

AperTO - Archivio Istituzionale Open Access dell'Università di Torino

Stakeholder perceptions of manure treatment technologies in Denmark, Italy, the Netherlands and Spain

This is the author's manuscript

Original Citation:

Availability:

This version is available <http://hdl.handle.net/2318/1634987> since 2021-12-15T16:39:32Z

Published version:

DOI:10.1016/j.jclepro.2016.10.162

Terms of use:

Open Access

Anyone can freely access the full text of works made available as "Open Access". Works made available under a Creative Commons license can be used according to the terms and conditions of said license. Use of all other works requires consent of the right holder (author or publisher) if not exempted from copyright protection by the applicable law.

(Article begins on next page)

1 Stakeholder perceptions of manure treatment technologies in Denmark, Italy, the
2 Netherlands and Spain

3
4
5
6
7
8
9
10 4 Y. Hou ^{a,*}, G. L. Velthof ^b, S. D. C. Case ^c, M. Oelofse ^c, C. Grignani ^d, P. Balsari ^d, L.
11 Zavattaro ^d, F. Gioelli ^d, M. P. Bernal ^e, D. Fangueiro ^f, H. Trindade ^g, L. S. Jensen ^c, O.
12 Oenema ^b

13
14
15
16
17
18
19
20
21 8 ^a Wageningen University, Soil Quality Group, P.O. Box 47, 6700 AA, The Netherlands

22
23
24 9 ^b Wageningen University and Research Centre, Alterra, P.O. Box 47, 6700 AA, The Netherlands

25
26
27 10 ^c Dept. Plant and Environmental Sciences, Faculty of Science, University of Copenhagen, Thorvaldsensvej 40,
28 DK-1871 Frederiksberg C, Denmark

29
30
31
32 12 ^d Dept. Agricultural, Forest and Food Sciences, University of Turin, L.go Paolo Braccini, 2, 10095 Grugliasco,
33 Italy

34
35
36
37 14 ^e Dept. Soil and Water Conservation and Organic Waste Management, Centro de Edafología y Biología Aplicada
38 del Segura, CSIC, P.O. Box 164, 30100, Spain

39
40
41
42 16 ^f LEAF, Instituto Superior de Agronomia, Ulisboa, Tapada da Ajuda, 1349-017 Lisboa, Portugal

43
44
45
46 17 ^g Centre for the Research and Technology of Agro-Environment and Biological Sciences, Dept. Agronomy,
47 Universidade de Trás-os-Montes e Alto Douro, Apartado 1013, 5001-801 Vila Real, Portugal

48
49
50
51
52
53
54
55

56 *Corresponding author. Tel. +31 317485083. Fax +31 317426101

57
58 E-mail address: yong.hou@wur.nl; houyong7514364@126.com (Y. Hou).

20 **Abstract**

21 Manure treatment technologies have been developed in Europe to better use animal manures
22 and to reduce their environmental impact, but the adoption of these technologies in practice is
23 regionally diverse and still limited. Also, little is known about the opinions of stakeholders
24 towards manure treatment. This study aimed to identify stakeholder perceptions of (1) which
25 factors can facilitate and hinder the implementation in practice, (2) which technologies have
26 the most potential for successful adoption, and (3) how farm characteristics and scale of
27 treatment operations affect priorities for technology adoption. This analysis used data from a
28 survey of various stakeholders engaged in manure treatment in four European countries
29 (Denmark, Italy, the Netherlands and Spain) that have large areas of high animal density, but
30 diverse socio-economic, political and environmental conditions. Pressure from governmental
31 regulations was perceived as a key factor that stimulated manure treatment in all four
32 countries (70% of respondents). Processing manure to produce bioenergy was considered
33 important in Denmark and Italy, but less important in Spain and the Netherlands. The major
34 barriers to technology adoption were related to economic factors -lack of investment capital
35 (60% of respondents), high processing cost (52%) and a long payback period (45%), while
36 there was relatively little concern regarding transport and noise burden and health risks.
37 Slurry separation and anaerobic digestion were perceived to have the greatest potential for a
38 common adoption. Other preferred technologies were more country-specific (e.g. acidification
39 in Denmark, composting in Spain, and drying and reverse osmosis in Netherlands). Manure
40 treatment was considered to be less applicable at small livestock farms. Separation,
41 composting and acidification were perceived to be more applicable at farm scale, while drying,
42 anaerobic digestion, reverse osmosis at large, industrial scales. Our results imply that manure
43 treatment will remain a regional activity. Policy measures and outreach strategies to alleviate
44 the main barriers to the adoption of manure treatment are suggested.

45 *Keywords:* Acidification; Anaerobic digestion; Economic barriers; Environmental regulations;

46 Separation; Survey

47

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

48 **1. Introduction**

1
2
3 49 Animal manures are valuable sources of plant nutrients, soil organic matter and bioenergy.
4
5 50 However, following the introduction of relatively cheap inorganic fertilizers from the 1950s
6
7
8 51 onwards, animal manures were increasingly considered as a waste, especially in affluent
9
10 52 countries (e.g., in Europe and North America; Van der Meer, 1987). Recently, inappropriate
11
12 53 use and inefficient recycling of animal manures, particularly in regions with high animal
13
14 54 density, have exerted a series of negative impacts on the environment, e.g. eutrophication of
15
16 55 ecosystems, soil acidification and global warming (Steinfeld et al., 2006). In Europe, the
17
18 56 livestock sector is currently responsible for about 80% of total European ammonia (NH₃)
19
20 57 emissions, 10-17% of greenhouse gas (GHG) emissions, 40-50% of diffuse nitrogen (N) and
21
22 58 70% of inorganic phosphorus (P) losses to inland and coastal water (Leip et al., 2015). In
23
24 59 response, a series of governmental policies have been implemented by the European Union
25
26 60 (EU) and some of its Member States to improve the utilization of manure nutrients in
27
28 61 agriculture and therefore decrease their environment impact (Oenema et al., 2011). These
29
30 62 policies have contributed towards the development of manure treatment technologies, which
31
32 63 are important for achieving cleaner production in livestock husbandry.
33
34
35
36
37
38
39
40 64 Historically, manure has always been treated and used for various purposes. Attempts to
41
42 65 produce biogas from manure date back to the 10th century B.C. (Bond and Templeton, 2011).
43
44 66 Efforts to recover specific nutrients or to increase the agronomic value of manure date from
45
46 67 the second half of the 20th century (Van der Meer, 1987). Manure has been dried and used as
47
48 68 fuel and building material probably as long as there has been animal agriculture. A wide range
49
50 69 of new manure treatment technologies have been developed and are now available in Europe.
51
52 70 These technologies are considered to be of great importance for the development of
53
54 71 sustainable agricultural systems and societies (Foged et al., 2011a; Sommer et al., 2013).
55
56 72 Several technologies (e.g. slurry acidification, anaerobic digestion) are used to decrease
57
58
59
60
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

73 ammonia and/or GHG emissions from animal manure, and thereby decrease the risk of
74 climate change and acidification of ecosystems. Technologies have been developed to
75 produce renewable energy from manure, for instance, through anaerobic digestion (i.e. biogas
76 production) and incineration (Billen et al., 2015; Kimming et al., 2015). Manure-based
77 bioenergy production decreases CO₂ emissions by substituting fossil fuel for power and
78 electricity production, and therefore is a crucial contributor to the development of bio-
79 economy. Other technologies (e.g. solid-liquid separation, drying, composting, reverse
80 osmosis) have been developed to improve manure handling and transportation characteristics
81 (Sommer et al., 2013). In addition, various manure-based products resulting from these
82 treatment technologies provide opportunities for better nutrient management in agriculture.
83 These products may reduce unnecessary mineral fertilizer use and so the associated resource
84 use and environmental pollution from fertilizer production (Sommer et al., 2013).

85 Implementation of manure treatment technologies in practice is however limited and
86 regionally scattered in the EU. Less than 10% of the total animal manure production
87 (excluding excreta of grazing animals) was processed in the EU-27 in 2010, with large
88 variations between countries (Foged et al., 2011a). The extent to which treatment technology
89 advances in a country can be influenced by governmental policies and the perceptions of key
90 stakeholders. Environmental policies and legislations vary between EU countries. Although
91 EU Directives set the framework in which all Member States must create legislations directed
92 at civilians/industries to attain the EU-scale objectives, Member States have some flexibility
93 to implement these Directives (Oenema et al., 2011). For example, there is flexibility in the
94 design of national action programs and the use of mitigation measures and techniques in the
95 Nitrates Directive (1991/676/EC) and National Emission Ceiling Directive (2001/81/EC). In
96 addition, differences in farming systems and environmental conditions in the EU, combined
97 with the complexity of manure management and nutrient recycling, can also affect the

1
2
3
4
5 98 adoption of treatment technologies (Sommer et al., 2013). To facilitate the proper
6
7 99 development of manure treatment technology, there is a need to improve understanding of the
8
9
10
11 100 reasons for the limited and scattered implementation of these treatment technologies in
12
13 101 practice in the EU, especially in regions with high animal density.
14
15
16
17
18
19

20 102 While extensive research has been conducted to evaluate the technical, environmental and
21
22 103 economic performance of manure treatment technologies in EU, stakeholder opinions
23
24 104 regarding the factors influencing manure treatment in practice have not received significant
25
26 105 consideration. The diffusion and exploitation of cleaner technologies relies on a combination
27
28 106 of factors including governmental policies, financial incentives, technical and service support,
29
30 107 and social acceptance (Montalvo, 2008). A better understanding of needs and perceptions of
31
32 108 stakeholders from both the supply and demand side is essential to allow for successful
33
34 109 innovations for sustainable production and consumption to be shared, spread and scaled up
35
36 110 (Blok et al., 2015). The development of manure treatment involves stakeholders across
37
38 111 government, industry, academia, extension services and agricultural production sectors.
39
40 112 Integration between policy fields, expert bodies and types of expertise is increasingly required
41
42 113 in framing and assessing these EU environmental policies (Kowarsch, 2015). Stakeholders
43
44 114 from different sectors may have diverse opinions regarding the objectives of a policy measure
45
46 115 as well as on the relevant actions needed to achieve it (Petit and van der Werf, 2003; Van
47
48 116 Dam and Junginger, 2011). Policy makers and researchers generally have a broad picture of
49
50 117 environmental issues and manure management at regional and national scales. In contrast, the
51
52 118 experience of individual farmers are more tied to a particular farm environment, and their
53
54 119 decisions are shaped mostly by local socio-economic conditions (Asai et al., 2014; Ingram,
55
56 120 2008). Agricultural advisors have an fair understanding of a group of farmers and their farms
57
58 121 through regular contact, enabling them to develop a geographically broad impression of the
59
60 122 farming community (Ingram, 2008). Increased understanding among stakeholders involved in
61
62
63
64
65

123 the system can help to overcome barriers to the adoption and exploitation of manure treatment
124 technologies.

125 Few studies have been conducted to investigate stakeholder perceptions of factors influencing
126 the adoption of manure treatment technologies. Examples include studies focusing on
127 composting (Viaene et al., 2016), slurry separation (Gebrezgabher et al., 2015) and anaerobic
128 digestion (Dahlin et al., 2015; Hoppe and Sanders, 2014) in several EU countries. A study in
129 the Netherlands reported that farmer attitudes toward the various properties of manure
130 separation technology were important determinants of adoption. Farmer attitudes were
131 positive towards the agronomic attributes of separation such as the ability to use nutrients (e.g.
132 N and P) in manure optimally, but the economic benefits were generally not appreciated
133 (Gebrezgabher et al., 2015). Barriers to on-farm composting in Belgium were studied based
134 on interviews with stakeholders, which found that strict regulation, considerable financial
135 investment, and lack of experience and knowledge were hindering on-farm composting
136 (Viaene et al., 2016). An analysis of stakeholder perceptions in the biogas production chain in
137 several EU countries indicated that biogas producers and digestate suppliers face many risks
138 and challenges, primarily linked to high financial cost (and sometimes little incentives), legal
139 constraints for operation and market barriers to digestate application (Dahlin et al., 2015;
140 Hoppe and Sanders, 2014). These studies have illustrated that the adoption of manure
141 treatment technology is likely to be affected by a wide range of diverse socio-political,
142 environmental and agronomic factors. There is a need for better understanding of stakeholder
143 perceptions of factors that currently influence manure treatment and also their perspectives
144 regarding successful adoption of these technologies in future.

145 This study aimed to provide empirical insights into: (1) what stakeholders perceive as
146 important to facilitate or hinder the implementation of manure treatment in practice, (2)
147 stakeholder views of the technologies that have the most potential for successful adoption,

148 and (3) how the preference of technologies with the most potential differs between farm types,
149 farm sizes, and scale of treatment operations. To achieve these objectives, a survey of
150 stakeholders from various groups was conducted in four EU countries: Denmark, Italy, the
151 Netherlands and Spain. All selected countries have large areas of high animal density, but
152 diverse political, farming and environmental contexts.

153 **2. Methods**

154 This section includes a description of the countries surveyed (Section 2.1), stakeholder
155 categories (Section 2.2), the questionnaire structure (Section 2.3) and the methods regarding
156 data collection and analysis (Section 2.4).

157 *2.1 Country selection and context*

158 Denmark (DK), Italy (IT), the Netherlands (NL) and Spain (ES) were selected to represent
159 European countries that have highly-intensive animal production, and as a result, large
160 pressure for manure handling and management (Fig. 1). Average livestock densities are 1.9
161 and 3.6 livestock units (LU) per ha of utilized agricultural area in DK and NL, respectively
162 (compared to the EU-27 average of 0.8 LU ha⁻¹). In the north of IT (e.g. Lombardy and
163 Veneto regions) and in some regions of ES (e.g. Catalonia and Murcia regions) livestock
164 densities are also higher than 1.5 LU ha⁻¹ (Fig. 1).

165 These four countries were also selected because they vary in governmental policies, manure
166 management systems and environmental conditions (Table 1). All four countries need to
167 comply with the Nitrates Directive, which aims to protect water quality by promoting good
168 farming practices and preventing the pollution of groundwater and surface waters by nitrate
169 from agricultural sources (including animal manure). The implementation of the Nitrates
170 Directive has had a great influence on manure management (Velthof et al., 2014). The whole
171 territories of DK and NL have been designated as the so-called "Nitrate Vulnerable Zones"

172 (NVZs), while the NVZs cover approximately 21% of total agricultural area in ES and 32% in
173 Italy. Derogations have been granted for specific regions/farms in DK, IT and NL, which
174 allow them to go beyond the limit of 170 kg N ha⁻¹ of manure application, while there is no
175 derogation in ES. Renewable energy action plans differ between these countries, e.g. the use
176 of animal manures for renewable energy production (Table 1). Soil organic matter is key to
177 soil quality and productivity, and plays a major role in modifying chemical, microbiological
178 and physical properties in ways that improve soil fertility. Mean organic carbon contents in
179 the top soils are < 15 g C kg⁻¹ in most regions of ES, while > 30 g C kg⁻¹ on average in NL (de
180 Brogniez et al., 2015; Reijneveld et al., 2009). The organic carbon content of the soil may
181 affect decisions about the most suitable use of manure as a source of organic matter to
182 improve soil quality (Diacono and Montemurro, 2010). In DK and NL, manure management
183 systems of dairy cattle are dominantly slurry-based, in contrast to the large fraction of solid-
184 based systems in ES and IT (Table 1).

185 ***2.2 Stakeholder groups***

186 Six stakeholder groups with expertise in the domain of manure treatment were chosen for this
187 study: (i) livestock farmers; (ii) members of the board of farmers' organizations; (iii)
188 agricultural advisors and consultants; (iv) developers and users of treatment technologies
189 from industry (also including contractors with manure treatment facilities); (v) employees of
190 public authorities (working on the development and control of agri-environmental policies);
191 and (vi) researchers from academic institutions (with expertise in animal manure treatment)
192 (Table 2).

193 ***2.3 Questionnaire design***

194 The survey consisted of 62 questions divided into five sections. Section 1 dealt with
195 respondents' experience in manure treatment. Section 2 related to opinions on factors that

196 stimulate and hinder the implementation of manure treatment in practice. The selection of
197 these factors (presented in the questionnaire) was based on peer-reviewed studies (e.g.,
198 Gebrezgabher et al., 2015; Hoppe and Sanders, 2014; Montalvo, 2008) and views of experts
199 (including the authors) in the research of farm-based studies in the surveyed countries.
200 Section 3 aimed to investigate stakeholder opinions about the technologies that have the most
201 potential for successful adoption. Eight common treatment technologies were listed in the
202 questionnaire: solid-liquid separation, acidification, anaerobic digestion, biological nitrogen
203 removal, composting, drying, combustion/ incineration, and membrane filtration/ reverse
204 osmosis (Foged et al., 2011a). A brief description of each of these technologies is provided in
205 Appendix A. For each technology there were four follow-up questions to investigate why,
206 how and where the selected technologies had the greatest adoption potential (considering farm
207 type, farm size and scale of operation, and the benefits of each technology). Section 4
208 collected demographic information, including employment categories (to distinguish between
209 stakeholder groups) and farm characteristics (in the case of farmers). The final section
210 allowed respondents to submit any other comments and to give contact information (if they
211 wished to receive the results of the study). Respondents could write additional comments and
212 suggestions for each question (under the response ‘other’).

213 ***2.4 Data collection and analysis***

214 The survey was performed through both face-to-face interviews and online questionnaires,
215 with support from the academic institutions that participated in the joint EU project
216 ReUseWaste¹. The electronic version of the questionnaire was designed using
217 SurveyMonkeyTM. The questionnaire used for face-to-face interviews was the same as that
218 used for the online survey. Data were collected between April 2014 and June 2015.

¹ ReUseWaste: <http://www.reusewaste.eu/>

219 Survey dissemination strategies differed between countries. In DK, surveys were
220 disseminated by researchers from the University of Copenhagen via an email that described
221 the purpose and background context of the survey, and included a link to the online survey.
222 Agricultural advisors were contacted via a database of advisors obtained from the Danish
223 agricultural extension service (110 advisors were randomly selected; 32 of them completed
224 the survey, Table 2). A list of other stakeholders was prepared via personal contacts. For
225 instance, the questionnaires were emailed to 18 researchers with expertise in manure
226 management and treatment (in Aarhus University, University of South Denmark and the
227 University of Copenhagen), 20 officers in local and national governmental department (e.g.
228 the Danish Environmental Production Agency, the Danish AgriFish Agency), and to the
229 chairmen of 45 farmers' organizations in DK. Similarly to DK, all surveys were disseminated
230 via email and completed online in NL. Requests were sent to target stakeholders (except for
231 researchers) via the secretaries of two large (branch) organizations i.e., CUMELA and
232 Nutrient Platform, and of the main farmers' organization LTO.² Furthermore, a selection of
233 20 researchers from Wageningen UR with expertise in manure management and treatment
234 were asked to complete the questionnaire. In ES, the questionnaire was completed online by
235 stakeholders from the research, industry, extension service and policy communities who were
236 selected and contacted by researchers from the Spanish National Research Council (CSIC).
237 The questionnaire was completed via face-to-face interviews with farmers, instead of by an
238 online survey, as it was considered that farmers would generally have limited access to the
239 internet and were not familiar with online questionnaires. Livestock farmers were visited at
240 their homes (one by one) in regions of high livestock density (Murcia and Catalonia) by
241 researchers from CSIC. These farmers were selected via the contact of local agricultural
242 advisors and also according to their willingness to participate. A hard copy of the

² CUMELA: <http://www.cumela.nl/>, Nutrient Platform: <http://www.nutrientplatform.org/>, LTO: <http://www.lto.nl/>.

243 questionnaire was presented to the respondents. Interviewers were instructed not to present
1
2 244 their own opinions, but only to clarify the questions in case farmers did not understand.
3
4 245 Results of the face-to-face interviews were uploaded to the SurveyMonkey™ database and
5
6
7 246 analyzed alongside online responses. In IT, stakeholders from the research, industry,
8
9
10 247 extension service, and policy communities were surveyed during two national agricultural
11
12 248 meetings held in 2014 (November and December), and further interviews were subsequently
13
14 249 conducted via personal contacts of researchers from the University of Turin. Efforts were
15
16
17 250 made to ensure the privacy of the face-to-face interviews, and respondents were interviewed
18
19 251 one by one. Respondents representing livestock farmers and members of farmers’
20
21
22 252 organizations came mainly from areas where animal husbandry is highly intensive, i.e.
23
24 253 Northern Italy (Piedmont, Lombardy, Emilia Romagna and Veneto). In total across all the
25
26
27 254 countries 291 surveys were completed; each stakeholder group had between 18 and 75
28
29 255 respondents (see details in Table 2 and Section 3).

30
31
32 256 A draft of the survey was sent to researchers (more than 20 in total) in the four countries to
33
34
35 257 improve clarity and reduce the chance of misinterpretation. Comments and suggestions on the
36
37 258 draft questionnaires were used to modify the survey before distribution. The same survey was
38
39
40 259 disseminated to the four target countries, but translated (into Danish, Dutch, Italian and
41
42 260 Spanish). The English version of the questionnaire is provided in the supplementary material
43
44
45 261 associated with this article.

46
47
48 262 Data downloaded from the SurveyMonkey™ were compiled and analyzed using R version
49
50 263 3.0.0 (e.g. Crosstab function) and Microsoft Excel 2010. The number of positive ticks to each
51
52 264 option of a question (i.e. the number of respondents) was recorded. Results were analyzed by
53
54
55 265 individual countries and also with the sum of all countries. Since there were multiple-response
56
57
58 266 questions in the questionnaire, the absolute number of respondents referring to each answer of
59
60 267 a question was converted to the percentage of the total number of respondents who answered
61
62
63
64
65

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
268 the question. This conversion allowed for the comparison of different variables listed in a
269 question, as well as a comparison between countries.

270 **3. Results**

271 Table 2 provides an overview of the number of respondents per stakeholder group and
272 country. In total, 291 questionnaires were completed: 28% in DK, 22% in ES, 23% in IT, and
273 27% in NL. A total of 82% of the respondents had experience with manure treatment (Table
274 2). More than 50% of those had experience with manure separation and anaerobic digestion,
275 except for respondents in ES (Fig. 2). Over 70% of respondents in DK had experience with
276 slurry acidification. In ES, most respondents (40%) had experience with composting.
277 Respondents from NL had more experience with manure drying and membrane filtration (or
278 reverse osmosis) (Fig. 2). Few respondents answered that they had experience with alternative
279 treatment technologies that were not offered as possible responses in the question, e.g.
280 ammonia stripping from liquid manure, phosphorus recovery, or evaporation of liquid manure.

281 **3.1 Factors that stimulate and hinder adoption**

282 Pressure from environmental policies was perceived to be the most important factor affecting
283 the implementation of manure treatment in practice (70% of total respondents), which was the
284 case for respondents from all four countries and all stakeholder groups (Fig. 3; Appendix B.1).
285 The need to facilitate the export of manure from the farm (47%, especially in DK and NL)
286 was also highlighted by many respondents. The need to achieve renewable energy targets by
287 producing bioenergy from manure was ranked relatively highly in DK and IT. Compared to
288 the other factors considered, the need to efficiently use manure nutrients due to increased
289 fertilizer prices was considered relatively important in ES. For all countries, controlling
290 diseases, pathogens and odor was considered the least important among the six factors defined
291 in the survey (Fig. 3).

292 Economic factors were the main barriers to the implementation of manure treatment in
293 practice, namely the lack of investment capital (60% of total respondents), high processing
294 costs (52%), and a long payback period (45%). These barriers were perceived to be important
295 for all countries (Fig. 4) and by all stakeholder groups (Appendix B.2). Legal constraints (32%
296 of all respondents, highest at 45% in NL) and lack of knowledge (32% of all respondents,
297 especially in ES and IT) were chosen by a moderate number of respondents. Transport, noise
298 burdens and health risks were not seen as important barriers among all stakeholder groups
299 (Fig. 4; Appendix B.2). Interestingly, livestock farmers and agriculture advisors had relatively
300 little concern about the market for manure processing products (Appendix B.2). This
301 suggested that these farmers were possibly interested in using processed organic fertilizers,
302 which is confirmed by the results from a parallel study on farmer perceptions of organic
303 fertilizers in Denmark (Case et al., unpublished results).

304 **3.2 Preferred treatment technologies**

305 Stakeholders indicated that manure separation and anaerobic digestion had the greatest
306 potential for a common adoption in practice (36% and 42% of total respondents, respectively).
307 Other technologies appear to be more country specific. There was a relatively high adoption
308 potential for slurry acidification in DK (47%) and composting in ES (44%), while drying of
309 solid manure fractions and membrane filtration (or reverse osmosis) of liquid fractions were
310 considered positively in NL (Fig. 5).

311 **3.3 Preferred farm structure and scale of operation**

312 Fig. 6 shows that livestock farms with a limited area of land were considered to have a
313 relatively high adoption potential for all of the manure treatment technologies considered with
314 the exception of slurry acidification (Fig. 6a). This exception is possibly due to the fact that
315 farms with sufficient land are more willing, or are required to use techniques that reduce

1
2 316 ammonia losses from on-farm storage and application of manures. Overall, manure treatment
3 317 was considered to be more applicable to pig and cattle farms than to poultry farms (Fig. 6a).

4
5 318 Manure treatment was considered to be less applicable to small livestock farms (i.e. <50 LU).

6
7 319 Drying and reverse osmosis technologies were perceived most appropriate for large livestock
8 320 farms (> 1000 LU) (Fig. 6b).

9
10 321 Stakeholders had different views regarding the optimal scale of the manure treatment plant
11 322 (Fig. 6c). Separation (67% of respondents), acidification (55%) and composting (52%) were
12
13 323 perceived to be most applicable at the farm scale. Anaerobic digestion, drying (pelletizing)
14 324 and membrane filtration were considered to be most applicable at the industrial scale and for
15
16 325 farmer cooperatives (Fig. 6c).

17 18 19 20 21 22 23 24 25 26 326 **3.4 Benefits of manure treatment**

27
28
29 327 Table 3 shows respondent perceptions of the benefits of manure treatment. A reduction in
30 328 manure disposal costs and an increase in the fertilizer value of separated liquid and solid
31
32 329 fractions were ranked as the main benefits of manure separation. For anaerobic digestion, the
33
34 330 main benefits included bioenergy production, the increased fertilizer nitrogen value of
35 331 digestate, and the reduction of odor and gaseous emissions during further processing and field
36
37 332 application. Mitigation of ammonia emissions during slurry storage and application, and the
38
39 333 increased fertilizer N value of slurry were ranked as the main benefits of slurry acidification.
40
41 334 Increased organic matter quality of manure and improved soil quality after field application
42
43 335 were ranked as the main benefits of composting.

44 45 46 47 48 49 50 51 52 336 **4 Discussion**

53
54
55 337 Currently, less than 10% of the animal manure produced in EU is treated and most farmers
56
57 338 have little knowledge about manure processing technologies (Foged et al., 2011a). In the
58
59
60
61
62
63
64
65

339 present study the survey was disseminated to stakeholders involved in manure treatment,
340 directly or indirectly. This explains why over 80% of the respondents described themselves as
341 having at least some experience with manure treatment. Most of the stakeholders contacted
342 within each group (farmers, farmers' organizations, extension service, industry, policy and
343 research) were considered to be forerunners in the whole domain of the development,
344 implementation and management of manure treatment technologies. By exploring the views
345 of these stakeholders engaged with manure treatment, a better understanding of the future
346 perspectives of manure processing is possibly achieved.

347 **4.1 Key factors that stimulate manure treatment in practice**

348 Pressure from environmental policies and regulations was identified as the most important
349 stimulus for the implementation of manure treatment systems (Fig. 3). This may reflect the
350 fact that current policies and regulations implemented in these four countries have influenced
351 stakeholder decisions on manure handling and management activities. A number of policies
352 have been implemented by the EU and United Nations (UN) bodies to reduce environmental
353 pollution from animal manures (Oenema et al., 2011), which play an important role in
354 stimulating manure treatment activities in Europe. The EU Nitrates Directive sets up the
355 maximum application limit of manure in NVZs, equivalent to 170 kg N ha⁻¹ year⁻¹ (European
356 Commission, 1991). This limit obliges livestock farms to treat and/or to transport the excess
357 manure to other farms. The EU National Emission Ceiling Directive (European Commission,
358 2001) aims to reduce emissions of ammonia (NH₃) (including from manures), and thereby
359 stimulate the development of certain manure treatment technologies (Bittman et al., 2014).
360 For example, acidifying slurry was introduced as one of the options for obligatory NH₃
361 mitigation measures by Danish regulations in response to these EU Directives. On the other
362 hand, the use of manure treatment may remain marginal in regions that have low pressure
363 from these regulations. The authors conducted also a similar survey in Portugal, but the

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

364 number of responses from targeted stakeholders was small and hence results are not shown. A
365 low response rate from Portuguese stakeholders (in particular farmers) may reflect that the
366 interest for manure treatment is low in regions that have sufficient land for application of the
367 manure produced, as well as low pressure from governmental legislation. These results
368 revealed that variations between countries in manure treatment have a strong relationship with
369 variations in livestock density and national policies.

370 Producing bioenergy from animal manures was identified as an important reason for the
371 adoption of manure treatment in practice, in particular in DK and IT (Fig. 3). Anaerobic
372 digestion produces biogas that can be used directly for heating, for combined thermal and
373 electricity generation, or to upgrade to bio-methane that has similar characteristics to natural
374 gas (Bernet and Béline, 2009). Using animal manures as feedstock for biogas production has
375 advantages compared to using energy crops, such as less competition with food production
376 and higher mitigation potential of greenhouse gas emissions (De Vries et al., 2012). Further,
377 the digestate can serve as an improved organic N fertilizer (Table 3). The development of
378 biogas production in European countries has been influenced strongly by environmental
379 regulations and the EU Renewable Energy Directive (Edwards et al., 2015). The growth of
380 anaerobic digestion in DK is largely due to policy incentives such as increased investment
381 support for construction of biogas plants, the implementation of fossil energy taxes or
382 renewable energy tariff subsidies and the government support strategies to increase
383 interactions between various social groups (Raven and Gregersen, 2007). Italy has also
384 witnessed an extraordinary growth in biogas generation from animal manures and other
385 agricultural biomass in the last few years, which is largely due to the biogas support programs
386 implemented in Italy (the introduction of Tradable Green Certificate and feed-in-tariff, and
387 increased investment subsidies) (Chinese et al., 2014). In comparison, manure-based biogas
388 producers in NL and ES face many financial and socio-political challenges (Fierro et al., 2014;

389 Hoppe and Sanders, 2014), which may explain why biogas production was perceived as less
1
2 390 attractive in these two countries (Fig.3 and Fig. 5).
3
4

5 391 The pressure from increased fertilizer price was perceived to be important factor for
6
7
8 392 stimulating manure treatment in ES (Fig. 3), which is in line with the conclusion from a study
9
10 393 that investigated the existing experience on manure treatment in Catalonia, a region with high
11
12 394 animal density in ES (Flotats et al., 2009). The increase in prices of mineral fertilizers could
13
14 395 explain the recent growth in composting facilities in Catalonia, in order to recover nutrients in
15
16 396 organic forms and produce soil organic amendments that are economically valuable (Flotats et
17
18 397 al., 2009). The need to facilitate off-farm manure export was considered to be relatively
19
20 398 important in NL and DK, where the average LU is high and a large portion of farms have
21
22 399 been involved in manure exchange (Asai et al., 2014); it appears to be less important in ES
23
24 400 partly due to the average low animal density (Fig. 1).
25
26
27
28
29
30

31 401 **4.2 Key barriers to manure treatment in practice**

32
33

34 402 The most important barriers to the implementation of manure treatment in practice were
35
36 403 related to economic factors (Fig. 4). This corresponded with findings from several other
37
38 404 studies. Results from a survey among 111 Dutch dairy farmers indicated that nearly half of
39
40 405 respondents strongly disagreed with the statement that low cost of manure separation is a
41
42 406 reason for them to consider the use of manure separation, while only 13% of respondents
43
44 407 agreed (Gebrezgabher et al., 2015). Substantial upfront investments, subsidies not being
45
46 408 granted, and increased price of co-feedstock were identified as important barriers for biogas
47
48 409 producers in NL (Hoppe and Sanders, 2014). In the present study, most respondents (who
49
50 410 perceived that anaerobic digestion had the most potential for adoption) stated that subsidies
51
52 411 for upfront investment and/or energy production were vital for anaerobic digestion of animal
53
54 412 slurries in practice (data not shown). This confirms results from previous studies that
55
56
57
58
59
60
61
62
63
64
65

1
2 413 subsidies play a large role in the profitability of biogas plants (Chinese et al., 2014;
3 414 Gebrezgabher et al., 2010; Riva et al., 2014).

4
5 415 A number of respondents brought up legal constraints as an important issue hindering the
6
7
8 416 implementation of manure treatment (Fig. 3). A Dutch respondent indicated that “Licensing
9
10 417 can be very restrictive in realizing initiatives, due to lack of objective knowledge (on manure
11
12 418 processing) among local residents and licensing authorities”. Likewise, a stakeholder study
13
14 419 indicated that legal permits to operate biogas plants were difficult to attain in NL, partly
15
16 420 because municipalities did not yet have specific biogas polices in place and therefore there
17
18 421 were few staff trained in how to deal with permit requests for co-digestion plants (Hoppe and
19
20 422 Sanders, 2014). A Danish respondent also stated that “It is difficult or impossible to get
21
22 423 authority approval for treatment operations, because of the resistance of the local community”.
23
24 424 Therefore, outreach strategies should be developed to provide more information to local
25
26 425 residents, authorities, and extension services regarding the benefits and risks of manure
27
28 426 treatment so as to increase social acceptability.
29
30
31
32
33
34

35 427 **4.3 Differences in priorities of technology adoption and operation structure**

36
37
38
39 428 The choice of prioritized technologies generally corresponded with the technologies for which
40
41 429 respondents had experience (see Fig. 2) and the status of manure processing activities in the
42
43 430 countries surveyed (Foged et al., 2011a). An EU inventory study reported that slurry
44
45 431 separation was used most in IT and ES; anaerobic digestion was predominantly applied in
46
47 432 Germany, followed by IT and DK; and slurry acidification was mainly adopted in DK, while
48
49 433 ES had the largest share of composting operations (Foged et al., 2011a). Slurry acidification
50
51 434 in DK is typically applied to raw animal slurries either in the animal house (thus reducing
52
53 435 emissions from both housing, slurry storage and field application) or immediately prior to
54
55 436 land application (in the slurry storage or on the slurry tanker during field application, thus
56
57
58
59
60
61
62
63
64
65

1 437 only reducing emissions in the field) (Fangueiro et al., 2015; Kai et al., 2008). In this study,
2 438 composting was identified to have considerable growth potential in ES (Fig. 5). This is partly
3
4 439 because of the low soil organic matter content of arable land in ES ($< 15 \text{ g C kg}^{-1}$; de
5
6
7 440 Brogniez et al., 2015) and the ability to improve soil quality following the application of
8
9 441 compost (Bernal et al., 2009). Composting was not ranked highly in DK and NL, where soil
10
11 442 organic matter contents are relatively high (de Brogniez et al., 2015). Solid-liquid separation,
12
13 443 drying of solid fractions and reverse osmosis of liquid fractions (to concentrates) were
14
15
16 444 considered as attractive technologies for livestock farms with a limited area of land in NL
17
18 445 (Fig. 5). This may have been chosen due to the need to comply with policy regulations.
19
20
21 446 Obligatory manure treatment was introduced in NL in 2013, which designated that livestock
22
23 447 farms with a manure surplus have to treat and/or to export a certain percentage of the surplus.
24
25
26 448 Thus, the need to transport manures can greatly increase the use of treatment technologies that
27
28 449 reduce the volume of liquid (separation and reverse osmosis) and solid fractions (drying and
29
30 450 pelletizing). Manure-based anaerobic digestion was prioritized in DK (Fig. 5), mainly due to
31
32
33 451 Danish government policy. The Danish government proposed a target of using 50% of the
34
35
36 452 manure produced in DK for renewable energy by 2020, which would need to be met through
37
38 453 a strong expansion of biogas plants and capacity (Danish Ministry of Food, Agriculture and
39
40 454 Fisheries, 2009).

41
42
43
44 455 Farm size and treatment plant operation structure are important for the implementation of
45
46 456 manure treatment technologies (Flotats et al., 2009; Gebrezgabher et al., 2015). Clearly, land-
47
48 457 limited large farms with $>300 \text{ LU}$ (representing farms with high animal density) have larger
49
50
51 458 potential (or need) for the adoption of manure treatment than small farms (Fig. 6). Separation
52
53 459 and composting were generally considered to be farm-scale treatment technologies, while
54
55
56 460 manure drying and reverse osmosis were considered most applicable at large, industrial scales
57
58 461 (Fig. 6). The complexity of the management and the costs of investments and processing

462 varied among treatment technologies (Foged et al., 2011b). This may explain why the
463 potential and suitability of technology adoption is related to the scale of farm and plant
464 operations. Solid-liquid mechanical separation and composting are generally considered to be
465 less complex in operation and of relatively low cost, compared to treatments such as
466 anaerobic digestion and reverse osmosis (Flotats et al., 2009; Foged et al., 2011b). The annual
467 total costs of processing (i.e. the total capacity and operational costs excluding subsidies) can
468 vary from 0.5-3 € t⁻¹ of inputs for mechanical separation and slurry acidification to 8-14 € t⁻¹
469 for anaerobic digestion and reverse osmosis; the net costs of processing (i.e. the total costs
470 minus the income from the sale and use of processed products and subsidies) vary from 0.5-8
471 € t⁻¹ of inputs, or on the basis of total N treated, 0.15-3 € kg⁻¹ of N (Foged et al., 2011b; Møller
472 et al., 2000). Processing manure in a cooperative form has advantages to reduce financial
473 risks (to individual farmers) and treatment costs, and to make manure treatment viable for
474 small- and mid-sized farms (Flotats et al., 2009; Møller et al., 2000; Swindal et al., 2010).

475 **5 Conclusions and recommendations**

476 Understanding the opinions of stakeholders closely engaged in manure treatment can enhance
477 the effectiveness of programs designed to stimulate diffusion and exploitation of these
478 technologies. Such an understanding is an essential part of attaining EU environmental and
479 renewable energy targets. Based on the main findings from the present study, policy
480 requirements, outreach strategies and future research needs are suggested.

481 *Policy requirements.* Pressure from governmental legislation was identified as the key
482 stimulant of technology adoption, while barriers to adoption were mainly related to economic
483 factors. It is recommended that policies for the promotion of manure treatment must be
484 economically appealing to attract new adopters (farmers and industries). Long-term financial
485 support schemes (e.g. subsidies) must be developed to encourage potential adopters to invest,

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

486 considering the long-term investment requirements of manure treatment. It is also necessary
487 to improve permit request procedures to facilitate their acquisition for operations. Large
488 variations in technology preference between countries, farm types and scale of operation were
489 observed in this study. These variations need to be considered when developing policy
490 support schemes and marketing strategies.

491 *Outreach strategies.* More information should be conveyed to livestock farmers (especially
492 those with large, land-limited farms) and other technology users regarding the different
493 aspects of a specific technology, i.e. financial viability, optimal operation conditions (e.g.
494 farm size, operation scale), regulations and incentives, and the agronomic and environmental
495 performance of the technology. Better dissemination of this information to users would
496 alleviate the lack of knowledge and experience and thus to assist with their decisions on
497 technology adoption. Resources should be allocated to enable face-to-face, direct mail contact,
498 as well as internet sources for dissemination of information. Outreach strategies need to be
499 developed to convey these important environmental benefits of manure treatment to local
500 residents so as to increase social acceptability.

501 *Future research needs.* This study emphasizes the importance of understanding stakeholder
502 perceptions in countries with large areas of high animal density where manure treatment
503 should be prioritized. However, manure treatment should not be limited to these regions,
504 considering the potential benefits of manure treatment (e.g. not only environmental but also
505 agronomic benefits). Thus, future research addressing the perceptions of stakeholders in
506 regions with contrasting farming systems and socio-political conditions will complement the
507 present findings and provide a more complete picture of the development of manure treatment.
508 Understanding stakeholder opinions about the development of manure treatment can assist in
509 the design of policies and outreach strategies, leading to a better use of animal manures and a

1
2 510 sustainable production and management chain. The results from this study can serve as a basis
3 for such efforts.
4

5
6 512 **Acknowledgments**
7

8
9 513 This research has received funding from the People Programme (Marie Curie Actions) of the
10 European Union's Seventh Framework Programme FP7/2007-2013/under REA grant
11
12 514 agreement no 289887. The results and conclusions achieved reflect only the authors' view and
13
14 515 the Union is not liable for any use that may be made of the information contained therein.
15
16
17
18
19

20 517 **Appendix A:..**
21
22

23
24 518 **Appendix B:..**
25
26

27
28 519 **Appendix C. Supplementary data**
29
30

31 520
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

521 **References**

- 1 522 Adhikari, K., Hartemink, A.E., Minasny, B., Bou Kheir, R., Greve, M.B., Greve, M.H., 2014.
2 523 Digital mapping of soil organic carbon contents and stocks in Denmark. *PLoS One* 9.
3 524 doi:10.1371/journal.pone.0105519
- 4 525 Asai, M., Langer, V., Frederiksen, P., Jacobsen, B.H., 2014. Livestock farmer perceptions of
5 526 successful collaborative arrangements for manure exchange: A study in Denmark. *Agric.
6 527 Syst.* 128, 55–65. doi:10.1016/j.agsy.2014.03.007
- 7 528 Bernal, M.P., Alburquerque, J.A., Moral, R., 2009. Composting of animal manures and
8 529 chemical criteria for compost maturity assessment. A review. *Bioresour. Technol.* 100,
9 530 5444–5453. doi:10.1016/j.biortech.2008.11.027
- 10 531 Bernet, N., Béline, F., 2009. Challenges and innovations on biological treatment of livestock
11 532 effluents. *Bioresour. Technol.* 100, 5431–5436. doi:10.1016/j.biortech.2009.02.003
- 12 533 Billen, P., Costa, J., Van der Aa, L., Van Caneghem, J., Vandecasteele, C., 2015. Electricity
13 534 from poultry manure: a cleaner alternative to direct land application. *J. Clean. Prod.* 96,
14 535 467–475. doi:10.1016/j.jclepro.2014.04.016
- 15 536 Bittman, S., Dedina, M., Howard, C., Oenema, O., Sutton, M., 2014. Options for Ammonia
16 537 Mitigation: Guidance from the UNECE Task Force on Reactive Nitrogen. Centre for
17 538 Ecology and Hydrology, Edinburgh, UK.
- 18 539 Blok, V., Long, T.B., Gaziulusoy, A.I., Ciliz, N., Lozano, R., Huisingh, D., Csutora, M., Boks,
19 540 C., 2015. From best practices to bridges for a more sustainable future: Advances and
20 541 challenges in the transition to global sustainable production and consumption:
21 542 Introduction to the ERSCP stream of the Special volume. *J. Clean. Prod.* 108, 19–30.
22 543 doi:10.1016/j.jclepro.2015.04.119
- 23 544 Bond, T., Templeton, M.R., 2011. History and future of domestic biogas plants in the
24 545 developing world. *Energy Sustain. Dev.* 15, 347–354. doi:10.1016/j.esd.2011.09.003
- 25 546 Chinese, D., Patrizio, P., Nardin, G., 2014. Effects of changes in Italian bioenergy promotion
26 547 schemes for agricultural biogas projects: Insights from a regional optimization model.
27 548 *Energy Policy* 75, 189–205. doi:10.1016/j.enpol.2014.09.014
- 28 549 Dahlin, J., Herbes, C., Nelles, M., 2015. Biogas digestate marketing: Qualitative insights into
29 550 the supply side. *Resour. Conserv. Recycl.* 104, 152–161.
30 551 doi:10.1016/j.resconrec.2015.08.013
- 31 552 Danish Ministry of Food, Agriculture and Fisheries, 2009. *Aftale om Grøn Vækst*.
32 553 Copenhagen, Denmark. [http://mst.dk/virksomhed-myndighed/landbrug/politiske-](http://mst.dk/virksomhed-myndighed/landbrug/politiske-aftaler/groen-vaekst/)
33 554 [aftaler/groen-vaekst/](http://mst.dk/virksomhed-myndighed/landbrug/politiske-aftaler/groen-vaekst/) [accessed 06.06.15]
- 34 555 de Brogniez, D., Ballabio, C., Stevens, A., Jones, R.J.A., Montanarella, L., van Wesemael, B.,
35 556 2015. A map of the topsoil organic carbon content of Europe generated by a generalized
36 557 additive model. *Eur. J. Soil Sci.* 66, 121–134. doi:10.1111/ejss.12193
- 37 558 De Vries, J.W., Vinken, T.M.W.J., Hamelin, L., De Boer, I.J.M., 2012. Comparing
38 559 environmental consequences of anaerobic mono- and co-digestion of pig manure to
39 560 produce bio-energy--a life cycle perspective. *Bioresour. Technol.* 125, 239–48.
40 561 doi:10.1016/j.biortech.2012.08.124
- 41 562 Diacono, M., Montemurro, F., 2010. Long-term effects of organic amendments on soil
42 563 fertility. A review. *Agron. Sustain. Dev.* 30, 401–422. doi:10.1051/agro/2009040
- 43 564 Edwards, J., Othman, M., Burn, S., 2015. A review of policy drivers and barriers for the use
44 565 of anaerobic digestion in Europe, the United States and Australia. *Renew. Sustain.*

- 566 Energy Rev. 52, 815–828. doi:10.1016/j.rser.2015.07.112
- 1 567 European Commission, 1991. Council directive of 12 December 1991 concerning the
2 568 protection of waters against pollution caused by nitrates from agricultural sources. Off. J.
3 569 Eur. Communisison 1–8.
- 5 570 European Commission, 2001. Directive 2001/81/EC of the European Parlia-
6 571 Council of 23 October 2001 on national emission ceilings for certain atmospheric
8 572 pollutants. Off J Eur Union L 309:22.
- 9 573 Eurostat. Online Database at: <http://epp.eurostat.ec.europa.eu>. [accessed 01.02.15]
- 11 574 Fangueiro, D., Hjorth, M., Gioelli, F., 2015. Acidification of animal slurry– a review. J.
12 575 Environ. Manage. 149, 46–56. doi:10.1016/j.jenvman.2014.10.001
- 14 576 Fierro, J., Gómez, X., Murphy, J.D., 2014. What is the resource of second generation gaseous
15 577 transport biofuels based on pig slurries in Spain? Appl. Energy 114, 783–789.
16 578 doi:10.1016/j.apenergy.2013.08.024
- 18 579 Flotats, X., Bonmatí, A., Fernández, B., Magrí, A., 2009. Manure treatment technologies: on-
19 580 farm versus centralized strategies. NE Spain as case study. Bioresour. Technol. 100,
20 581 5519–5526. doi:10.1016/j.biortech.2008.12.050
- 22 582 Foged, H.L., Flotats, X., Blasi, A.B., Palatsi, J., Magri, A., Schelde, K.M., 2011a. Inventory
23 583 of manure processing activities in Europe. Technical Report No. I concerning 'Manure
24 584 Processing Activities in Europe' to the European Commission, Directorate General
25 585 Environment. 138 pp.
- 27 586 Foged, H.L., Flotats, X., Blasi, A.B., Schelde, K.M., Palatsi, J., Magri, A., Juznik., Z., 2011b.
28 587 Assessment of economic feasibility and environmental performance of manure
29 588 processing technologies. Technical Report No. IV concerning 'Manure Processing
30 589 Activities in Europe' to the European Commission, Directorate General Environment.
31 590 130 pp.
- 34 591 Gebrezgabher, S.A., Meuwissen, M.P.M., Kruseman, G., Lakner, D., Oude Lansink, A.G.J.M.,
35 592 2015. Factors influencing adoption of manure separation technology in the Netherlands.
36 593 J. Environ. Manage. 150, 1–8. doi:10.1016/j.jenvman.2014.10.029
- 38 594 Gebrezgabher, S.A., Meuwissen, M.P.M., Prins, B.A.M., Lansink, A.G.J.M.O., 2010.
39 595 Economic analysis of anaerobic digestion—A case of Green power biogas plant in The
40 596 Netherlands. NJAS - Wageningen J. Life Sci. 57, 109–115.
41 597 doi:10.1016/j.njas.2009.07.006
- 43 598 Hoppe, T., Sanders, P.T.M., 2014. Agricultural green gas demonstration projects in the
44 599 Netherlands. A stakeholder analysis. Environ. Eng. Manag. J. 13, 3083–3096.
- 46 600 Ingram, J., 2008. Are farmers in England equipped to meet the knowledge challenge of
47 601 sustainable soil management? An analysis of farmer and advisor views. J. Environ.
48 602 Manage. 86, 214–228. doi:10.1016/j.jenvman.2006.12.036
- 50 603 Kai, P., Pedersen, P., Jensen, J.E., Hansen, M.N., Sommer, S.G., 2008. A whole-farm
51 604 assessment of the efficacy of slurry acidification in reducing ammonia emissions. Eur. J.
52 605 Agron. 28, 148–154. doi:10.1016/j.eja.2007.06.004
- 54 606 Kimming, M., Sundberg, C., Nordberg, Å., Baky, A., Bernesson, S., Hansson, P.-A., 2015.
55 607 Replacing fossil energy for organic milk production – potential biomass sources and
56 608 greenhouse gas emission reductions. J. Clean. Prod. 106, 400–407.
57 609 doi:10.1016/j.jclepro.2014.03.044
- 59 610 Kowarsch, M., 2015. COMMENTARY : Policy assessments to enhance EU scientific advice.

- 611 Nat. Clim. Chang. 6, 1–5. doi:10.1038/nclimate2835
- 1 612 Leip, A., Billen, G., Garnier, J., Grizzetti, B., Lassaletta, L., Reis, S., Simpson, D., Sutton,
2 613 M.A., de Vries, W., Weiss, F., Westhoek, H., 2015. Impacts of European livestock
3 614 production: nitrogen, sulphur, phosphorus and greenhouse gas emissions, land-use, water
4 615 eutrophication and biodiversity. *Environ. Res. Lett.* 10, 115004. doi:10.1088/1748-
5 616 9326/10/11/115004
- 6 617 Møller, H., Lund, I., Sommer, S., 2000. Solid–liquid separation of livestock slurry: efficiency
7 618 and cost. *Bioresour. Technol.* 74, 223–229. doi:10.1016/S0960-8524(00)00016-X
- 8 619 Montalvo, C., 2008. General wisdom concerning the factors affecting the adoption of cleaner
9 620 technologies: a survey 1990-2007. *J. Clean. Prod.* 16, 7–13.
10 621 doi:10.1016/j.jclepro.2007.10.002
- 11 622 Nolan, T., Troy, S.M., Gilkinson, S., Frost, P., Xie, S., Zhan, X., Harrington, C., Healy, M.G.,
12 623 Lawlor, P.G., 2012. Economic analyses of pig manure treatment options in Ireland.
13 624 *Bioresour. Technol.* 105, 15–23. doi:10.1016/j.biortech.2011.11.043
- 14 625 Oenema, O., Bleeker, A., Braathen, N.A., Budňáková, M., Bull, K., Geupel, M., Hicks, K.,
15 626 Hoft, R., Kozlova, N., Leip, A., Valli, L., Velthof, G., Winiwarter, W., 2011. Nitrogen in
16 627 current European policies, in: Sutton, M.A., Howard, C.M., Erisman, J.W., Billen, G.,
17 628 Bleeker, A., Grennfelt, P., Grinsven, H. van, Grizzetti, B. (Eds.), *The European Nitrogen*
18 629 *Assessment*. Cambridge University Press, Cambridge, UK, pp. 62–81.
- 19 630 Petit, J., van der Werf, H.M.G., 2003. Perception of the environmental impacts of current and
20 631 alternative modes of pig production by stakeholder groups. *J. Environ. Manage.* 68, 377–
21 632 386. doi:10.1016/S0301-4797(03)00105-1
- 22 633 Raven, R.P.J.M., Gregersen, K.H., 2007. Biogas plants in Denmark: successes and setbacks.
23 634 *Renew. Sustain. Energy Rev.* 11, 116–132. doi:10.1016/j.rser.2004.12.002
- 24 635 Reijneveld, A., van Wensem, J., Oenema, O., 2009. Soil organic carbon contents of
25 636 agricultural land in the Netherlands between 1984 and 2004. *Geoderma* 152, 231–238.
26 637 doi:10.1016/j.geoderma.2009.06.007
- 27 638 Riva, C., Schievano, A., D’Imporzano, G., Adani, F., 2014. Production costs and operative
28 639 margins in electric energy generation from biogas. Full-scale case studies in Italy. *Waste*
29 640 *Manag.* 34, 1429–1435. doi:10.1016/j.wasman.2014.04.018
- 30 641 Steinfeld, H., Gerber, P., Wassenaar, T., Castel, V., Roslaes, M., De Haan, C., 2006.
31 642 *Livestock’s long shadow*. Environ. issues options. FAO report, Rome, Italy. 390 pp.
- 32 643 Sommer, S.G., Christensen, M.L., Schmidt, T., Jensen, L.S, 2013. *Animal Manure Recycling:*
33 644 *Treatment and Management*. first ed. West Sussex, UK.
- 34 645 Swindal, M.G., Gillespie, G.W., Welsh, R.J., 2010. Community digester operations and dairy
35 646 farmer perspectives. *Agric. Human Values* 27, 461–474. doi:10.1007/s10460-009-9238-
36 647 1
- 37 648 Van Dam, J., Junginger, M., 2011. Striving to further harmonization of sustainability criteria
38 649 for bioenergy in Europe: Recommendations from a stakeholder questionnaire. *Energy*
39 650 *Policy* 39, 4051–4066. doi:10.1016/j.enpol.2011.03.022
- 40 651 Van Der Meer, H.G., 1987. *Animal Manure on Grassland and Fodder Crops: Fertilizer or*
41 652 *Waste?*, Martinus Nijhoff, Dordrecht
- 42 653 Velthof, G.L., Lesschen, J.P., Webb, J., Pietrzak, S., Miatkowski, Z., Pinto, M., Kros, J.,

654 Oenema, O., 2014. The impact of the Nitrates Directive on nitrogen emissions from
1 655 agriculture in the EU-27 during 2000-2008. *Sci. Total Environ.* 468-469, 1225–1233.
2 656 doi:10.1016/j.scitotenv.2013.04.058
3

4
5 657 Viaene, J., Van Lancker, J., Vandecasteele, B., Willekens, K., Bijttebier, J., Ruyschaert, G.,
6 658 De Neve, S., Reubens, B., 2016. Opportunities and barriers to on-farm composting and
7 659 compost application: A case study from northwestern Europe. *Waste Manag.* 48, 181–
8 660 192. doi:10.1016/j.wasman.2015.09.021
9

10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

661 **Fig. 1.** Livestock density in the EU-27, expressed in livestock units (LU) per ha utilized
1 662 agricultural area (UAA). Data were from Eurostat (2010) for the year 2010.

2 663
3
4
5 664 **Fig. 2.** Response to the question: “please indicate the treatment technique(s) in which you are
6 665 involved.” (multiple responses permitted). The number of respondents per country with
7 666 experience in manure treatment is shown in the legend.

8
9
10 667
11
12 668 **Fig. 3.** Responses to the questions (expressed as % of respondents from all survey countries):
13 669 “please indicate the top three reasons that can stimulate farmers to apply manure treatment
14 670 techniques.” The number of respondents that answered this question is shown in the legend.

15 670
16
17 671
18
19
20 672 **Fig. 4.** Responses to the questions (expressed as % of respondents from all survey countries):
21 673 “please indicate the three most important constraints / barriers to the adoption of manure
22 674 treatment technologies.” The number of respondents that answered this question in each
23 675 country is shown in the legend.

24 675
25
26 676
27
28 677 **Fig. 5** Response to the question: “which techniques have the most potential to be applied in
29 678 your country during the next 10 years?” The number of respondents (answered this question)
30 679 for each country is shown in the legend. The number of respondents for each
31 680 technology/answer is indicated in the Y-axis label.

32
33 680
34
35 681
36
37 682 **Fig. 6.** Responses to indicate which farm types (a), sizes (b, LU=livestock unit) and operation
38 683 scales (c) have the most potential for adoption of respective technologies (multiple answers),
39 684 expressed as % of respondents for all four countries. The number (n) of respondents is shown
40 685 for each technology. Results referring to biological nitrogen removal and incineration
41 686 treatment are not shown due to limited number of responses.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

688 **Table 1** A comparison of political and agri-environmental characteristics selected for the four European countries.

	DK	NL	ES	IT
Policies				
-Nitrates Directive				
NVZs (% total agricultural area)	100	100	21	32
Derogation granted	Yes	Yes	No	Yes
-Renewable Energy (RE) Directive ^a				
RE from manure in 2006 (ktoe, kilotonne of oil equivalent)	~70	0	~1.6	n.a.
Estimates in 2020 (ktoe)	~145	~98	~143	n.a.
Agri-environmental conditions				
-Average soil organic carbon in top soils (g kg ⁻¹) ^b	20-30	30-40	<15	15-20
-Manure management systems (% of manure N from housing) ^c				
Dairy cow	92% as slurry	99% as slurry	70% as solid	60% as solid
Other cattle	60% as solid	83% as slurry	99% as solid	60% as slurry
Pigs	95% as slurry	99% as slurry	90% as slurry	99% as slurry

689 ^aSource from National renewable energy action plans; No information (n.a.) available for Italy

690 ^bAdhikari et al., 2014; de Brogniez et al., 2015; Reijneveld et al., 2009

691 ^cInformation from National inventory reports (NIR) to UNFCCC (the United Nations Framework Convention on Climate Change) for the year 2010.

692

693 **Table 2** Overview of respondents as number per country

	DK	NL	ES	IT	Total
Do you have experience in manure treatment? ^a					
Yes	73	66	55	45	239
No	9	13	7	23	52
Total	82	79	62	68	291
What is your job? ^b					
Farmer	10	18	35	12	75
Representative in a farmer organization	10	3	3	5	21
Agricultural advisor	32	6	1	10	49
Technology developer/ user in company	17	30	4	7	58
Employee in the public authority	8	1	3	6	18
Researcher	7	18	15	23	63
Respondents skipped this question	7	9	3	6	25

694 ^a single answer

695 ^b multiple answers

696

Table 3 Summary of responses to the questions asking about the benefits of each respective technology (for all four countries), measured in % of the total number of respondents for each question.

	% of respondents
What are the top three benefits of separation? (Number of respondents: n =102)	
To reduce cost of manure disposal	47
To increase fertilizer value of liquid fractions	39
To increase fertilizer value of solid fractions	34
To use solid fractions for biogas production	27
To use solid fractions for composting	25
To reduce ammonia emissions from liquid fractions after field application	18
To use solid fractions as bedding materials	16
What are the top three benefits of anaerobic digestion? (n=120)	
To produce bioenergy	88
To increase fertilizer nitrogen value of digestate	58
To reduce odor and gaseous emissions during processing	43
To reduce odor and gaseous emissions after field application of digestate	42
To increase soil quality after field application of digestate	13
To increase fertilizer phosphorus value of digestate	8
What are the top three benefits of acidification? (n=44)	
To reduce ammonia emissions during field application	82
To reduce ammonia emissions during storage	73
To increase fertilizer nitrogen value of slurry	68
To increase fertilizer sulfur value of slurry	27
To reduce greenhouse gas emissions during storage	25
What are the top three benefits of composting? (n=48)	
To improve the organic matter quality	54
To remove pathogens	46
To reduce the volume and mass of the manure	42
To improve soil quality after field application of compost	42
To increase economic value as compost products	40
To reduce ammonia emissions after field application of compost	19
To homogenize the manure	13
What are the top three benefits of drying/ pelletizing? (n=34)	
To facilitate export	59
To increase the market value of the manure	53
To reduce costs of transporting manure surplus off farm	41
To increase soil quality after field application of dried products	29
To decrease ammonia emissions after field application of dried products	26
What are the top three benefits of membrane filtration/ reverse osmosis? (n=25)	
To increase fertilizer effectiveness of nitrogen as concentrates	64
To make a K fertilizer	48
To reduce cost of transporting phosphorus surplus off farm	44
To remove organic matter from liquid manures	28
To reduce ammonia emissions after field application of concentrates	12

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

Figure 1

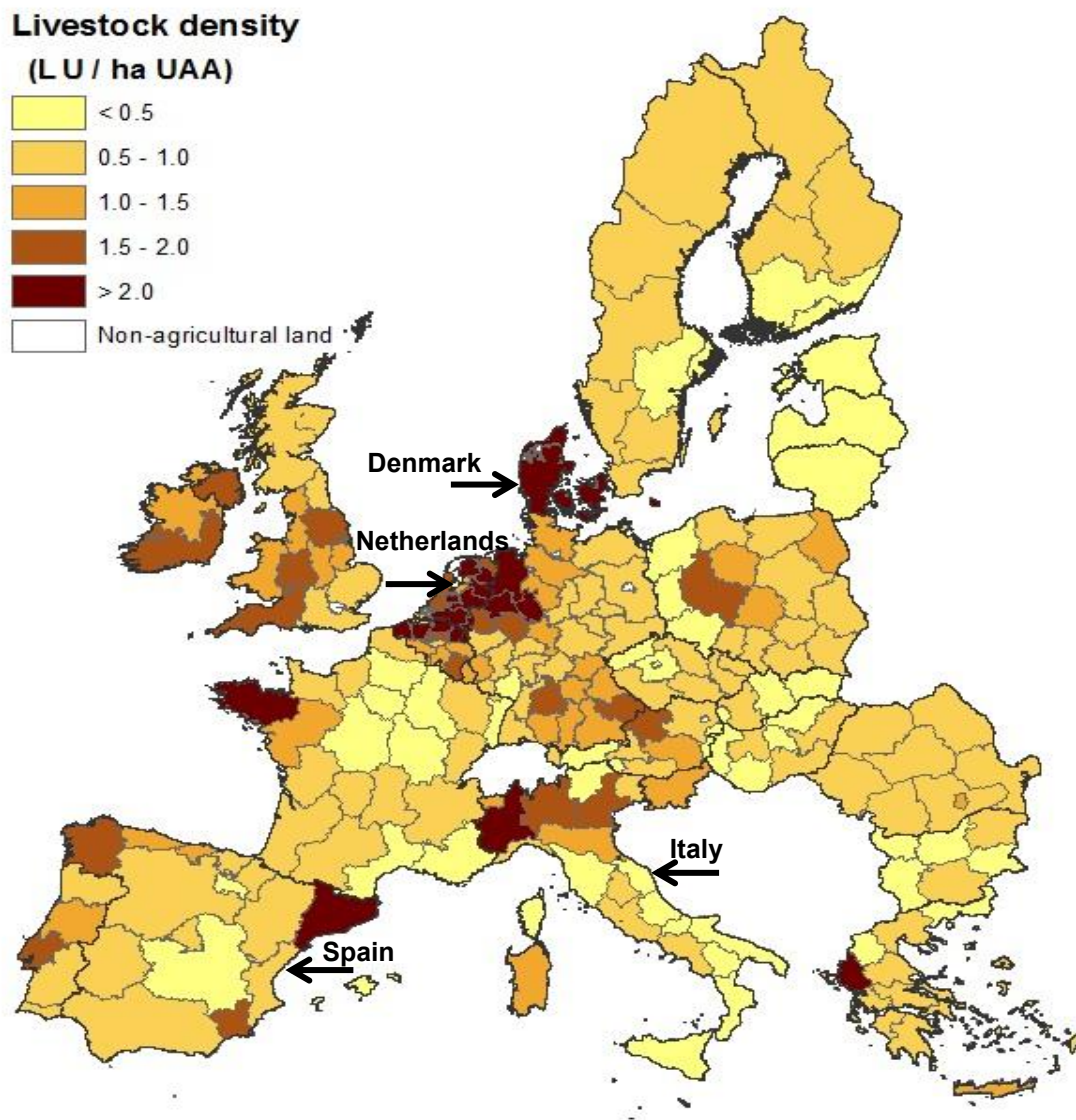


Figure 2

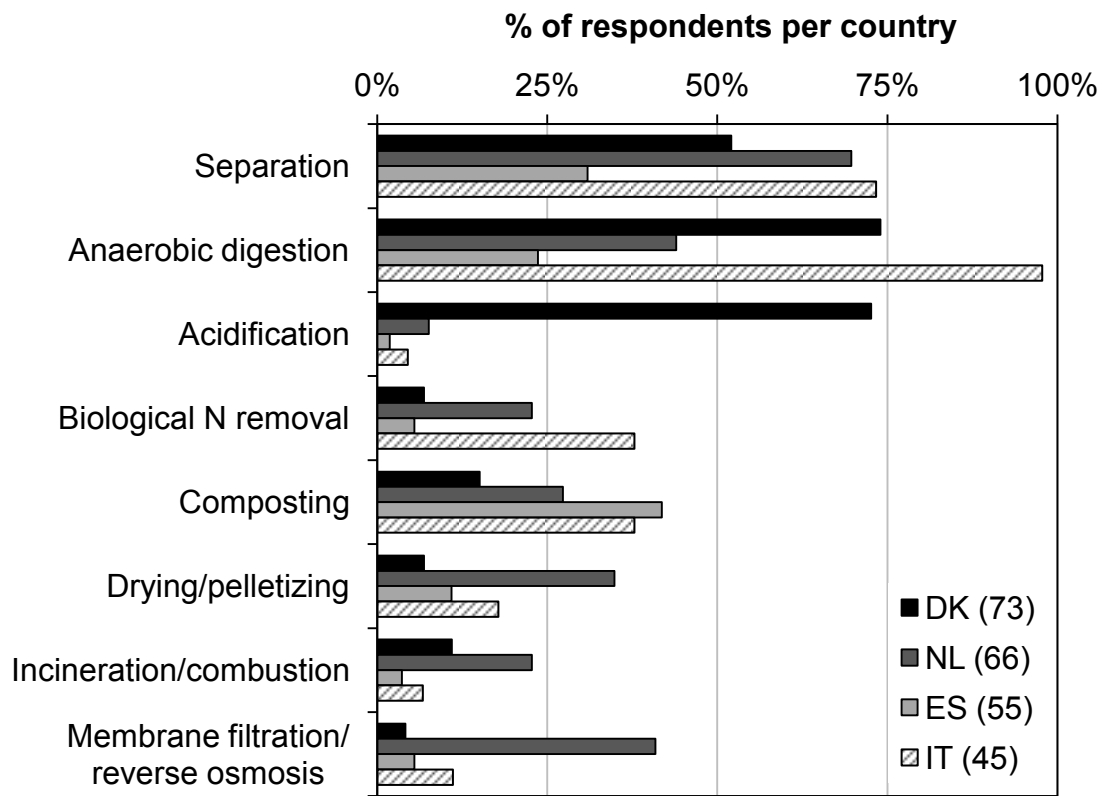


Figure 3

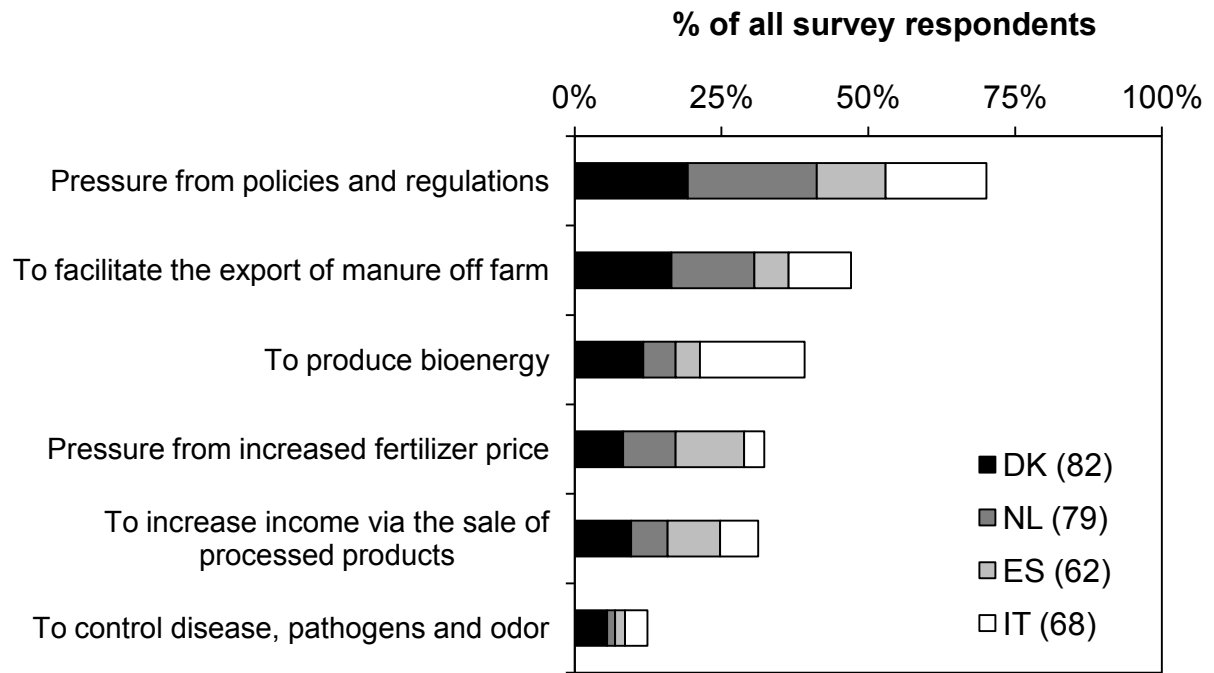


Figure 4

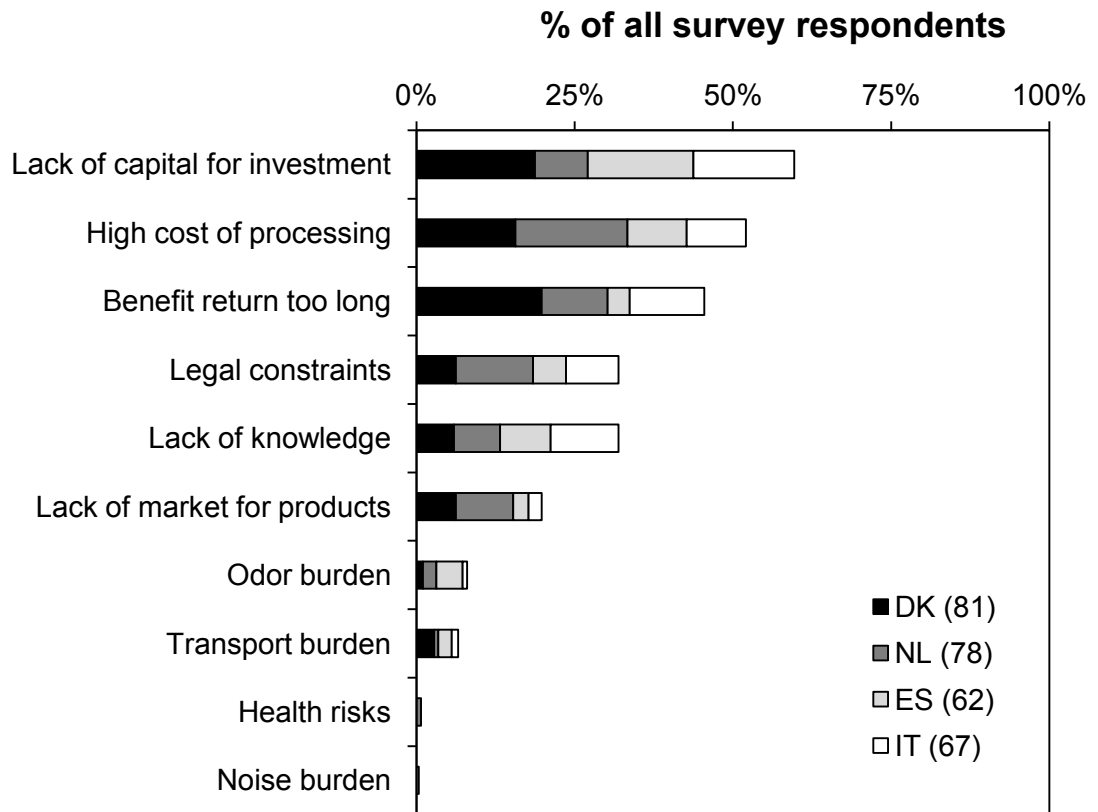


Figure 5

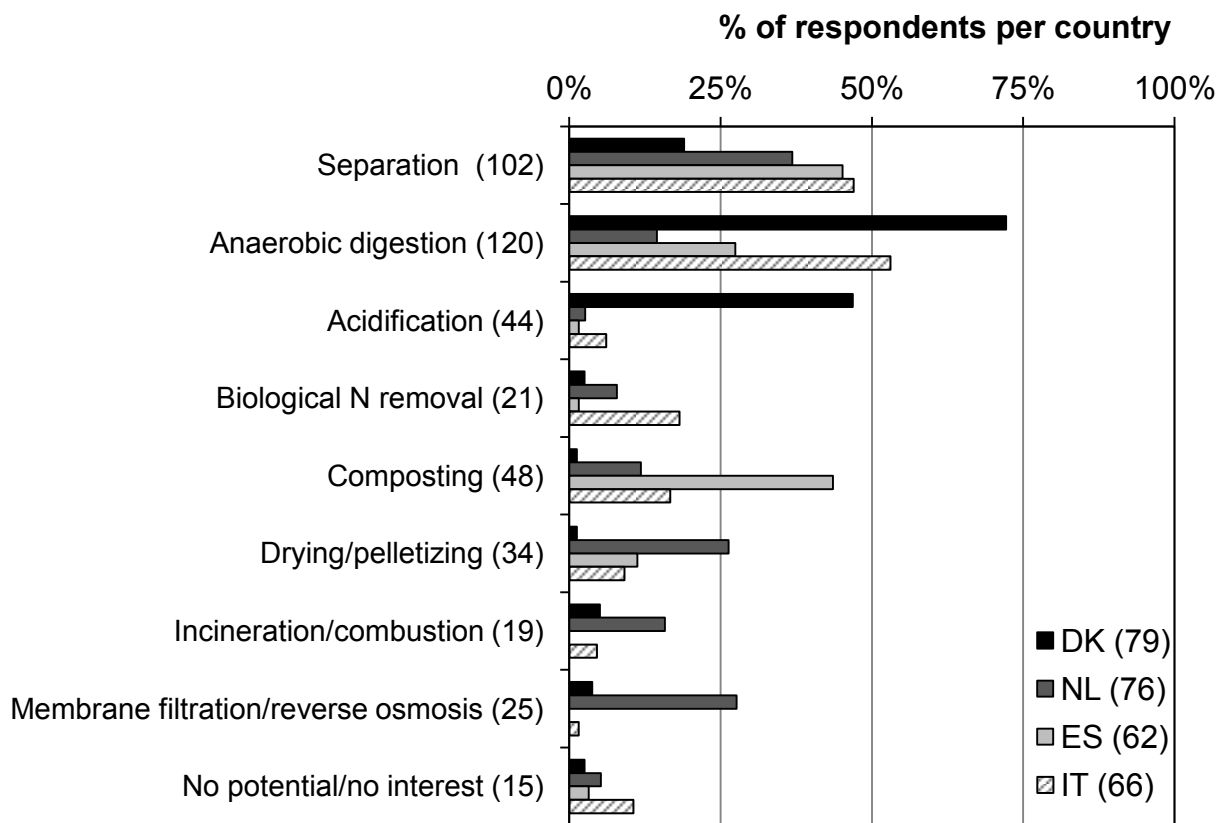
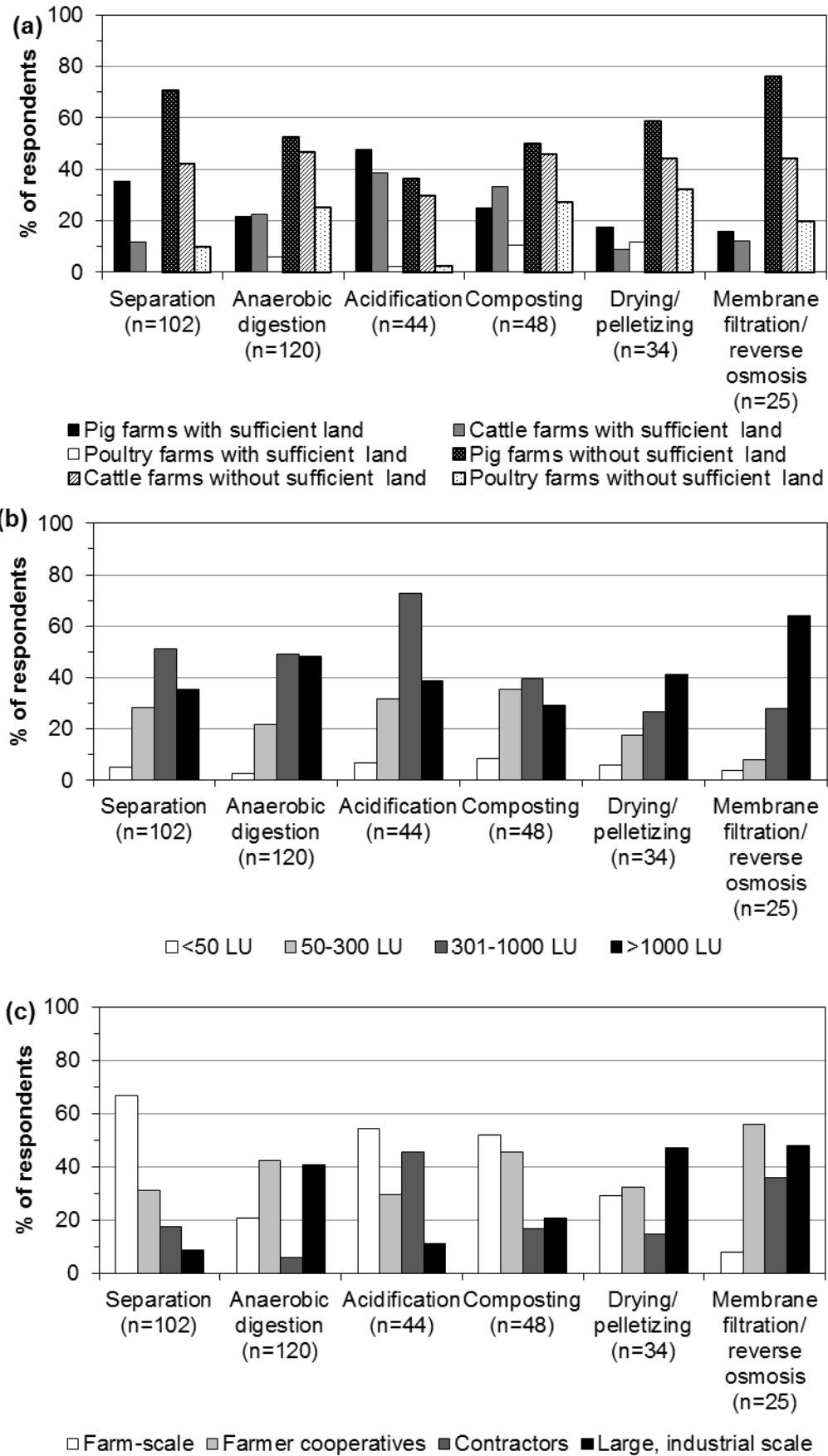


Figure 6



Appendix A

[Click here to download e-component: Appendix A.docx](#)

Appendix B

[Click here to download e-component: Appendix B.docx](#)

Supplementary info

[Click here to download e-component: Appendix- Supplementary info.pdf](#)