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1 **Species interactions increase the temporal stability of community productivity in *Pinus***
 2 ***sylvestris-Fagus sylvatica* mixtures across Europe**

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57 Running headline:

58 Species interactions enhance stability in mixed forests

59

60 **Abstract**

61

62 1. There is increasing evidence that species diversity enhances the temporal stability of
63 community productivity in different ecosystems, although its effect at population and tree
64 levels seems to be negative or neutral. Asynchrony between species was found to be one of
65 the main drivers of this stabilizing process. However, scarce research in this area has been
66 undertaken in forest communities, so determining the effect of species mixing on the stability
67 of forest productivity as well as the identity of the main drivers involved still poses a
68 challenging task.

69 2. We investigate the way in which mixing species influences the temporal stability of
70 productivity in *Pinus sylvestris* L. and *Fagus sylvatica* L. forests, and attempt to determine the
71 main drivers. We used a network of 93 experimental plots distributed across Europe to
72 compare the temporal stability of basal area growth over a 15-year period (1999-2013) in
73 mixed and monospecific forest stands at different organizational levels, namely community,
74 population and individual tree levels. Overyielding, asynchrony between species, and species
75 interactions were explored as possible drivers of temporal stability of productivity.

76 3. Mixed stands showed a higher temporal stability of basal area growth than monospecific
77 stands at community level, but not at population or individual tree levels. Asynchrony
78 between species growth in mixtures was related to temporal stability, but neither overyielding
79 nor asynchrony between species growth in monospecific stands were linked to temporal
80 stability. Therefore, species interactions modify between-species asynchrony in mixed stands.
81 Accordingly, temporal shifts in species interactions were related to asynchrony and to the
82 mixing effect on temporal stability.

83 4. Synthesis. Our findings confirm that species mixing can stabilize productivity at
84 community level whereas there is a neutral or negative effect on stability at population and
85 individual tree level. The contrasting findings as regards the relationships between temporal
86 stability and species asynchrony in mixed and monospecific stands suggest that the main
87 driver in the stabilizing process is the temporal niche complementarity between species rather
88 than differences in species specific responses to environmental conditions.

89

90 **Keywords**

91

92 Temporal variability; mixed-species forests; plant-plant interactions; overyielding;
93 asynchrony; niche complementarity; organizational levels;

94

95 **Introduction**

96

97 Mixed-species stands are widely thought to provide many forest functions and services more
98 effectively than monocultures (Hector & Baghi 2007; Gamfeldt *et al.* 2013; van der Plas *et*
99 *al.* 2016). The superior level and stability of productivity in mixed forests is of interest for
100 most functions and services, as well as being a precondition for the promotion of this

101 alternative in forestry practice. Much evidence exists that mixed-species stands often produce
102 greater yields than monocultures (Piotto 2008; Paquette & Messier 2011; Vilà *et al.* 2013;
103 Pretzsch *et al.* 2015; Liang *et al.* 2016) although contradictory findings of underyielding
104 (Chen *et al.* 2003; Carvard *et al.* 2010) discourage generalization. Many studies show that
105 mixing may improve different aspects related to the stability of productivity (Jucker *et al.*
106 2014; Pretzsch, Schütze & Uhl, 2013; de Dios-García, Pardos & Calama, 2015; Metz *et al.*
107 2016), but again, the findings of other research suggest the opposite (Grossiord *et al.* 2014;
108 Merlin *et al.* 2015). Among the probable reasons for these varying and seemingly inconsistent
109 findings are differences in the complementarity of the analyzed species assemblages (Toïgo *et al.*
110 2015) as well as the underlying site conditions with their specific growth limiting factors
111 (Forrester 2014). Findings may also differ depending on the level of analysis, as mixing
112 effects in forest communities are frequently studied at stand, species, or individual tree level;
113 the results not necessarily being the same (Forrester & Pretzsch 2015). The conservation and
114 management of productive, stable, and resource-use efficient mixed-species stands requires an
115 improved understanding of the mechanisms involved, which could also contribute towards
116 theory development and greater generalization with regard to these forests.

117 The term ‘stability’ in ecosystems includes several concepts such as resistance, resilience or
118 temporal stability of productivity, all of which address diversity-stability relationships
119 (McCann 2000; Ives & Carpenter 2007). In the case of forests, temporal variability of
120 community productivity is an important ecological property because stability of productivity
121 is an indicator of sustainability of both forest functioning and the delivery of ecosystem
122 services (Blüthgen *et al.* 2016). Temporal variability is usually measured by the coefficient of
123 variation or its inverse, i.e temporal stability then depends on the mean and standard deviation
124 (Tilman, Lehman & Bristow 1998). Different statistical and biological mechanisms have been
125 identified as possible causes of increasing temporal stability as regards species diversity.
126 These include overyielding, species asynchrony and species interactions (Hector *et al.* 2010,
127 Loreau & Mazancourt 2013; Blüthgen *et al.* 2016). Overyielding means higher productivity in
128 mixtures than in the corresponding monospecific systems, which may lead to a stabilizing
129 effect by a higher mean if other factors remain constant (Tilman 1999). Species asynchrony
130 exists when the temporal responses of the species are not perfectly positively correlated. Such
131 increases in the variability of responses may result in a reduction in the community
132 variability. Asynchrony of species-specific responses to environmental fluctuations has been
133 reported as a key factor in temporal stability (Loreau & de Mazancourt 2008; Hector *et al.*
134 2010), in accordance with the insurance hypothesis (Yachi & Loreau 1999). However, species
135 interactions can also trigger species asynchrony by compensatory dynamics between species
136 (Tilman, Lehman & Bristow 1998; Morin *et al.* 2014), which might result in less variation at
137 community level (Loreau & de Mazancourt 2013). Species interactions may also involve
138 temporal stability as a consequence of their effect on overyielding, and at the same time
139 overyielding may be linked to species asynchrony (Allan *et al.* 2011). These direct and
140 indirect relationships make it difficult to disentangle the key mechanisms and therefore the
141 relative importance of the different mechanisms on the diversity-stability relationship is still
142 poorly understood (Loreau & de Mazancourt 2013).

143 In general, diversity has been found to have a stabilizing effect on productivity at community
144 level, but a destabilizing effect at population levels by increasing competitive interactions
145 (Hector *et al.* 2010; Gross *et al.* 2014). However, contrasting results have been obtained at
146 population level (Jiang & Pu 2009), even among the scarce studies undertaken in forest
147 communities (Jucker *et al.* 2014; Morin *et al.* 2014). This trade-off between the effects at
148 different organizational levels might be crucial in ecosystems with few species, where the
149 species specific dynamic can be of major interest, as in many European temperate mixed
150 forests comprising only two or three species.

151 Diversity- roductivity relationships in forests have been found to depend on environmental
152 gradients (Pretzsch *et al.* 2010; Toïgo *et al.* 2015; Jucker *et al.* 2016), since the result of the
153 interactions among species changes depending on the growing conditions (Forrester, 2014;
154 Forrester & Bauhus, 2016). The growth response of tree species to climatic conditions as well
155 as temporal variation in climate-growth relationships also vary considerably among sites
156 (Lloyd & Fastie 2002; Tardif *et al.* 2003). Therefore, differences in diversity-stability
157 relationships might also be expected along ecological gradients, with the relative importance
158 of different mechanisms varying along the gradients (Hallet *et al.* 2014; Xu *et al.* 2015).

159 The number of studies concerning the relationship between diversity and temporal stability of
160 productivity based on empirical data in forests is far fewer than in grasslands communities.
161 This is due to the inherent arduousness involved in carrying out experiments with tree species,
162 due to their long life span, as well as the difficulties of conducting observational studies in
163 natural ecosystems, where many often uncontrollable factors interact. In a recent study, Jucker
164 *et al.* (2014) analysed several monospecific and mixed forests of 16 target species in Europe
165 (Jucker *et al.* 2014) and found a positive effect of species diversity on the stability of wood
166 productivity. However, a previous study found the opposite for conifer mixed forests in Sierra
167 Nevada, California (DeClerck, Barbour & Sawyer 2006). Therefore, further research is
168 required to elucidate the mixing effect on temporal stability of productivity and the
169 underlying mechanisms for different forest species assemblages and sites.

170 In this study we focus on two tree species, *Pinus sylvestris* L. and *Fagus sylvatica* L., growing
171 in mono-specific and mixed forests across a large range of their distribution. This design
172 allows us to infer the general effect of this admixture on the temporal stability of productivity
173 while considering the large spatial variability in site conditions across Europe. This mixture
174 was selected because it includes a combination of species with highly complementary traits,
175 including an early and a late-successional species, a light-demanding as opposed to a shade-
176 tolerant species, and a conifer with a broad-leaved species. Actually, the mixture between *P.*
177 *sylvestris* and *F. sylvatica* was found to shown significant mixing effects in terms of
178 productivity and structural heterogeneity (Pretzsch *et al.* 2015, 2016). It may serve as a model
179 system for other widespread species combinations of comparable spatial and temporal
180 complementarity in traits.

181 The main hypotheses in this study are that: (i) temporal stability of productivity is higher in
182 mixed than in mono-specific stands at community level but not at population and individual
183 tree levels; (ii) in this model mixture, the dynamics of species interactions is one of the
184 drivers in stabilizing productivity due to the complementary traits of these species; and (iii)

185 the mixing effect on temporal stability depends on site conditions. Our main objective is
 186 therefore to explore whether mixing species of contrasting traits increases the temporal
 187 stability of productivity at different organizational levels and if so, to elucidate the main
 188 underlying mechanisms in order to better understand the inter-specific dynamics of the
 189 *P.sylvestris* - *F.sylvatica* and comparable mixtures.

190

191 MATERIAL AND METHODS

192 Field data and study design

193 The study data come from a transect of plots in mixed and monospecific forest stands of *P.*
 194 *sylvestris* (Scots pine) and *F. sylvatica* (European beech) along an environmental gradient.
 195 The transect was established voluntarily and nationally-funded by members of the COST
 196 Action FP1206 EuMIXFOR (see www.mixedforests.eu). The main aim of the initiative was
 197 to study the variability of over-yielding, structural properties and stability under different
 198 environmental conditions in monocultures and mixtures (see for example Pretzsch *et al.* 2015;
 199 2016). The study design was based on the ‘triplet’ concept (Pretzsch *et al.* 2014), i.e. at each
 200 location three plots were established, one in a mixed-species stand and two in the respective
 201 monocultures, with similar site conditions (soil and topographic conditions) in order to allow
 202 meaningful comparisons between mixtures and monocultures. A total of 31 triplets (93 plots)
 203 were set up across the main distribution area of this mixture in Europe (Fig. 1), covering a
 204 large environmental gradient, mainly determined by water supply. Climate data were gathered
 205 from all available meteorological stations in the proximity of each triplet (see Table S1 in
 206 Supporting Information for detailed information about climate and site conditions).

207 The three plots for each triplet were installed in even-aged, fully-stocked forest stands of
 208 similar age in which thinning treatments had not been recently applied (for details see Table
 209 S2 and Pretzsch *et al.* 2015, 2016). The mixed plots represent tree-wise mixtures with species
 210 proportions that range from 18% to 72 % of pine, although in most of them the proportion is
 211 around 50%. Plots are rectangular with varying sizes from 0.02 to 1.55 ha. In each plot, the
 212 tree species, tree diameter, height and height to the crown base were recorded for all trees. In
 213 a sub-sample of 20 trees per plot and species two increment cores were extracted at a stem
 214 height of 1.30 m for tree ring analysis. Annual growth series were cross-dated and the
 215 arithmetic means of the annual ring widths of the two cores were used for further analysis. A
 216 description of the main stand characteristics in mixed and monospecific stands are provided in
 217 Table S2.

218 Productivity data at different organizational levels

219 *Community level*

220 As a proxy to represent community biomass productivity we use stand basal area growth per
 221 hectare, as it is closely linked to measured variables in the field. In contrast to other studies
 222 which focused on aboveground biomass growth when studying diversity- roductivity and/or
 223 diversity-stability relationships (Paquette & Messier 2011; Jucker *et al.* 2014, 2016), we
 224 relied on basal area growth. Calculation of stand biomass growth would have required height-
 225 diameter functions and tree biomass allometric functions for all sites. However, it is well
 226 known that such calculations could lead to additional uncertainty at least in mixed stands
 227 (Toïgo *et al.* 2015) as the respective functions were derived from data of monospecific stands.
 228 Using these functions may had caused biased estimations of biomass growth as mixing tree
 229 species can modify tree allometry (Pretzsch 2014) as well as between-tree growth partitioning

230 (Binkley *et al.* 2003; Pretzsch & Schütze 2014), suggesting the need of specific functions for
 231 mixtures (Forrester & Pretzsch 2015; Río *et al.* 2016).

232 Our study focuses on the temporal stability and over-yielding during the 15 year period prior
 233 to the inventory (1999-2013). This period was chosen because it covers sufficient years to
 234 provide meaningful information on temporal variability in growth, whilst avoiding bias from
 235 unknown tree mortality or tree removal which could have interfered the results as mixing may
 236 change species-specific mortality rates (Zhao *et al.* 2006; Condés & Río 2015).

237 Stand basal area was calculated as the sum of the cross sectional area (at 1.3 m above ground
 238 level) of all the trees measured at a given time. Stand basal area increments per year were
 239 determined based on cored trees and non-cored trees. In the case of sampled trees, we used
 240 tree ring series to reconstruct tree diameters over bark for each of the 15 years of the study
 241 period. To estimate the diameter increments of non-cored trees we fitted diameter increment
 242 functions for each plot and species per year, based on diameter increments and tree diameters
 243 of cored trees (31 triplets * 4 (two tree species in mixed and monospecific stand) * 15 years =
 244 1980 functions for the studied period 1999-2013). We used log-log models ($\ln(id)=a_0+a_1 \times$
 245 $\ln(d)$), where id is the tree diameter increment for that year (cm year^{-1}) and d is the tree
 246 diameter at breast height (cm).

247 *Population level*

248 To study the productivity at population level we additionally calculated the annual basal area
 249 increment (BAI) per species in mixed plots. In order to compare species behavior in mixed
 250 and monospecific stands we scaled up the species specific basal area increment series in
 251 mixed stands to one hectare using species basal area proportions. As species proportion can
 252 change from one year to another due to the different annual basal area increments between
 253 species we calculated species proportions per year through the estimated annual basal area per
 254 species.

255 *Individual tree level*

256 At individual tree level we used the measured tree ring widths from cored trees transformed to
 257 individual tree basal area increments. As the tree growth response to variability in
 258 environmental conditions and to intra- and inter-competition level depends on tree social
 259 status (Martín-Benito *et al.* 2008; Zang, Pretzsch & Rothe 2012; Río, Condés & Pretzsch
 260 2014) we used only dominant and codominant trees (1691 trees), selected through the
 261 diameter and height distributions per species and plot.

262 **Data evaluation and analysis**

263 *Temporal Stability at different organizational levels*

264 Temporal stability (TS) at the different organizational levels was calculated as the inverse of
 265 coefficient of variation for the 15 year study period, i.e. the ratio of mean basal area increment
 266 to its standard deviation. This measure is often preferred to the coefficient of variation, as the
 267 latter decreases with stability and when the stability increases it approaches zero (Lehman &
 268 Tilman 2000). Statistics of the mean, standard deviation and temporal stability of annual basal
 269 area increment at the different organizational levels are presented in Table S3.

270 The effect of mixing species on temporal stability of productivity at community and
 271 population level was analyzed using a mixed linear model including the species composition
 272 of the plot as a fixed factor. First we compared mixed vs monospecific stands, and in a second
 273 step we considered species identity of monospecific plots. Data were log-transformed to
 274 correct heteroscedasticity in residuals.

$$275 \quad \text{Ln}(TS_{ij}) = (a_0 + a_{0j}) + a_1 \cdot \text{composition} + \varepsilon_{ij} \quad \text{eqn 1}$$

276 where TS_{ij} is the temporal stability of the annual basal area increment for the plot i in the
 277 triplet j ; composition is a dummy variable with two levels, mixed and monospecific, or three
 278 levels, mixed, monospecific pine and monospecific beech; a_0 and a_1 are parameters to be
 279 estimated. We included a random effect (a_{0j}) due to the hierarchical structure of the data to
 280 account for possible correlation of the three plots within a triplet. Covariates potentially
 281 influencing TS included climatic attributes and their interaction with species composition
 282 were tested. At tree level we fitted a similar model but taking also the effect of tree size on
 283 temporal stability into account.

284 In order to study the effect of mixing on TS at different organizational levels we first defined
 285 the mixing effect as the ratio of TS in mixed stands to TS in monospecific stands
 286 ($TS_{\text{mixed}}/TS_{\text{mono}}$) and then we analyzed the correlation between the ratios at community,
 287 population and individual tree levels.

288 *Overyielding*

289 The over- or under-yielding values per triplet were estimated using the ratio of productivity
 290 (RPP) (Harper, 1977), $RPP = \sum P_{i,\text{mix}}/P_{i,\text{mono}}$, where $P_{i,\text{mix}}$ is the observed productivity (i.e. basal
 291 area increment) of species i in the mixed stand and $P_{i,\text{mono}}$ is the productivity of species i in
 292 the monospecific stand. We estimated the RPP per year and triplet for the 15 year study
 293 period and then averaged them per triplet.

294 To estimate the overyielding at population level we used the relative productivity per species
 295 (RP_i) (Pretzsch *et al.* 2013; Río *et al.* 2016), i.e. the ratio of the observed productivity of
 296 species i in the mixed stand (up-scaled to one hectare) to the observed productivity of the
 297 respective species in the monoculture, $RP_i = (P_{i,\text{mix}}/m_i)/P_{i,\text{mono}}$, where m_i is the species
 298 proportion estimated by the proportion of species i in the stand basal area for a given year. As
 299 for RPP, RP_i were estimated per year and later averaged for the 15 years in order to consider
 300 the possible influence of temporal changes on species proportion. We tested whether the mean
 301 RPP and RP_i were significantly different from one, i.e. significant over- or under-yielding,
 302 using a t-student test, and the possible relationship between overyielding and temporal
 303 stability at different levels through simple linear models. At community level we studied the
 304 possible influence of RPP on the temporal stability in mixed stands (TS_{mixed}) and on the
 305 mixing effect ($TS_{\text{mixed}}/TS_{\text{mono}}$). At population level we related the RP_i to the mixing effect, i.e.
 306 ratio of TS at population level.

307 *Asynchrony*

308 To estimate the species asynchrony we used the coefficient of correlation between the growth
 309 series of the two species growing in mixed stands (r_{mixed}); a value of -1 means complete
 310 asynchrony between species' growths and +1 indicates complete synchrony. This approach is
 311 similar to that proposed by Gross *et al.* (2014), although in its simplest version of a mixture
 312 composed of only two species. Additionally, we studied the correlation between the basal area
 313 increment series of the two species growing in monocultures (r_{mono}), as this correlation might
 314 express the differences or the similarity in the dependence of the two species on inter-annual
 315 environmental conditions, i.e. the asynchrony of the intrinsic response of each species to
 316 environmental fluctuations (Loreau & de Mazancourt 2013). Species asynchrony was
 317 estimated at the community level by stand basal area increment series of the two species. At
 318 tree level it was studied by species specific mean tree basal area increment series.

319 We explored the role of species asynchrony in TS in a similar way than for overyielding, i.e.
 320 by using linear models for relating TS_{mixed} and the ratios of $TS_{\text{mixed}}/TS_{\text{mono}}$ to r_{mixed} and r_{mono} at

321 different levels. Furthermore, we tested whether there was any relationship between species
322 asynchrony and overyielding.

323 *Temporal variation in species interactions*

324 To study the inter-annual variation in species interactions depending on annual growing
325 conditions we used a similar approach to that used in Río, Schütze & Pretzsch (2014). We
326 compared the annual productivity in mixed stands to the respective reference productivity.
327 The latter reflects conditions where no mixing effect takes place, which is calculated as the
328 sum of the productivities of the two species in monospecific stands times their proportion in
329 the mixed stand ($\sum P_i \cdot m_i$) (Pretzsch *et al.* 2013; Río *et al.* 2016). When the annual basal area
330 increment in the mixed stand is higher than the reference basal area increment, there is a
331 positive species interaction or overyielding; whereas if one year it is lower this indicates that
332 there is negative interaction or underyielding. In this section, as the aim is to study the
333 temporal variation in species interaction but not the net effect or overyielding, we
334 standardized the observed and reference basal area increment series by dividing them by the
335 mean and we built the respective basal area growth indices series (IBAI_{mixed} and IBAI_{ref}) to
336 remove the net overyielding effect for the 15 year period (see Fig S1).

337 A year was considered to have favorable growing conditions when the IBAI was high and
338 unfavorable when the IBAI was low. To test whether annual species interactions vary
339 depending on growing conditions we fitted a linear model relating the two growth indices
340 series (IBAI_{mixed} = f(IBAI_{ref})). If the slope is not different from one, the temporal variation in
341 species interaction does not depend on annual growing conditions (i.e variation is similar in
342 good and bad years), whereas if the slope is different from one it means that the interactions
343 depend on annual growing conditions (see Fig S1). As the two variables are assumed to be
344 measured with the same error and we were interested in the slope value and not in predicting
345 new IBAI values, we used a major regression to estimate the slope per triplet and then
346 explored if the slope values were related to TS.

347 **RESULTS**

348 **Temporal stability at different levels: community, species and individual tree level**

349 *Community level*

350 Temporal stability of annual stand basal area increment was lower in the monospecific stands
351 than in the mixed stands ($P = 0.010$), the observed mean being TS=5.14 and 6.08 respectively.
352 When the composition of monospecific stand was considered the TS in monospecific
353 European beech plots was lower than the mixed plots ($P = 0.012$), whereas for Scots pine it
354 was also lower although the difference was smaller ($P = 0.052$) (Table S4). We tested the
355 possible influence of climatic variables but found no significant relationships. When
356 analyzing the mean and the standard deviation of stand BAI there were no statistical
357 differences between compositions.

358 *Population level*

359 There were no statistical differences between the TS of annual basal area growth in mixed
360 (expanded to hectare) and in monospecific stands at population levels. For pine, both the
361 mean of annual basal area increments and the standard deviation were significantly lower in
362 mixed than in monospecific stands, whereas for beech the mean and the standard deviation
363 were significantly higher in mixed than in monospecific stands. Climatic variables did not
364 explain TS variability for either of the two species.

365 *Individual tree level*

366 TS in annual tree basal area increment was significantly different between pure and mixed
 367 plots for pine ($P < 0.001$), being greater in monospecific stands. The inclusion of the tree size
 368 or site covariates did not improve the basic model. The increase in TS in monospecific stands
 369 was due to a higher mean tree BAI, as the differences in the mean were significant between
 370 monospecific and mixed stand whereas in the case of the standard deviation they were not.
 371 For beech, there were no differences in tree TS between mixed and monospecific stands, but
 372 the tree size had a significant effect on tree TS (Table S4). Both the mean and the standard
 373 deviation were significantly higher in the mixed compared to the monospecific stands.

374 *Overall effect*

375 The results showed that at community level the mixture leads to stability of productivity, but
 376 this effect disappears at population level while at tree level the opposite effect was observed
 377 in the case of pine. The stability is lower at population level than at community level,
 378 particularly for beech (Fig. 2a). The mean ratios $TS_{\text{mixed}}/TS_{\text{mono}}$ at community level were 1.31
 379 and 1.28 for beech and pine respectively, whereas at population level they were not
 380 significantly different from one. There is a positive correlation (r) between the mixing effect
 381 on stability at the two levels for both species ($r = 0.763$ $P < 0.0001$ for pine and $r = 0.716$ P
 382 < 0.0001 for beech). If we compare the mixing effect on stability at individual tree, population
 383 and community level we observe that there is no correlation between the effects of mixing on
 384 stability at tree level with the corresponding effects at the other two organizational levels (Fig.
 385 2b).

386 **Overyielding**

387 The mean RPP of all triplets was 1.12 and it was statistically different from 1. This indicates
 388 that there was a general overyielding in stand basal area growth although the variability
 389 among triplets was large with some triplets showing underyielding (Fig S2). The RPP was not
 390 related to any of the site variables analyzed, nor to the TS in mixed stands. Accordingly,
 391 overyielding was not related to any of the mixing effects of TS at community level (ratio of
 392 TS in mixed stands to monospecific stands) (Fig. S2).

393 At population level we found overyielding in the case of beech (Relative productivity ($RP_{\text{be}} =$
 394 1.49) and underyielding for pine ($RP_{\text{pi}} = 0.87$), both significantly different from one (note that
 395 there was no correlation between the RP_i of the two species). $TS_{\text{mixed}}/TS_{\text{mono}}$ ratio at
 396 population level (i.e. mixing effect on stability) was negatively related to the relative
 397 productivity by species (RP_i). Thus, with increasing overyielding stability decreased in mixed
 398 stands (Fig 3). This suggests that at population level, under-yielding is linked to higher
 399 stability for pine, but it is important to highlight the absence of differences between mixed
 400 and monospecific stands in TS at this level.

401 **Species asynchrony**

402 The mean coefficient of correlation between basal area increment series of beech and pine in
 403 the mixed stand (r_{mixed}), or species synchrony at community level, was 0.37, but there was a
 404 high variability among triplets ranging from -0.62 to 0.89 (Fig. 4). The observed high
 405 negative values revealed the presence of a high species asynchrony at community level for
 406 some triplets. The respective mean correlation in monospecific stands (r_{mono}) was similarly
 407 0.37 with a narrower range (-0.39 to 0.87), which indicates that in some triplets the two
 408 species use the annually available site resources differently whereas in other cases the
 409 response to the interannual fluctuations in environmental conditions is quite similar.
 410 However, it is important to highlight that the relationship between r_{mixed} and r_{mono} was not
 411 significant (Fig. S3), reflecting that the mixture changes the species-specific responses to

412 annual environmental conditions. No effect of any site characteristic on correlation between
413 species' basal area increments was found.

414 The temporal stability of community productivity in mixed stands was partially explained by
415 the species asynchrony in mixed plots (Fig. 4), following a quadratic model ($R^2 = 0.40$;
416 $P < 0.001$). For coefficients of correlation higher than 0.6 the TS_{mixed} decreases notably.
417 Therefore, when the species asynchrony was lower, the stability in the mixture was lower.
418 However, this relationship was not significant when considering the correlation in
419 monocultures instead of in mixtures (Fig. S4). The mixing effect on stability at community
420 level (ratio $TS_{\text{mixed}}/TS_{\text{mono}}$) increased in the case of pine when the species asynchrony in
421 mixed stands was higher ($R^2 = 0.25$; $P = 0.004$), but this effect was not significant for beech
422 (Fig. S5).

423 At individual tree level the mean correlation between the mean tree basal area growth series
424 of beech and pine was 0.41 in mixtures, varying between -0.65 to 0.91, whereas the respective
425 mean correlation in monocultures was 0.32 with a narrower range (-0.35 to 0.77). In contrast
426 to the results observed at community level, the coefficients of correlation in mixed and
427 monospecific stands are correlated ($r = 0.43$, $P < 0.0161$). The coefficients of correlation at
428 tree level and at community level are positively correlated in mixed stands ($r = 0.58$, $P <$
429 0.0005) and in monocultures ($r = 0.74$, $P < 0.0001$). The asynchrony at tree level was not
430 related to temporal stability at individual tree and species level.

431 The relationship between overyielding (RPP) and species asynchrony in mixed stands at
432 community level was significant ($R^2 = 0.20$; $P = 0.011$), the overyielding increasing with the
433 species asynchrony (Fig. 5). However, this relationship was not significant when relating RPP
434 to the coefficient of correlation in monocultures. Therefore, the species asynchrony in mixed
435 stands has an influence on the temporal variability and quantity of productivity at community
436 level.

437 **Species interactions**

438 The results of the major regression per triplet, relating the observed and reference stand basal
439 area growth indices, indicated that the slope was statistically different from one in 10 out of
440 the 31 triplets ($P < 0.05$), 5 having a slope higher than one and 5 with a slope lower than one.
441 The relationship between the temporal stability in mixed stands (TS_{mix}) and the slope values
442 was negative ($R^2 = 0.21$; $P = 0.010$). Hence, higher temporal stability seems to be linked to
443 slopes lower than one and lower stability to higher slopes. As with other variables, site
444 characteristics were not significant.

445 Accordingly the slopes were also negatively related to the mixing effect on stability
446 ($TS_{\text{mixed}}/TS_{\text{mono}}$). In Fig. 6 it can be seen that lower slopes are linked to triplets where the TS
447 is higher in mixed than in monospecific stands and this is particularly notable for pine ($R^2 =$
448 0.32 ; $P = 0.001$ for beech; and $R^2 = 0.53$; $P < 0.001$ for pine). Thus, the reduction in temporal
449 variation of productivity in mixed stands compared to monocultures is linked to a temporal
450 variation in species interaction, this interaction being more positive in years with low growth
451 rates and more negative in years with high growth rates. In triplets where the stability is
452 higher in monospecific stands, the slopes tend to be greater than one, which means more
453 positive interactions in years with high growth and more negative interactions in years with
454 low growth rates.

455 The slopes explained part of the variability in the coefficient of correlation between basal area
456 increment series of beech and pine in the mixed stand (r_{mixed}) ($R^2 = 0.16$; $P = 0.027$). The
457 positive relationships between them suggest that part of the asynchrony observed in mixed
458 stands is due to temporal changes in species interactions.

459

460 **DISCUSSION**

461 Our findings show that species mixing can stabilize productivity at community level but not at
 462 population level. This stabilizing effect was mainly explained by species asynchrony in the
 463 mixed stands, which was influenced by the species interactions. This result along with the
 464 lack of any relationships between temporal stability and species asynchrony in monospecific
 465 stands suggests that the main driver in the stabilizing process was the temporal niche
 466 complementarity between species rather than differences in species-specific responses to
 467 environmental conditions. Overyielding was not linked to temporal stability but to species
 468 asynchrony in mixed stands, highlighting the important contribution of temporal niche
 469 complementarity to the level and stability of forest productivity.

470 **Drivers of temporal stability and the level of productivity**471 *Overyielding*

472 Overyielding was found to contribute to the stabilization of productivity in different types of
 473 communities (Hector *et al.* 2010; Isbell, Polley & Wilsey 2009, Jucker *et al.* 2014). Our
 474 analysis showed a significant overyielding at community level, but it was not linked to the
 475 temporal stability of productivity (Fig. S2). This result for our two species mixture is contrary
 476 to the findings of Jucker *et al.* (2014) for tree mixtures of 2-4 species. Based on long-term
 477 simulations, Morin *et al.* (2014) reported that temporal stability was weakly driven by
 478 overyielding, which is in line with our results. However, it is important to consider that the
 479 stabilizing effect of overyielding may increase with species diversity, and may therefore have
 480 a relatively small effect in two-species mixtures, such in our case (Hector *et al.* 2010).

481 *Asynchrony*

482 The important role of species asynchrony in community stability has been highlighted
 483 recently in many studies (Roscher *et al.* 2011; Blüthgen *et al.* 2016). The results from our
 484 study confirm that asynchrony in species growth is an important driver of temporal stability
 485 (Fig. 4). Asynchrony of temporal responses to varying environmental conditions between
 486 species has also been identified as a stabilizing factor (Loreau & de Mazancourt 2013).
 487 However, it should be noted that in our case, species asynchrony in monospecific stands was
 488 not related to stability (Fig. S4), indicating that intrinsic species-specific responses to
 489 environmental fluctuations observed in monospecific stands are not necessarily a good
 490 indicator of the stabilizing effect that emerges when species are mixed (Gross *et al.* 2014).
 491 The mixing of Scots pine and European beech therefore changes the intrinsic species
 492 responses to yearly environmental variations at community level in comparison to
 493 monospecific stands, and temporal shifts in species interactions linked to temporal niche
 494 complementarity seem to play a key role in this change. Previous studies concerning forests
 495 have reported changes in the growth response to extreme droughts between mixed and
 496 monospecific stands (Lebourgeois *et al.* 2013; Pretzsch *et al.* 2013), although the results
 497 depended on species composition (Merlin *et al.* 2015; Grossiord *et al.* 2014). Nevertheless,
 498 those studies were either mainly based on tree level growth analyses or made no attempt to
 499 link the tree and community level analyses. Our results indicate that the changes in species
 500 asynchrony between mixed and monospecific stands were considerably lower at tree than at
 501 community level, but also that the asynchronies at the two levels were correlated, the latter
 502 suggesting that differences in species specific responses to variability in environmental
 503 conditions may also affect temporal stability. These results underline the need for further
 504 studies at community level and the importance of linking both levels.

505 The asynchrony-overyielding relationship identified in this study (Fig. 5) suggests that
 506 temporal niche complementarity is one of the most important mechanisms driving
 507 overyielding in this mixture. These results contradict the hypothesis stated by Jucker *et al.*
 508 (2014), who argued that asynchrony might not influence overyielding because it would
 509 require a rapid response in forest dynamics to environmental conditions. However, our study
 510 assumed no diversity effect on mortality, although significant effects of mixing on tree
 511 mortality, self-thinning lines and stand density indices have been reported previously (Binkley
 512 1984, 2003; Condés & Río 2015; Pretzsch & Biber 2016; Woodall, Milles & Vissage 2005),
 513 and may influence overyielding as well as stability.

514 *Species interactions*

515 We found the higher temporal stability in mixed stands to be linked to shifts in species
 516 interactions that influenced the growth response of a given species to inter-annual
 517 environmental conditions. That is, the temporal variation in niche complementarity between
 518 species, which results in compensatory dynamics between species, is one of the main factors
 519 underlying the increase in temporal stability. These results provide an empirical corroboration
 520 of the simulation-based findings of Morin *et al.* (2014), which pointed to the greater
 521 importance of species interactions as opposed to species-specific differences in responses to
 522 environmental conditions. However, the temporal scale and the compensatory dynamics
 523 considered in the simulations are not directly comparable to our approach.

524 **Temporal stability and overyielding at different levels**

525 The different stabilizing effects of species mixing at different organizational levels are in
 526 accordance with theory-based expectations (Tilman 1999; Loreau & de Mazancourt 2013) and
 527 show that the general pattern found in diversity-temporal stability relationships at community
 528 level also occur in the case of mixed forests with two species. Generally, species diversity
 529 increases the temporal stability of productivity at community level, but a high variability in
 530 this effect was reported at population level (Jiang & Pu 2009). In our study, we found a
 531 stabilizing effect at community level, but a neutral effect at population level. This lack of any
 532 destabilizing effect at population level might be explained by the slower growth dynamics of
 533 forests along with the long periods that are often required before any change in relative
 534 species abundance occurs, this factor playing an important role in diversity-opulation
 535 stability (Roscher *et al.* 2011). Accordingly, a negative diversity effect on forest species
 536 stability was found by Morin *et al.* (2014) based on long-term simulations from a process-
 537 based succession model.

538 At population level, we found underyielding for pine and overyielding for beech when
 539 growing in the mixed stands. These changes in mean productivity in comparison to
 540 monospecific stands were also associated with comparable relative changes in the standard
 541 deviation, resulting in similar temporal stabilities. Nevertheless, mixing species resulted in a
 542 destabilizing effect on individual pines, mainly due to the lower mean productivity, whereas
 543 in the case of beech, a neutral effect was found. The differences between the population and
 544 individual-tree level responses for pine may be due to the fact that only dominant and
 545 codominant trees were explored at tree level. Temporal variation in tree growth is generally
 546 lower as tree size increases, as indicated by the increasing stability of beech with tree size,
 547 even within the dominant and codominant trees included in this study. Similarly, tree
 548 responses to drought can vary among trees of different social status within a stand (Martín-
 549 Benito *et al.* 2008).

550 Mixing effects that were evident at the mean tree or population levels do not necessarily have
 551 any far-reaching practical relevance at community level. Studies that apply an individual tree
 552 level approach may overlook any compensation effects at population or community levels and

553 lead to questionable predictions when the results from individual dominant trees were scaled
 554 up to community level responses. It is important to underline the possible mixing effect on
 555 size distributions (Pretzsch & Schütze 2014, 2015), which can be one cause of contrasting
 556 effects at different levels, and contribute to misleading results if not taken into account when
 557 up-scaling.

558 Our results clearly show that the behaviour of mixed species stands cannot be derived simply
 559 by assuming additive effects between the combined species (e.g., based on the traits or
 560 dynamics of the species in monocultures). Both the overyielding of mixed-species stands at
 561 community level and the differences in growth stability at the community, population, and
 562 individual tree levels point to a multiplicative character of mixing effects. Modelling
 563 approaches cannot derive mixed stand dynamics from the weighted mean of the respective
 564 monocultures and should be able to reproduce the spatial and temporal inter-specific
 565 interactions between the combined species (Pretzsch, Forrester & Rötzer 2015).

566 **Environmental drivers**

567 The experimental design of our study was originally developed to examine whether the
 568 temporal variability of productivity in monocultures and mixed species stands is higher at
 569 sites with lower mean water supply. Many dendrochronological studies suggest that trees at
 570 drought prone sites may frequently suffer water limitation and therefore present more distinct
 571 fluctuations between high- and low-growth years (Fritts 2001). However, we found no
 572 statistical effect of precipitation or de Martonne aridity index on the temporal stability of
 573 productivity. This finding may be due to the typical lack of *ceteris paribus* conditions in field
 574 experiments, such that many factors may change along the transect other than the water
 575 supply and humidity. These factors could modify the effect of water supply and confound any
 576 productivity-water relationship. Indeed, the high variability in species asynchrony observed in
 577 monospecific stands along the transect at both stand and mean tree levels (from negative
 578 values to almost one), suggests that different environmental factors might be influencing
 579 species-specific growth at the different sites. Similarly, species over- or under-yielding (RPP_i)
 580 were not correlated, indicating that different environmental factors influence the mixing effect
 581 for each species.

582 Few studies have quantified the effects of European beech and Scots pine interactions on
 583 water, light or nutrient availability, uptake or use-efficiencies. In the same plots as those used
 584 in this study, the RP for light absorption at stand level generally increased due to a
 585 combination of more stratified canopy structures, changes in diameter-crown allometric
 586 relationships and increases in mean tree size in the mixtures (Forrester *et al.* in prep). Water-
 587 related interactions may also play a role as a result of inter-specific differences in interception
 588 (Nihlgård 1970; Augusto *et al.* 2002; Gerrits, Pfister & Savenije. 2010; Staelens *et al.* 2006;
 589 Van Nevel 2015), the isohydric behavior of pine *vs.* the anisohydric behavior of beech
 590 (Hartman 2011) and contrasting vertical root distributions and litter layers (Bonnemann 1939;
 591 Heinsdorf 1999; Knapp 1991), which may influence the vertical profile of water availability
 592 and uptake. These differences could improve nutrient availability in the mixtures compared
 593 with the pine monocultures. The seasonality of resource-use by a given species can also be
 594 modified by mixing, as shown for transpiration and light (Forrester *et al.* 2010; Sapjanskas *et*
 595 *al.* 2014). Further studies on the water and nutrient pools and fluxes might be required to
 596 determine their contribution to the temporal niche complementarity effects in these pine and
 597 beech mixtures.

598 **Concluding remarks**

599 Spatial and temporal species' complementarity in structure or functioning seems to be
 600 essential to increase the level and stability of productivity in mixed compared with

601 monospecific stands. In our two-species mixture, species asynchrony in mixed stands
 602 improved the level and stability of productivity , while our results with regard to temporal
 603 shifts in species interactions highlight the role of temporal niche complementarity in the
 604 stabilizing process. This species assemblage may provide a model example for other
 605 widespread species combinations as regards the degree of spatial and temporal
 606 complementarity. Other common conifer-broadleaved mixtures of early and late successional
 607 species or shade intolerant and tolerant species may behave similarly in terms of level and
 608 stability of productivity. We found the stability of productivity to be superior at most of the
 609 sites, regardless of the water supply and humidity, suggesting that the stabilization results
 610 from various complementarity effects together.

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616 **Data accessibility**

617 Data available from the Dryad Digital Repository

618 **References**

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803 **Supporting Information**

804 Additional Supporting Information may be found in the online version of this article:

805 **Table S1.** Overview of the 31 mixed *Pinus sylvestris-Fagus sylvatica* triplets included in this
806 analysis807 **Table S2.** Stand characteristics of monospecific and mixed-species stands of the triplets.808 **Table S3.** Description of the mean, standard deviation and stability of the annual basal area
809 increment at the different organizational levels observed in monospecific and mixed-species
810 stands.811 **Table S4.** Fixed effect results at stand level, species level and individual tree level for the
812 prediction of temporal stability, mean and standard deviation of annual basal area increment.

813 **Figure S1.** Example of the process of standardization and analysis of temporal variation in
 814 species interactions

815 **Figure S2.** Relationship between the mixing effect on stability and overyielding

816 **Figure S3.** Relationship between the coefficient of correlations of species stand basal area
 817 increments at community level in mixed and monospecific stands

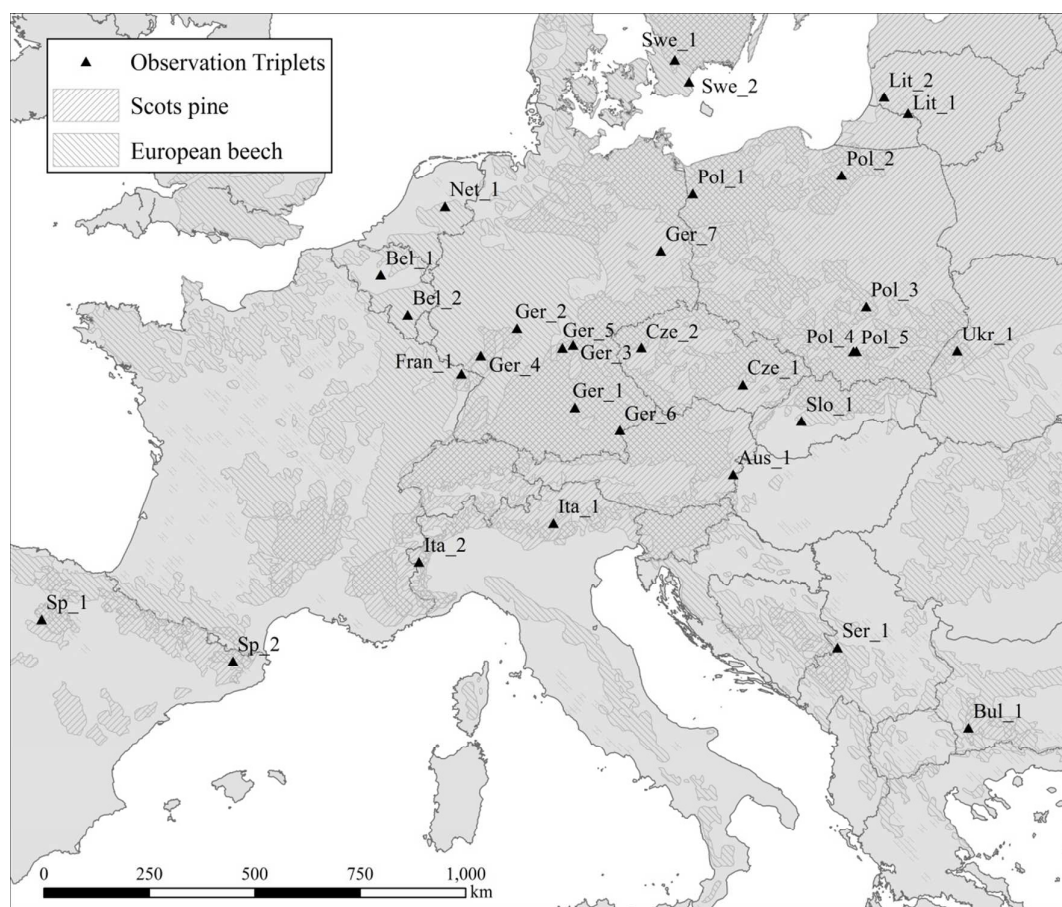
818 **Figure S4.** Relationships between temporal stability of stand basal area increment in mixed
 819 stands and species asynchrony in mixed and monospecific stands

820 **Figure S5.** Relationship between the mixing effect on temporal stability at community level
 821 and species asynchrony in mixed stands

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824 **Figures**

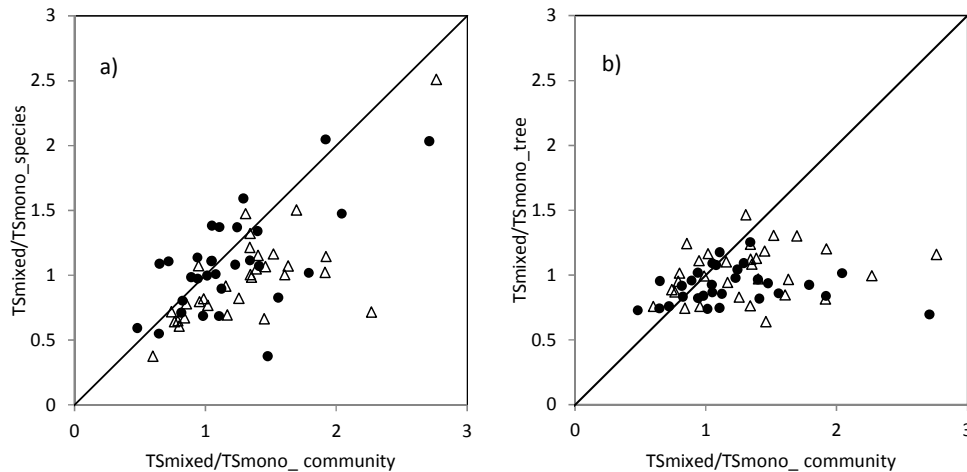


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826 Fig 1. Location of the 31 triplets of monospecific and mixed stands of Scots pine and
 827 European beech over the distribution of *Pinus sylvestris* and *Fagus sylvatica* according to
 828 EUFORGEN (<http://www.euforgen.org/distribution-maps/>)

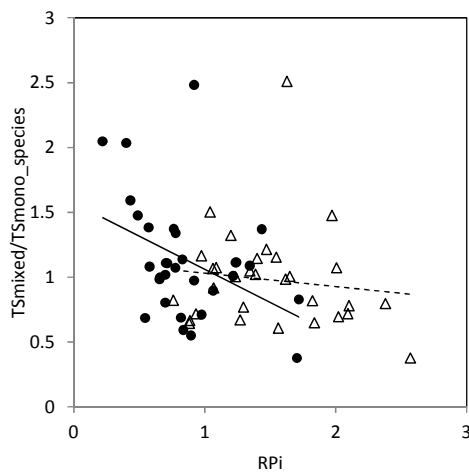
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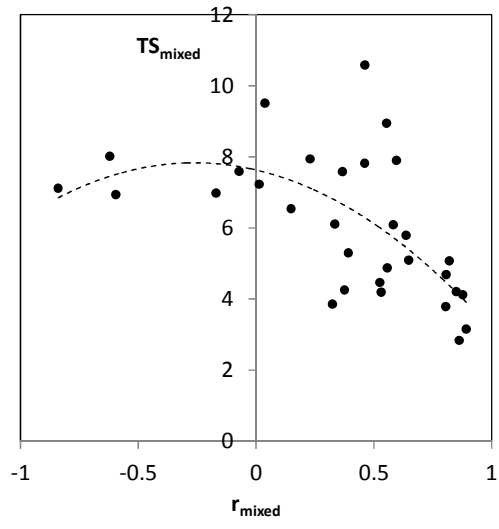
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832 Fig 2. Relationship between mixing effects on temporal stability in basal area increment
 833 (TS_{mixed}/TS_{mono}) at different organizational levels for *F. sylvatica* (white triangles) and *P.*
 834 *sylvestris* (black circles); a) species vs. community levels; b) individual tree vs. community
 835 levels.



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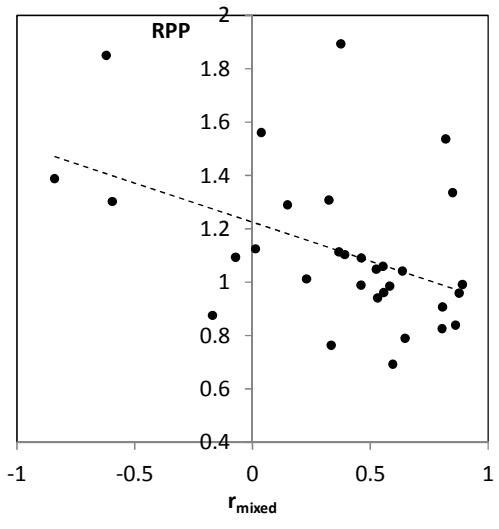
837 Fig. 3. Relationship between mixing effects on temporal stability in basal area increment at
 838 species level (TS_{mixed}/TS_{mono}) and relative productivity (R_{pi}) for *F. sylvatica* (white triangles)
 839 and *P. sylvestris* (black circles). Straight lines are the linear trend lines, dashed for *F.*
 840 *sylvatica* (NS) and continue for *P. sylvestris* ($R^2=0.17$; $P=0.023$)



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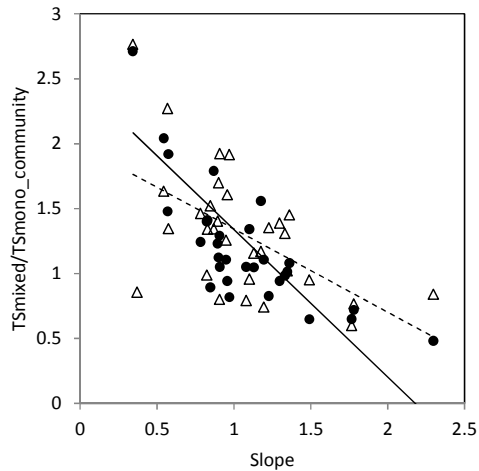
842 Fig 4. Temporal stability in stand basal area increment (TS_{mixed}) as a function of the
843 coefficient of correlation between species increments in mixed stands (r_{mixed}) ($R^2=0.40$;
844 $P<0.001$).

845



846

847 Fig 5. Relationship between overyielding (RPP) and the coefficient of correlation between
848 species increments in mixed stands (r_{mixed}) ($R^2=0.20$; $P=0.011$).



849

850 Fig. 6. Relationship between mixing effects on temporal stability in basal area increment
 851 ($TS_{\text{mixed}}/TS_{\text{mono}}$) at community level and slope of the major regression between observed and
 852 reference stand basal area growth indices in mixed stands ($IBAI_{\text{mixed}}=a+b \cdot IBAI_{\text{reference}}$; see text
 853 and Fig. S1 for additional information) for *F. sylvatica* (white triangles) and *P. sylvestris*
 854 (black circles). Straight lines are the linear trend lines, dashed for beech ($R^2=0.32$; $P=0.001$)
 855 and continue for pine ($R^2=0.53$; $P<0.001$).