |
| 2008 | on site (DE) | n.a. | n.a. | n.a. | ^a | ^a | 27 th April |
| 2009 | ex-situ (DE) | 4 th April | 4 th April | 17 th April | ^a | ^a | 14 th April |
| 2009 | on site (DE) | 6 th April | 6 th April | n.a. | ^a | ^a | n.a. |
| 2010 | ex-situ (DE) | 23 rd March | 23 rd March | 3 rd April | ^a | ^a | n.a. |
| 2010 | on site (DE) | n.a. | n.a. | n.a. | ^a | ^a | after 25 th April |
| 2011 | ex-situ (DE) | 6 th February | 20 th March | n.a. | ^a | ^a | after 25 th March |
| 2011 | on site (SLO) | 2 nd April | ^a | ^a | 2 nd April | 2 nd April | ^a |
| 2012 | ex-situ (DE) | 16 th March | 18 th March | 21 st April | ^a | ^a | n.a. |
| 2012 | ex-situ (SLO) | 24 th March | ^a | ^a | 25 th March | n.a. | ^a |
| 2013 | ex-situ (DE) | 13 th April | 13 th April | 1 st May | ^a | ^a | n.a. |

n.a. = data not available

^aPlant species not existent in the study site

Table 7 Egg-laying plants used by *C. oedippus* in Europe known before present study

| Plant species | Country | Reference |
|------------------------------|----------------------|---|
| <i>Molinia caerulea</i> | SLO, IT, DE, PL, CRO | Čelik 1997, Čelik et al. 2009a, Bonelli et al. 2010, Bräu et al. 2010, Sielezniew et al. 2010, Šašić 2010 |
| <i>Carex panicea</i> | SLO, DE | Čelik 1997, Čelik et al. 2009a, Bräu et al. 2010 |
| <i>Carex davalliana</i> | SLO, PL | Čelik et al. 2009a, Sielezniew et al. 2010 |
| <i>Carex hostiana</i> | SLO | Čelik 1997, Čelik et al. 2009a |
| <i>Carex gracilis</i> | PL | Sielezniew et al. 2010 |
| <i>Carex tomentosa</i> | CRO | Šašić 2010 |
| <i>Carex flacca</i> | SLO | Rakar, Čelik, Vreš unpubl. |
| <i>Gratiola officinalis</i> | SLO | Čelik 1997, Čelik et al. 2009a |
| <i>Angelica sylvestris</i> | PL | Sielezniew et al. 2010 |
| <i>Inula hirta</i> | SLO | Rakar, Čelik, Vreš unpubl. |
| <i>Lotus corniculatus</i> | SLO | Rakar, Čelik, Vreš unpubl. |
| <i>Cirsium palustre</i> | PL | Sielezniew et al. 2010 |
| <i>Calluna vulgaris</i> | IT | Bonelli et al. 2010 |
| <i>Rhamnus catharticus</i> * | DE | Bräu et al. 2010 |
| <i>Cotinus coggygria</i> | SLO | Rakar, Čelik, Vreš unpubl. |

CRO = Croatia, DE = Germany, IT = Italy, PL = Poland, SLO = Slovenia

* This plant was erroneously given as *Rhamnus frangula* (*Frangula alnus*) in Bräu et al. (2010)