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1 **Wood distillate as an alternative bio-based product against lichens on sandstone**

2 Elisabetta Bianchi <sup>a</sup>, Renato Benesperi <sup>b</sup>, Paolo Giordani <sup>c</sup>, Luca Martire <sup>d</sup>, Sergio Enrico Favero Longo  
3 <sup>e,\*</sup>, Stefano Loppi <sup>a</sup>

4 <sup>a</sup> Dipartimento di Scienze della Vita, Università di Siena, Via Mattioli 4, 53100, Siena, Italy;  
5 [elisabetta.bianchi@unisi.it](mailto:elisabetta.bianchi@unisi.it), [stefano.loppi@unisi.it](mailto:stefano.loppi@unisi.it)

6 <sup>b</sup> Dipartimento di Biologia, Università di Firenze, Via La Pira 4, 50121, Firenze, Italy;  
7 [renato.benesperi@unifi.it](mailto:renato.benesperi@unifi.it)

8 <sup>c</sup> Dipartimento di Farmacia, Università di Genova, Viale Cembrano 4, 16148, Genova, Italy;  
9 [giordani@difar.unige.it](mailto:giordani@difar.unige.it)

10 <sup>d</sup> Dipartimento di Scienze della Terra, Via Valperga Caluso 35, Università di Torino, 10125, Torino,  
11 Italy; [luca.martire@unito.it](mailto:luca.martire@unito.it)

12 <sup>e</sup> Dipartimento di Scienze della Vita e Biologia dei Sistemi, Università di Torino, Viale Mattioli 25,  
13 10125, Torino, Italy; [sergio.favero@unito.it](mailto:sergio.favero@unito.it)

14

15 \* Corresponding author

16 Sergio E. Favero Longo, PhD  
17 Università degli Studi di Torino  
18 Dipartimento di Scienze della Vita e Biologia dei Sistemi  
19 Viale Mattioli 25, 10125 Torino, Italy  
20 Tel. +390116705972  
21 Fax +390116705962  
22 [sergio.favero@unito.it](mailto:sergio.favero@unito.it)  
23 [orcid.org/0000-0001-7129-5975](https://orcid.org/0000-0001-7129-5975)

## 24 **Abstract**

25 The use of traditional biocides to halt or reduce biodeterioration is increasingly deterred, due to risks  
26 for human health and the environment, as well as for potential interference with stone materials.  
27 Alternative and eco-friendly substances are needed to limit these issues. Here we aim to evaluate the  
28 devitalization of lichens by a new bio-based product: wood distillate (also known as pyroligneous  
29 acid), a by-product of the use of plant biomass to produce bioenergy by pyrolysis without the addition  
30 of synthetic chemicals. We compared cellulose poultice applications of wood distillate at a  
31 concentration of 10% and two common chemical biocides against four epilithic lichen species on Pietra  
32 Serena, a sandstone widely used in Europe. The efficiency of devitalization was measured in terms of  
33 lichen vitality expressed by chlorophyll *a* fluorescence emission  $F_v/F_M$  and  $F_0$ . Furthermore, we  
34 evaluated the effects of wood distillate on physical properties of the stone material of relevance for its  
35 conservation, including colour, resistance to dissolution, and surface hardness. Wood distillate was as  
36 effective as chemical biocides in devitalizing the thalli and did not cause any relevant interference with  
37 the assayed sandstone, although a limited dissolution of its calcite cement was detected.

38

39 **Keywords:** CIELAB colour measurement; Circular economy; Devitalization; Lichen; Stone cleaning;  
40 Wood distillate.

41

## 42 **1. Introduction**

43 Natural and man-made building materials of ancient to contemporary exterior architectural structures  
44 are inevitably colonized by living organisms and susceptible to biodeterioration (Cwalina, 2011;  
45 Favero-Longo and Viles, 2020; Sanmartín et al., 2021). Among phototrophic colonizers, saxicolous  
46 (i.e. rock-dwelling) lichens play a primary role as agents of stone biodeterioration, causing aesthetic,  
47 chemical, and physical decay within a relatively short timescale (Caneva et al., 2008; Pinna, 2017). A  
48 debate is ongoing on the need to remove lichens, at least in cases where the deterioration effect is rather  
49 negligible and/or their presence positively contributes to the aesthetics or represents a biodiversity  
50 value (Pinna, 2014; Favero-Longo and Viles, 2020). A bioprotective effect has been even demonstrated  
51 for certain combinations of lichen species, rock substrates and climate conditions (Carter and Viles,  
52 2005; Pinna, 2021) Nevertheless, their removal is still generally considered necessary for the  
53 conservation of archaeological and monumental sites as well as in the maintenance of exterior surfaces

54 of every building (Seaward, 2015; Cappitelli et al., 2020). Although widely adopted in routine  
55 cleaning, the sole use of mechanical tools and (pressurized) water leave largely unremoved, and may  
56 even spread, lichen structures on and within the stone, and is thus usually followed by rapid  
57 recolonization dynamics (Pinna, 2017). In recent years, physical methods, such as laser treatments,  
58 have been proposed, showing promising results (e.g. Mascalchi et al., 2015; Sanz et al., 2015; Rivas et  
59 al., 2018). However, the optimization and practical applicability of these cleaning techniques are still  
60 pending, and some critical issue has emerged, including mineral melting and perceptible color change  
61 of stone surfaces (Sanmartín et al., 2019; Pozo-Antonio et al., 2019, 2022). Consequently, the effective  
62 practice most frequently used to remove lichens is based on their devitalization with chemical biocides  
63 (Kakakhel et al., 2019), followed by mechanical removal of thalli and, finally, the possible application  
64 of products (often the same biocides) aimed at limiting the recolonization (Pinna, 2017; Capitelli et al.,  
65 2020). However, biocides traditionally used in the control of biodeterioration, such as quaternary  
66 ammonium salts and isothiazolinones, give rise to increasing concerns due to health hazards,  
67 environmental persistence with consequent microbial adaptation, and/or potential nitrogen supply  
68 favouring recolonization (e.g. Bastian et al., 2012; Poursat et al., 2019; Silva et al., 2020).

69 In Europe, the use of biocidal products is regulated by the Regulation No 528/2012 of the European  
70 Parliament and of the Council (EU, 2012), which defines the implementation needed for the  
71 improvement of health and safety at work for humans and for the reduction of impacts on the  
72 environment. In particular, the removal or reduction of the use of hazardous products remains the  
73 primary objective in accordance with the United Nations' agenda for 2030 for sustainable development,  
74 namely the Third goal: "Ensuring health and security for all and for all ages" (United Nations, 2015). In  
75 addition, the EU's chemicals strategy for sustainability towards a toxic-free environment foresees a  
76 specific action dedicated to boosting the investment and innovative capacity for production and use of  
77 chemicals that are safe and sustainable by design, and throughout their life cycle (European  
78 Commission, 2020). The proposal of biocompatible and "eco-friendly" strategies to control  
79 biodeterioration has thus increased, also considering the application of plant-derived bioproducts  
80 instead of synthetic chemicals, although their natural origin does not necessarily imply that they are not  
81 toxic to humans and the environment (Lo Schiavo et al., 2020; Cappitelli and Villa, 2021). Many  
82 alternative products are mostly based on essential oils and secondary metabolites produced by plants  
83 against pathogens and predators (Palla et al., 2016; Caneva and Tescari, 2017; Jeong et al., 2018), and

84 their devitalizing effect on lichens has recently been investigated (Favero-Longo et al., 2021). Besides  
85 the devitalizing action, to exclude unacceptable corrosive or discolouring effects, a successful  
86 bioproduct should not interfere with the substrate (Pinna, 2017; Fidanza and Caneva, 2019).

87 Here we test a bio-based product, wood distillate (WD), also known as pyroligneous acid, a by-product  
88 of the use of plant biomass to produce bioenergy by pyrolysis. During this process, no synthetic  
89 chemical is added and only the physiological water present in the sapwood is used for the extraction  
90 and subsequent condensation (Wei et al., 2010; Mathew and Zakaria, 2015). Based on the different  
91 productive conditions (e.g. nature of the raw material, moisture content of biomass, temperature,  
92 contact time), WD may feature a different fine chemical composition, but the typical major constituents  
93 of WD are water, acetic acid, esters and phenolic compounds (Marumoto et al., 2012; Cai et al., 2012).  
94 In particular, the content of phenolic compounds, carbonyls and organic acids likely accounts for its  
95 known antimicrobial action (Velmurugan et al., 2009; Wei et al., 2010; Suresh et al., 2019). Anti-  
96 bacterial and insecticide effects have been shown at dilutions in the range 1:10-1:100 in deionized  
97 water (Mmojieje and Hornung, 2015) and 1:100 in 10 mM MgSO<sub>4</sub> (Misuri and Marri, 2021).

98 Although WD may be a promising biological alternative for the control of biodeterioration, to the best  
99 of our knowledge, the devitalization efficacy of WD against saxicolous lichens and its potential  
100 interference with stone materials have never been investigated. Hence, the aim of this study was to  
101 evaluate: i) the devitalization activity of WD using as reference two commercial chemical biocides  
102 widely used in Europe; ii) the interferences of WD with properties of relevance for the conservation of  
103 sandstone substrate, namely colour, resistance to dissolution, and surface hardness.

104

## 105 **2. Materials and Methods**

### 106 ***2.1 Sites and materials***

107 WD and biocide applications on lichens were carried out, *in situ*, at the Botanical Garden of the  
108 University of Siena [Site A, Siena, Italy: WGS84N 43.858537, E 11.303751; 322 m a.s.l.] and the park  
109 of Pratolino at Vaglia [Site B, Florence, Italy: N 43.859492, E 11.304735; 451 m a.s.l.] (Fig. 1a, e).  
110 Treatments were performed on the epilithic crustose-placodioid species *Protoparmeliopsis muralis*  
111 (Schreb.) M. Choisy, for both sites, and the epilithic crustose-areolate *Verrucaria nigrescens* Pers.,  
112 *Circinaria hoffmanniana* (S. Ekman and Fröberg ex R. Sant.) A. Nordin. and *Blastenia crenularia*

113 (With.) Arup, Søchting and Frödén, in the latter site only (Fig. S1). The selected species are common  
114 from the sub-Mediterranean to the montane belt of Italy on natural and man-made stone surfaces  
115 (Nimis et al., 2016).

116 In the first site, the substrate was a brick; in the second one the substrate was Pietra Serena, a sandstone  
117 lithology widely used in heritage sites as well as nowadays in Italy, composed of quartz with accessory  
118 plagioclase, calcite, K-feldspar, apatite, dolomite, and variable amounts of clay minerals and calcite  
119 cement as binders, and with an effective porosity around 3-5% (Fratini et al., 2014). Slabs of such rock  
120 material were obtained from a stone shop in Florence (Cosi & Bechelli S.N.C., Pontassieve) and used  
121 to assess the potential interference of WD with properties of relevance for conservation. The  
122 composition of the slabs was confirmed by X-ray powder diffraction, displaying quartz and subordinate  
123 calcite and plagioclase (Fig. S2); their effective porosity was estimated around 3.5% by water  
124 absorption under vacuum (Robin et al., 2016).

125 A chestnut (*Castanea sativa* Mill.) wood distillate (WD) produced in Val di Chiana (Arezzo, Italy) by  
126 Esperia s.r.l. (RM Group Energy Solutions) and distributed by BioDea© was used at different dilutions,  
127 as subsequently detailed for each experiment. Our available data indicate a composition in line with the  
128 general richness of hundreds of organic compounds of this product, including acetic acid as main  
129 constituent (up to 30%) and several phenols, polyphenols, and tannins, with very low concentrations of  
130 toxic compounds such as PAHs and PCBs and trace elements like As and Cr (Wei et al., 2010;  
131 Filippelli et al., 2021). The following compounds were selected as reference chemical biocides: (i)  
132 benzalkonium chloride (BAC), prepared as 3% water solution of Preventol RI50 (alkyl dimethyl benzyl  
133 ammonium chloride, approx 50%, and isopropyl alcohol, 2%, in water; Lanxess, Köln, Germany), and  
134 (ii) N-octyl-isothiazolinone and didecyl-dimethyl ammonium chloride (OIT-DDAC), prepared as 3.0%  
135 solution of BiotinT (OIT, 7-10%, DDAC, 40-60%, formic acid 2.0-2.5%, and isopropyl alcohol, 15-  
136 20%, in water; CTS, Altavilla Vicentina, Italy). A bottled water with low mineral content (fixed residue  
137 at 180°C of 22-43 mg L<sup>-1</sup>; Fonti di Vinadio, Vinadio, Italy) was used through the experiments, both for  
138 the dilutions and as negative control.

139

140

141

## 142 **2.2 Wood distillate and biocides applications**

143 A preliminary WD dose-effect experiment was carried out at site A, on bricks colonized by thalli of *P.*  
144 *muralis* (Fig. 1b-d). Different wood distillate concentrations (0.50%, 0.75%, 1%, 5%, 10%), selected  
145 with reference to the range of effectiveness reported against other biological targets (e.g. Mmojieje and  
146 Hornung, 2015; Misuri and Marri, 2021), were applied with a cellulose poultice (Arbocel BC 1000, JR  
147 Pharma, Rosenberg, Germany), approx. 1 cm thick, containing ca. 12 mL cm<sup>-3</sup> of solutions, after  
148 having moistened the thalli with bottled water. To preserve the humidity, the cellulose poultice was  
149 covered with a sheet of aluminium foil for 4 hours and was later gently removed; thereafter, thalli were  
150 moistened with bottled water again (Favero-Longo et al., 2020).

151 To assess the species-specific effectiveness in comparison with traditional biocides, WD at the  
152 concentration of 10% -selected on the basis of the results of the above-described preliminary  
153 experiment (Table S1; Fig. 2)-, BAC and OIT-DDAC were applied at site B, on the horizontal  
154 sandstone balustrades of a monumental stairway (Fig. 1f-h). All the products were applied following  
155 the same protocol adopted at site A.

156

## 157 **2.3 Lichen vitality**

158 In both experiments, five thalli (statistical replicates) for each treatment were examined for their  
159 photosynthetic performance as target of the devitalization effectiveness. The vitality of lichens was  
160 checked by measurements of chlorophyll *a* fluorescence (Chl<sub>a</sub>F), using a Handy-PEA fluorimeter  
161 (Plant Efficiency Analyser, Hansatech instruments Ltd., Norfolk, England). Measurements were  
162 performed on dark-adapted moistened thalli, previously humidified by bottled water, and covered with  
163 a black cotton fabric. Five measurements were taken for each thallus, positioning the sensor head at 90°  
164 over its surface, inducing Chl<sub>a</sub>F by a red light (peak at 650 nm), and recording the data after a  
165 saturating light pulse (3000 μmol s<sup>-1</sup> m<sup>-2</sup>) of 1s (Bianchi et al., 2019). At site A, analyses were carried  
166 out 4 and 16 hours after the application of the wood distillate, to screen for the potential of the different  
167 dilutions on the short term. At site B, the analyses were performed one day (T1) and fifteen days (T2)  
168 after the treatments, to assess short term effects and the potential recovery after a couple of weeks  
169 (Tretiach et al., 2012; Favero-Longo et al., 2017). The maximum quantum efficiency of PSII, that is  
170  $F_V/F_M$  (where  $F_V = F_M - F_0$ ), where  $F_V$  is the variable fluorescence yield,  $F_M$  is the maximal fluorescence

171 yield and  $F_0$  is the minimal fluorescence yield, was calculated (van Kooten and Snel, 1990). According  
172 to previous research on the effectiveness of biocidal treatments against lichens (e.g., Tretiach et al.,  
173 2012; Favero-Longo et al., 2020) the maximum quantum efficiency of PSII and variations in  $F_0$ , related  
174 to chlorophyll content of the light harvesting complex (Baruffo and Tretiach, 2007), were used to check  
175 the vitality of the thalli and PSII efficiency.

176

## 177 **2.4 Colour measurements**

178 In the laboratory, WD at the concentration of 10% was applied on the Pietra Serena sandstone slabs, cut  
179 to a dimension of 4×3×1 cm using a diamond saw. The application was performed with cellulose  
180 poultice as previously described. Moreover, pure WD and tap water were also applied for comparison.  
181 The pH of pure and 10% solution of WD, measured using a pH-meter ORP - HI2002 (Hanna  
182 Instruments, Italy), were 3.2 and 3.6, respectively.

183 Colour measurements on sandstone (Pietra Serena) were carried out by a portable spectrophotometer  
184 (Konica Minolta CM-23d), under the following conditions: D65 illuminant, 2° observer and a target  
185 area of 8 mm diameter, following Prieto and colleagues (2010). The CIELAB colour system (CIE,  
186 1986) was used to analyse the data: each colour is defined by three Cartesian or scalar coordinates: the  
187  $L^*$  parameter represents the lightness, ranging from 0 (absolute black) to 100 (absolute white);  $a^*$   
188 represents the chromatic variations from red to green; and  $b^*$  represents the chromatic variations from  
189 yellow to blue. To analyse the colour after WD treatment and after washing with bottled water, the total  
190 colour difference ( $\Delta E^*_{ab}$ ) was calculated as follows:

$$191 \Delta E^*_{ab} = [(\Delta L^*)^2 + (\Delta a^*)^2 + (\Delta b^*)^2]^{1/2},$$

192 where:  $\Delta L^* = L^*_i - L^*_0$ ;  $\Delta a^* = a^*_i - a^*_0$ ;  $\Delta b^* = b^*_i - b^*_0$ , the subscript i denotes the colour parameter  
193 after 4 hours and after washing, and the subscript 0 denotes the colour parameter at the beginning of  
194 experiment. Seven measurements were made directly on the threatened surface of three slabs per  
195 treatment ( $n=21$ ).

196 Saturation [ $C^*_{ab}=(a^{*2}+b^{*2})^{1/2}$ ] and hue [ $h^*_{ab}=\arctan(b^*/a^*)$ ], together with their respective  $\Delta$  values,  
197 were also calculated.



198 To perceive differences in colour, the following ranges were considered based on Mokrzycki and Tatol  
199 (2012): E1:  $0 < \Delta E^*_{ab} < 1$  CIELAB units, observer does not notice the difference; E2:  $1 < \Delta E^*_{ab} < 2$   
200 CIELAB units, only experienced observer can notice the difference; E3:  $2 < \Delta E^*_{ab} < 3.5$  CIELAB  
201 units, unexperienced observer also notices the difference; E4:  $3.5 < \Delta E^*_{ab} < 5$  CIELAB units, clear  
202 difference in colour is noticed; E5:  $\Delta E^*_{ab} > 5$  CIELAB units, observer notices two different colours.

203

## 204 ***2.5 Resistance to dissolution***

205 To evaluate the impact of the WD acidity on the durability of Pietra Serena sandstone, incubations  
206 assays were performed. Four slabs (4×3×1 cm) were weighted with a Kern EG420-3NM (Kern and  
207 Sohn GmbH, Balingen, Germany) before and after their immersion for 4 hours in a static 10% WD  
208 solution and their subsequent drying on a heating plate, thus quantifying the acidolysis-driven mass  
209 loss. Slab immersion in distilled deionized water was carried out as negative control.

210 To evaluate the persistence of mineral constituents, observations of the surface and cross sections of the  
211 slabs were performed using cathodoluminescence microscopy (CL). CL was carried out using a CITL  
212 8200 mk3 equipment (operating conditions of about 17 kV and 400  $\mu$ A). Moreover, CL observations  
213 were also performed on slabs incubated in stirred 10% solutions of WD and on slabs on which the 10%  
214 solution of WD was applied with cellulose poultice.

215

## 216 ***2.6 Surface hardness measurements***

217 Hardness measurements were carried out on Pietra Serena sandstone slabs (9×8×4 cm), before and after  
218 the application of WD at the concentration 10% with the cellulose poultice ( $n=3$ ). Stone surface  
219 hardness was measured using a Proceq Equotip Piccolo 2, DL-type (Proceq, Switzerland). A  
220 combination of two measuring procedures [single impact method (SIM) and repeated impact method  
221 (RIM)] was adopted for each stone sample measuring area (Yilmaz, 2013; Wilhelm et al., 2016).  
222 Firstly, to evaluate the elastic and plastic properties of the rock surfaces after the treatment, a series of  
223 45 randomly distributed readings (SIM) was carried out. Furthermore, a second series of 20 repeated  
224 measurements (RIM) on ten points was taken to characterize the elastic and plastic properties of the  
225 surface and subsurface of the rock, informative on strength characteristics such as the consolidation of

226 mineral grains, the looseness of the original rock surface, and the degree of compaction due to repeated  
227 impacts (Aoki and Matsukura, 2007). The robust hybrid dynamic hardness ( $\text{HDH}_{\text{robust}}$ , sensu Wilhelm  
228 et al., 2016) was calculated as follows:

$$229 \text{HDH}_{\text{robust}} = (\text{HLDL}_{\text{S:med}})^2 / (\text{HLDL}_{\text{R:med}})$$

230 where:  $\text{HLDL}_{\text{S:med}}$  is the median value of the SIM series and  $\text{HLDL}_{\text{R:med}}$  represents the median of the  
231 maximum values of the ten RIM series.

232

## 233 **2.7 Data analysis**

234 A dose-response regression model was applied to describe the effect of WD at different concentrations  
235 on photobiont vitality of *P. muralis* thalli at Site A, using WD concentration as independent variable  
236 and the effect on  $F_V/F_M$  values as dependent variable. The effective dose was checked using the *drc* R  
237 package version 3.0-1 (Ritz et al., 2015). In *drc* the function ED was used to calculate arbitrary  
238 effective dose values ED10, ED50, ED90 and ED95 based on the model fit, where 95% confidence  
239 intervals are obtained using the delta method. A logistic curve was used to describe the response of  
240 fluorescence measurements against WD doses. A Linear Mixed Effect Model (LMEM), as a Repeated  
241 Measurement ANOVA design, was applied for each lichen species to describe the effects of the WD  
242 and biocide treatments on photobiont vitality ( $F_V/F_M$  and  $F_0$ ), using thallus identity as a random effect  
243 factor.  $F_V/F_M$  and  $F_0$  were used as response variables and treatment (WD and traditional biocides),  
244 species, and time as explanatory variables in a full factorial design. We evaluated the significance of  
245 the fixed effects and of associated interaction factors using a type III ANOVA, using the Satterthwaite  
246 approximation. For each analysis, data normality of the residuals was checked with the Shapiro-Wilk  
247 test. LMEM computations were performed using the *lmer* function of the *lmerTest* R package version  
248 3.1-3 for fitting the models. The means and standard deviations (SD) of colours and the medians of  
249 hardness measurements on sandstone were checked by one-way ANOVA and Tukey post hoc test was  
250 used for a post-hoc comparison of individual means in all analysis (with at least  $p < 0.05$  as the  
251 significance level). The analysis was run using the statistical program R Core Team (2021).

252

253

## 254 **3. Results**

### 255 **3.1 Efficacy of devitalization treatments in situ**

256 Wood distillate at the concentration of 10% was effective at devitalizing 95% (ED95) of thalli of *P.*  
257 *muralis* (Table S1) almost zeroing  $F_V/F_M$  values after 4 hours from treatment at site A, with no  
258 recovery being observed after 16 hours (Fig. 2).

259 Treatments at site B with 10% WD, BAC and OIT-DDAC induced several physiological alterations in  
260 all species over time as shown by the results of LMEM analysis (Table S2a).  $F_V/F_M$  values of all  
261 species were significantly lower than those of controls both at T1 and T2 (Fig. 3), showing values  
262 below the viability threshold of 0.15 (Favero-Longo et al., 2017). Only in the cases of application of  
263 OIT-DDAC on *B. crenularia* and BAC on *P. muralis* at T2,  $F_V/F_M$  values showed a partial recovery  
264 over the viability threshold compared to the other treatments (mean $\pm$ SD: 0.244 $\pm$ 0.18 and 0.162 $\pm$ 0.12,  
265 respectively; Fig. 3). At T2, the physiological parameters of all species treated with WD showed the  
266 strongest decrease, with  $F_V/F_M$  of WD-treated *B. crenularia* and *C. hoffmanniana* significantly  
267 ( $p < 0.05$ ) lower than values reached with traditional biocides (Fig. 3).

268 After all treatments,  $F_0$  values changed significantly over time in all species (Table S2b). Upon WD  
269 treatment, at T2,  $F_0$  values of all species were significantly ( $p < 0.05$ ) lower than those observed in  
270 controls (Fig. 4), but significantly ( $p < 0.05$ ) higher than those obtained with traditional biocides (Fig. 4).

271

### 272 **3.2 Effects of wood distillate on the properties of sandstone**

273 Table 1 reports the quantitative description of surface colourimetry measures on pure and 10% WD  
274 treated Pietra Serena sandstone and of those treated with tap water (TW) as control. Parameter  $\Delta L^*$ ,  
275 related to the lightness of the colour, indicated a general darkening of stone samples after both pure (-  
276 4.25 CIELAB units) and 10% (-2.35 CIELAB units) WD applications with respect to TW (-1.27  
277 CIELAB units). The washing step caused a significant reduction in darkening in stone samples treated  
278 with pure WD changing  $\Delta L^*$  to -2.51 CIELAB units while the values of those treated with TW and  
279 10% WD only showed a slight variation to -1.84 and -2.79 CIELAB units, respectively. About the  
280 parameter  $\Delta a^*$ , associated with greenness (-) – redness (+) changes, samples treated with TW showed  
281 values around zero, while samples treated with WD showed similar values around 0.50 CIELAB units.  
282 In both cases,  $\Delta a^*$  after the washing step slightly lowered to 0.35 and 0.42 CIELAB units respectively.

283  $\Delta b^*$  values, associated with blueness (-) – yellowness (+) changes, increased to 1.64 after pure WD and  
284 to 2.53 CIELAB units after 10% WD, significantly higher than the values of the samples treated with  
285 TW (around 0.20 CIELAB units). A significant reduction in yellowing began after the stone was  
286 cleaned with tap water (0.47 and 0.70 CIELAB units, respectively). Saturation ( $C^*_{ab}$ ) and hue ( $h_{ab}$ )  
287 showed higher variation after 1:10 WD application ( $\Delta C^*_{ab} = 2.42$ ,  $\Delta h^*_{ab} -14.82$  CIELAB units) than  
288 with pure WD ( $\Delta C^*_{ab} = 1.52$ ,  $\Delta h^*_{ab} -10.78$  CIELAB units), but a significant recovery towards the  
289 original values was observed for both the treatments after the washing step. The total colour change  
290 ( $\Delta E^*_{ab}$ ) after pure WD treatment was 4.58 CIELAB units (range E4; *sensu* Mokrzycki and Tatol, 2012)  
291 and significantly decreased to 2.58 after the washing step (E3) The  $\Delta E^*_{ab}$  in sandstone treated with  
292 10% WD was 3.51, and 2.93 CIELAB units (E3) after cleaning, while for samples treated with TW it  
293 was 1.28 CIELAB units and reached 1.84 after the washing step (E2). In each case,  $\Delta E^*_{ab}$  parameter  
294 did not exceed the threshold of 5 CIELAB units (E5).

295 Changes were not observed in the weight of sandstone samples soaked in static 10% WD solution (-  
296  $0.05\% \pm 0.02$ ) compared to the control ones soaked in water ( $-0.04 \pm 0.03\%$ ; Table S3). Nevertheless, CL  
297 observations showed partial dissolution of calcite cement at the surface, recognizable in terms of loss of  
298 orange luminescence signal with respect to controls in deionized water (Fig. 5a-b). Similarly, the slabs  
299 treated with 10% WD applied with cellulose poultice showed a partial calcite dissolution (Fig. 5c),  
300 while in the slabs incubated in the stirred solution the calcite cement at the surface completely  
301 disappeared (Fig. 5a). However, no treatment determined a microscopically detectable dissolution  
302 through the cross-sectioned slab profiles (Fig. 5d-g).

303 Equotip measurements showed that 10% WD applied with poultice cellulose did not affect the Pietra  
304 Serena sandstone surfaces showing also no significant variation compared with TW-treated slabs (Fig.  
305 S3).

#### 306 **4. Discussion**

307 The search for innovative natural products to devitalize biodeteriogens is one of the hottest areas of  
308 interest for the cleaning of stone, particularly in the field of cultural heritage, and has led to several  
309 studies exploring their potential application (e.g. Lo Schiavo et al., 2020; Cappitelli et al., 2020). In this  
310 work, we showed that WD effectively devitalizes thalli of different lichen species, which are  
311 commonly found on architectural stone surfaces, showing results comparable with those obtained with  
312 conventional chemical biocides based on QACs and OIT. Such effectiveness was demonstrated when

313 WD was applied at a concentration of 10%, whereas more diluted concentrations were less effective for  
314 the devitalization of vitality parameters of *P. muralis*, although this lichen species has somewhat been  
315 recognized as less resistant to biocide treatments than others (Favero-Longo et al., 2017). A 10% WD  
316 treatment has been shown to have an inhibition effect against insects and some gram-positive bacteria  
317 (Lee et al., 2010; Mmojieje and Hornung, 2015), but other laboratory studies showed that even lower  
318 concentrations of 1-2% may be sufficient to inhibit the growth of some fungi and bacteria (Jung et al.,  
319 2007; Lee et al., 2010; Misuri and Marri, 2021). With this regard, the poor sensitivity of *P. muralis* to  
320 5% and lower WD concentrations may depend on a higher resistance of saxicolous lichens with respect  
321 to other (micro-)organisms, but also on the fact that our experiments were run in the field under real  
322 conditions.

323 With the application of 10% WD no signs of recovery were seen after 15 days from treatment,  
324 consistently with cellulose poultice application of QACs and OIT-based traditional biocides (Favero-  
325 Longo et al., 2017). In addition, in the case of *B. crenularia* and *C. hoffmanniana*, the best  
326 devitalization performance was obtained with WD, which always maintained  $F_V/F_M$  below the viability  
327 threshold set at 0.15.

328 Some difference between traditional biocides and WD in  $F_V/F_M$  and  $F_0$  at T1 and T2 may be suggestive  
329 of a different toxicity process. In the case of BAC and OIT-DDAC, the zeroing of  $F_V/F_M$  at T1 was  
330 associated with a remarkable, but incomplete, drop of  $F_0$ , which instead more remarkably decreased at  
331 T2. Such pattern is compatible with a loss of chlorophyll determined by damage to thylakoidal  
332 membrane caused by QACs, as BAC and DDAC (Wessels and Ingmer, 2013). However, the non-  
333 disrupted photobionts maintained some potential for recovery, as indicated by the slight increase of  
334  $F_V/F_M$  at T2 with respect to T0 observed for all the species here, in agreement with the observations of  
335 e.g., Tretiach et al. (2012). Differently, in the case of WD,  $F_0$  showed a slight increase at T2 with  
336 respect to T1 (except for the less resistant lichen *P. muralis*), which is the same pattern observed for  
337 controls, likely related to different environmental conditions at the time of measuring.  $F_V/F_M$  values did  
338 not show any recovery, suggesting that the photobionts were fully inactivated. This pattern suggests an  
339 inhibitory effect on the physiological functionality of the photobiont, which is compatible with the high  
340 content of (poly-)phenolic compounds in WD and their known effects on other (micro-)organisms  
341 (Pimenta et al., 2018). This does not exclude a WD-driven structural damage influencing the  
342 chlorophyll content, which is likely prominent in the case of *P. muralis*. The identification of WD  
343 components responsible for its devitalization effectiveness, which is currently a shared objective in the

344 case of its potential herbicidal, fungicidal, and bactericidal properties considered of valuable interest  
345 for agriculture (e.g., Pimenta et al., 2018; Aguirre et al., 2020), will be a future research step also in the  
346 case of lichens. Although the hazard profile of WD for humans and the environment similarly goes far  
347 beyond the aim of this study, our unpublished data indicate that levels of substances of possible  
348 toxicological concern such as PAHs and PCBs, as well as some toxic trace element like As and Cr, are  
349 very low in the used WD. Moreover, based on the effects on different human cell lines, a safe profile of  
350 WD emerged for short time usage, but caution is necessary following persistent product exposure  
351 (Filippelli et al., 2021).

352

353 In this work we also surveyed the interference of WD with the physical properties of Pietra Serena.  
354 Despite the dark brown colour of pure WD, the treated stone surfaces did not show important changes  
355 in the chromatic appearance, being in the range of those noticed only by experienced observers  
356 (Mokrzycki and Tatol, 2011). In particular, the  $\Delta E^*_{ab}$  values for slabs treated with 10% WD (and even  
357 pure) decreased after the application and subsequent washing with blotted water to values in the E2-E3  
358 ranges (sensu Mokrzycki and Tatol, 2011). They were thus well beneath the threshold of 5.0 CIELAB  
359 units over which an observer perceives two different colours and which is the normal limit of  
360 perception considered in industrial or technical applications (Palazzi, 1995; Eyssautier-Chuine et al.,  
361 2016). They were also beneath the more stringent threshold of 3.0 CIELAB units recently considered  
362 as upper limit of rigorous colour tolerance or noticeable change in colour following cleaning  
363 intervention on stone heritage surfaces (Sanmartín et al., 2020), and may be acceptable according to the  
364 Italian guidelines for the restoration of stone buildings (Bergamonti et al., 2018).

365 Incubation assays showed that WD, despite its low pH, did not determine any remarkable weight loss  
366 of the assayed sandstone slabs due to acidolysis, although the expected dissolution of calcite was  
367 detected with CL observations. Indeed, calcite dissolution appeared a remarkable feature at the surface  
368 of slabs incubated in stirred solutions, but poorly detectable in their interior. The phenomenon was  
369 even more contained for both the static incubation and the, similarly static, poultice application.  
370 Accordingly, calcite dissolution in static conditions is remarkably lower than in flowing solutions,  
371 particularly in the case of rock structural features limiting a reactive fluid infiltration (e.g. Brand et al.  
372 2017; Pearce et al. 2019). Therefore, combination of single and repeated impact measures with Equotip  
373 did not show changes in Pietra Serena sandstone after the poultice application of 10% WD, compared  
374 with those treated with tap water. Such a stability of hybrid dynamic hardness, which is informative on

375 elastic and plastic properties of stone surfaces and sub surfaces and is a proxy of open porosity  
376 (Wilhelm et al., 2016), further accounts for a low impact of WD on the physical durability of the  
377 examined sandstone material.

378 By summarizing negative and positive issues, the low pH and the consequent (expected) dissolution of  
379 calcite cement in the examined sandstone, although contained by the static application, reasonably  
380 discourage the WD application on sculpted surfaces and fine details of the stone cultural heritage.  
381 However, the high devitalization efficacy and the limited impact on physical properties of the  
382 examined sandstone as colour and surface hardness may be compatible with the WD treatment of less  
383 delicate architectural elements, as pavements or unrefined stone blocks, overcoming emerging  
384 drawbacks of traditional biocidal products (see section 1) and the pending technical limits of physical  
385 approaches. Such potency may be particularly supported if the WD treated experimental surfaces will  
386 show a long-term preservative effect of WD against recolonization processes (monitoring of the  
387 assayed parcels in progress), which may exclude the necessity of repeated applications.

388 Based on its acidity, more remarkable negative interferences may be instead expected on carbonate  
389 substrata, as marble and limestone. It is nevertheless remarkable and encouraging that the  
390 neutralization of WD does not affect its activity against biological targets others than lichens (Mmojieje  
391 and Hornung, 2015).

392

## 393 **5. Conclusions**

394 This work showed that 10% chestnut wood distillate is effective for the devitalization of lichens on  
395 sandstone surfaces, and that its application is compatible with keeping stone colour and surface  
396 hardness, and thus appears as a promising plant-based product for the control of biodeterioration on  
397 similar lithologies, although some dissolution of calcite cement suggests to exclude its application on  
398 delicate architectural elements.

399

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407

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596 **Tables**

597 **Table 1.** Values of CIELAB colour parameters ( $L^*$ ,  $a^*$ ,  $b^*$ ,  $C^*_{ab}$ ,  $h_{ab}$ ) and CIE total colour difference ( $\Delta L^*$ ,  $\Delta a^*$ ,  $\Delta b^*$ ,  $\Delta E^*_{ab}$ ,  $\Delta C^*_{ab}$ ,  $\Delta h^*_{ab}$ ) of Pietra  
 598 Serena sandstone after tap water (TW) and WD (1:1 and 1:10). For each treatment measurements were carried out before (T0) and after treatment  
 599 (After TW, After WD) and after washing step with blotted water (After WW). Different letters indicate statistically significant differences ( $p < 0.05$ ) for  
 600 each treatment (lowercase letters) and between treatments (uppercase letters). For  $\Delta E^*_{ab}$  values, E1-E5 ranges are indicated following Mokrzycki and  
 601 Tatol (2011).

602

CIELAB colour parameters											
	$L^*$	$a^*$	$b^*$	$C^*_{ab}$	$h^*_{ab}$	$\Delta L^*$	$\Delta a^*$	$\Delta b^*$	$\Delta C^*_{ab}$	$\Delta h^*_{ab}$	$\Delta E^*_{ab}$
<b>TW</b>											
<b>T0</b>	66.67 ± 1.42 a	-0.79 ± 0.04 a	1.92 ± 0.17 a	2.08 ± 0.15 a	112.70 ± 1.92 a						
<b>After TW</b>	65.40 ± 1.33 a	-0.79 ± 0.02 a	2.12 ± 0.19 a	2.26 ± 0.18 a	110.50 ± 1.91 a	-1.27 ± 1.11 a, A	0.00 ± 0.03 a, A	0.20 ± 0.18 a, A	0.18 ± 0.17 a, A	-2.13 ± 1.90 a, A	1.28 ± 1.08 a, A (E2)
<b>After WW</b>	64.83 ± 1.36 b	-0.82 ± 0.05 b	2.10 ± 0.14 a	2.26 ± 0.14 a	111.50 ± 2.44 a	-2.84 ± 1.52 a, A'	-0.03 ± 0.07 a, A'	0.18 ± 0.23 a, A'	0.18 ± 0.20 a, A'	-1.19 ± 3.59 a, A'	1.84 ± 1.45 a, A' (E2)
<b>WD1:1</b>											
<b>T0</b>	58.49 ± 1.16 a	-0.95 ± 0.03 a	3.36 ± 0.14 a	3.49 ± 0.13 a	105.94 ± 1.06 a						
<b>After WD</b>	54.24 ± 0.38 b	-0.44 ± 0.01 b	5.00 ± 0.10 b	5.02 ± 0.10 b	95.10 ± 0.31 b	-4.25 ± 0.86 a, B	0.51 ± 0.01 a, B	1.64 ± 0.14 a, B	1.52 ± 0.14 a, B	-10.78 ± 0.91 a, B	4.58 ± 0.78 a, B (E4)
<b>After WW</b>	56.00 ± 0.13 c	-0.60 ± 0.01 c	3.78 ± 0.06 c	3.83 ± 0.06 c	99.04 ± 0.34 c	-2.51 ± 1.08 b, A'	0.35 ± 0.01 b, B'	0.47 ± 0.11 b, A'	0.33 ± 0.10 b, A'	-6.89 ± 0.80 b, B'	2.58 ± 1.01 b, A' (E3)
<b>WD1:10</b>											
<b>T0</b>	61.61 ± 2.33 a	-0.97 ± 0.08 a	2.65 ± 0.45 a	2.78 ± 0.32 a	110.55 ± 4.24 a						
<b>After WD</b>	59.26 ± 2.00 a	-0.52 ± 0.07 b	5.18 ± 0.28 b	5.21 ± 0.27 b	95.70 ± 0.84 b	-2.35 ± 0.97 a, B	0.45 ± 0.08 a, B	2.53 ± 0.50 a, C	2.42 ± 0.39 a, C	-14.82 ± 3.65 a, C	3.51 ± 0.89 a, B (E4)
<b>After WW</b>	58.81 ± 1.55 b	-0.54 ± 0.04 b	3.35 ± 0.3 c	3.40 ± 0.30 c	99.31 ± 1.29 c	-2.79 ± 1.06 a, A'	0.42 ± 0.11 a, B'	0.70 ± 0.29 b, A'	0.61 ± 0.15 b, B'	-11.24 ± 3.95 a, B'	2.93 ± 0.95 a, A' (E3)

603



604 **Figure captions**

605 **Fig. 1.** Lichen devitalization assays in sites A (a-d) and B (e-h). (a) Bricks delimiting a flowerbed in the  
606 Botanical Garden of the University of Siena (Italy); (b-e) thallus of *Protoparmeliopsis muralis* (b) before, (c)  
607 during and (d) after the WD poultice application; (e) sandstone balustrade of a monumental stairway in the park  
608 of Pratolino at Vaglia (Florence, Italy); (f-h) parcels colonized by targeted lichen species (f) before, (g) during  
609 and (h) after the poultice application of WD and traditional biocides.

610 **Fig. 2.** Dose-response curves of WD at different concentrations (0%; 0.5%; 0.75%; 1%; 5%; 10%) in terms of  
611 variation in *P. muralis* maximum quantum efficiency of Photosystem II photochemistry ( $F_V/F_M$ ) after 4 hours (a)  
612 and after 16 hours (b) the treatment. See Table S1 for estimates, and the corresponding estimated standard errors,  
613 and possibly lower and upper confidence intervals to find the effective dose (ED) values.

614 **Fig. 3.** Maximum quantum efficiency of Photosystem II photochemistry ( $F_V/F_M$ ) variation in thalli of *B.*  
615 *crenularia*, *C. hoffmanniana*, *P. muralis* and *V. nigrescens* in site B measured 1-day (T1) and 15 (T2) days after  
616 the biocides application (Control: blotted water; OIT-DDAC; BAC and WD). See Table S2a for rANOVA  
617 results. For each species: lower case letters denote significant differences between biocides and control during  
618 each time; individual treatment differences over time were marked \*.

619 **Fig. 4.**  $F_0$  variations (scaled to the maximum relative to each species) in thalli of *B. crenularia*, *C. hoffmanniana*,  
620 *P. muralis* and *V. nigrescens* measured after 1-day (T1) and 15 days (T2) the biocides application in site B  
621 (Control: blotted water; OIT-DDAC; BAC and WD). See Table S2b for rANOVA results. For each species:  
622 lower case letters denote significant differences between biocides and control during each time; individual  
623 treatment differences over time were marked \*.

624 **Fig. 5.** Photomicrographs in cathodoluminescence (CL) of sandstone (Pietra Serena), with orange luminescence  
625 marking the presence and abundance of calcite cement and granules, distinguishable from other mineral phases  
626 (feldspars, blue; apatite, green; dolomite, red; quartz, no/poor luminescence). (a-c) Surface of slabs incubated (a)  
627 in water (negative control, left) and stirred WD (at concentration 1:10, right); (b) in static WD (at concentration  
628 1:10) and (c) treated with WD (at concentration 1:10) with a cellulose poultice application; (d-g) cross-sectioned  
629 profiles of slabs incubated (d) in water, (e) stirred and (f) static WD (at concentration 1:10) and (g) treated with  
630 WD (at concentration 1:10) with a cellulose poultice application. Scale bars: 1 mm (a-c), 0.5 mm (d-g).

631









