

AperTO - Archivio Istituzionale Open Access dell'Università di Torino

Clastogenic effects of bisphenol A on human cultured lymphocytes

This is the author's manuscript

Original Citation:

Availability:

This version is available <http://hdl.handle.net/2318/1624718> since 2018-05-02T15:19:22Z

Published version:

DOI:10.1177/0960327117693069

Terms of use:

Open Access

Anyone can freely access the full text of works made available as "Open Access". Works made available under a Creative Commons license can be used according to the terms and conditions of said license. Use of all other works requires consent of the right holder (author or publisher) if not exempted from copyright protection by the applicable law.

(Article begins on next page)

Clastogenic Effects of Bisphenol-A on Human Cultured Lymphocytes

Journal:	<i>Human and Experimental Toxicology</i>
Manuscript ID	HET-16-0460.R1
Manuscript Type:	Original Article
Date Submitted by the Author:	n/a
Complete List of Authors:	Santovito, Alfredo; University of Turin, Department of Life Sciences and Systems Biology Cannarsa, Elio; University of Turin, Department of Life Sciences and Systems Biology Schleicherova, Dáša ; University of Turin, Department of Life Sciences and Systems Biology Cervella, Piero; University of Turin, Department of Life Sciences and Systems Biology
Keyword:	endocrine disruptors, micronucleus assay, chromosomal aberrations assay, human lymphocytes, BP-A
Abstract:	Bisphenol-A is an endocrine disrupting compound widely used in the production of polycarbonate plastics and epoxy resins. It is ubiquitously present in the environment, mostly in aquatic environments, with consequent risks to the health of aquatic organisms and humans. In the present study we analysed the cytogenetic effects of bisphenol-A on human lymphocytes using in vitro chromosomal aberrations and micronuclei assays. Lymphocyte cultures were exposed to five different concentrations of BP-A (0.20, 0.10, 0.05, 0.02 and 0.01 µg/mL) for 24 h (for chromosomal aberrations test) and 48 h (for micronuclei test). The concentration of 0.05 µg/mL represents the reference dose established by United States Environmental Protection Agency; 0.02 µg/mL represents the higher concentration of unconjugated BP-A found in human serum and 0.01 µg/mL represents the tolerable daily intake established by European Union. Data obtained with both assays showed significant genotoxic effects of the bisphenol-A at concentrations of 0.20, 0.10 and 0.05 µg/mL, whereas at the concentration of 0.02 µg/mL we observed only a significant increase in the micronuclei frequency. Finally, at the concentration of 0.01 µg/mL no cytogenetic effects were observed, indicating this latter as a more tolerable concentration for human health with respect to 0.05 µg/mL, the reference dose established by US EPA.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

SCHOLARONE™
Manuscripts

For Peer Review

1
2
3
4 **Title: Clastogenic Effects of Bisphenol-A on Human Cultured Lymphocytes**

5
6 **Short Title: Effects of Bisphenol-A on Human Lymphocytes**

7
8
9
10
11 Authors: Alfredo SANTOVITO*, Elio CANNARSA, Dáša SCHLEICHEROVA, Piero
12
13 CERVELLA

14
15
16
17
18 University of Turin, Department of Life Sciences and Systems Biology, Via Accademia Albertina
19
20 n. 13, 10123 Torino (Italy)

21
22
23
24 *Corresponding Author:

25
26 Alfredo SANTOVITO

27
28 Department of Life Sciences and Systems Biology

29
30 Via Accademia Albertina n. 13

31
32 10123 – Torino (Italy)

33
34 Tel.: +39-0116704554

35
36 Fax: +39-0116704508

Summary

Bisphenol-A is an endocrine disrupting compound widely used in the production of polycarbonate plastics and epoxy resins. It is ubiquitously present in the environment, mostly in aquatic environments, with consequent risks to the health of aquatic organisms and humans. In the present study we analysed the cytogenetic effects of bisphenol-A on human lymphocytes using *in vitro* chromosomal aberrations and micronuclei assays. Lymphocyte cultures were exposed to five different concentrations of BP-A (0.20, 0.10, 0.05, 0.02 and 0.01 µg/mL) for 24 h (for chromosomal aberrations test) and 48 h (for micronuclei test). The concentration of 0.05 µg/mL represents the reference dose established by **United States Environmental Protection Agency**; 0.02 µg/mL represents the higher concentration of unconjugated BP-A found in human serum and 0.01 µg/mL represents the tolerable daily intake established by **European Union**. Data obtained with both assays showed significant genotoxic effects of the bisphenol-A at concentrations of 0.20, 0.10 and 0.05 µg/mL, whereas at the concentration of 0.02 µg/mL we observed only a significant increase in the micronuclei frequency. Finally, at the concentration of 0.01 µg/mL no cytogenetic effects were observed, indicating this latter as a more tolerable concentration for human health with respect to 0.05 µg/mL, the reference dose established by US EPA.

Keywords:

BP-A, endocrine disruptors, micronucleus assay, chromosomal aberrations assay, human lymphocytes.

1. Introduction

In the last decades, global production of plastic has continued to rise. Recovery and recycling, however, remain insufficient and millions of tons of plastic compounds continue to accumulate in the environment. Most of these compounds have adverse effects on the biome and pose huge risks for the human health. Some of them are reported to be endocrine disrupting compounds (EDCs), mimicking or blocking natural hormone action and altering the normal functioning of the endocrine system of invertebrate and mammals species.¹

Among EDCs, bisphenol A (BP-A: 4,4'-(propane-2,2-diyl)diphenol) has received much attention due to its high production volume and widespread human exposure.² BP-A is an important synthetic compound used in the production of polycarbonate plastics and as a precursor of epoxy resins. A wide range of consumer products contain BP-A, including water and baby bottles, food packaging materials, internal coatings of cans and drums, toys, water pipes, sports equipment, medical and dental devices, and household electronics.³

Due to the daily use, BP-A was found in all environmental matrices such as air, water, soil, sediment, and sewage sludge. It is released into the environment either through sewage treatment effluent, landfill leachate, untreated urban and industrial wastewater or natural degradation of polycarbonate plastics.⁴ For these reasons, BP-A was found at relatively high concentrations (up to hundreds of $\mu\text{g/L}$) in several rivers and lakes worldwide.⁵

Being BP-A used for food packaging, bottles, and coatings for tins, human exposure to this compound is extremely variable depending on life-styles, principally dietary consumption habits.

In general, humans are chronically exposed to low doses of the compound, mainly through food and drinking water.⁶ However, the release of BP-A from polycarbonates is accelerated by UV light, aging, heating, or following pH changes due to the contact with acidic or basic compounds.⁷ It has been shown that upon long-term heating BP-A can leach from plastic baby bottles into an aqueous solution.⁸ For these reasons, firstly in Canada, Denmark and next in all European countries, BP-A

1
2 has been banned in infant formula bottles.⁹
3

4 Nevertheless, many plastic food and drinking containers still contain BP-A and its presence has been
5 reported in human serum and in 95% of the urine samples obtained from a reference population in
6 the United States,¹⁰ as well as in maternal and fetal plasma,¹¹ placental tissue,¹² and in the milk of
7 lactating mothers.¹³ Several studies, using different analytical techniques, have also measured the
8 unconjugated BPA concentrations in human serum, reporting values ranging from 0.2 to
9 20 ng/mL.² Finally, the estimated exposure of the general population to BP-A is supposed to reach
10 a body burden of up to 9 µg/kg/day.¹⁴
11
12
13
14
15
16
17
18
19

20 In humans, BP-A is rapidly absorbed from the gastrointestinal tract and conjugated with glucuronic
21 acid in the liver. The formed glucuronide is then cleared from blood by elimination with urine
22 within 24 h after oral administration. This rapid excretion results in a low body burden of the BP-A
23 in humans following oral absorption of low doses.¹⁵ However, despite its rapid rate of excretion, as
24 consequence of its ubiquitous and widespread distribution, BP-A has arisen worldwide concerns
25 about its possible associations with human diseases such as obesity⁶ and cancer.¹⁶ At reproductive
26 system level, exposure to ECDs seems to be associated with decreased fertility and increased risk of
27 testicular or prostate cancer in men,¹⁷ whereas in women exposure to ECDs seems to increase the
28 risk of endometriosis, reproductive or other endocrine-related cancers.¹⁸
29
30
31
32
33
34
35
36
37
38
39

40 From the genetic point of view, the genotoxicity of BP-A has been evaluated in several *in vitro* and
41 *in vivo* studies, although with controversial results. BP-A was found to be able to induce increase of
42 chromosome aberrations (CAs) and DNA adducts in Syrian hamster embryo cells,¹⁹ micronuclei
43 (MNs) in human MCL-5 cells²⁰ and in fish erythrocytes,²¹ MNs and CAs formation in CHO
44 cells.²² A significant genotoxic effect by BP-A was also seen on bovine peripheral lymphocytes *in*
45 *vitro*, but only at the highest concentration of 1×10^{-4} M.²³ Moreover, Tiwari *et al.*²⁴ observed a
46 significant increase of MNs in polychromatic erythrocytes, structural CAs in bone marrow cells and
47 DNA damage in lymphocytes of rats exposed to BP-A. Interestingly, in HEK293 cells, BP-A was
48 found to alter the expression of some genes involved in important biological processes including
49
50
51
52
53
54
55
56
57
58
59
60

1
2 ion transport, cysteine and glycogen metabolic processes, apoptosis, DNA damage repair and many
3
4 others²⁵.

5
6
7 On the other hand, BP-A failed to induce gene mutations in Syrian hamster cells,¹⁹ polyploidy or
8
9 MNs in male mice,²⁶ and sister chromatid exchanges in mouse lymphoma cells.²⁷

10
11 Nevertheless, based on published data, the United States Environmental Protection Agency (US
12
13 EPA) has set a reference dose (RD) for BP-A oral exposure of 0.05 mg/kg body weight
14
15 (BW)/day.²⁸ More recently, the European Union (EU)¹⁴ established a temporary tolerable daily
16
17 intake (TDI) for BPA of 10 µg/kg BW/day, based on a consideration of much toxicological data,
18
19 including the effects of BPA on reproductive organs and functions, and hormonal effects in
20
21 animals.

22
23
24 No data are reported in literature about *in vitro* cytogenetic effects of BP-A on human lymphocytes.
25
26 For this reason, in the present study we assessed the possible *in vitro* clastogenic and aneugenic
27
28 effects of BP-A exposure on human lymphocyte cells, by CAs and MNs assays.

29
30 It has been shown that a high frequency of the chromosomal damage is a predictive value of health
31
32 risk.²⁹ In this scenario, CAs and MNs assays are important tools in the measurement of the
33
34 genotoxic potential of many chemicals, including BP-A. The CAs assay allows the detection of
35
36 cells carrying unstable aberrations (chromosome and chromatid breaks, deletions, fragments, rings,
37
38 dicentrics and chromatid exchanges) that will lead to cell death during proliferation.³⁰ Previous
39
40 studies have provided strong evidence in support of the hypothesis that high CAs frequency in
41
42 peripheral blood lymphocytes is a powerful predictor of cancer risk. Indeed, it is known that cancer
43
44 incidence among healthy individuals of a population increases with increased levels of CAs in their
45
46 circulating lymphocytes.²⁹ On the other hand, MNs are used as a fast and reliable assay for
47
48 detecting both potential clastogenic (chromosome breakage) or aneugenic (chromosome lagging
49
50 due to dysfunction of mitotic apparatus) effects of different xenobiotics. In particular, MNs
51
52 represent acentric chromosomal fragments or whole chromosomes left behind during mitotic cell
53
54 division and appear in the cytoplasm of interphase cells as small additional nuclei. Bonassi *et al.*³¹
55
56
57
58
59
60

1
2 provided evidences that also increased MNs frequencies in peripheral blood lymphocytes are
3
4 positively associated with early events in carcinogenesis.
5

6
7 In the present study, we tested concentrations of 0.20, 0.10, 0.05, 0.02 and 0.01 µg/mL, where 0.05
8
9 µg/mL represents the RD established by US EPA; ²⁸ 0.20 µg/mL and 0.10 µg/mL are multiples of
10
11 this RD in order to establish a possible dose-dependent increase of the genomic damage; 0.020
12
13 µg/mL represents the higher concentration of unconjugated BP-A found in literature in human
14
15 serum, ² and 0.01 µg/mL is the TDI established by EU. ¹⁴
16
17

18 19 20 **2. Methods**

21 22 23 *2.1 Chemicals and Media*

24
25 The IUPAC name of Bisphenol-A (BP-A, CAS n. 80-05-7) is: 4,4'-(propane-2,2-diyl)diphenol. The
26
27 BP-A was dissolved in Dimethyl Sulfoxide (DMSO, CAS no. 67-68-5). Gibco RPMI 1640 cell
28
29 culture media supplemented with L-glutamine, foetal calf serum, phytohemagglutinin (PHA), and
30
31 antibiotics were purchased from Invitrogen-Life Technologies, Milan, Italy. Cytochalasin-B,
32
33 Mitomycin-C (MMC), BP-A and Giemsa stain solution were obtained from Sigma-Aldrich, Milan,
34
35 Italy. Methanol, Acetic acid, and conventional microscope slides were purchased from Carlo Erba
36
37 Reagenti, Milan, Italy. Potassium chloride (KCl) and Sørensen buffer were obtained from Merck
38
39 S.p.A., Milan, Italy. Vacutainer blood collection tubes were from Terumo Europe, Rome, Italy.
40
41
42
43
44

45 46 *2.2 Subjects*

47
48 Peripheral venous blood was collected from 5 healthy **female** subjects (mean age±S.D., 27.60±5.22,
49
50 range 22-30 years), non-smoking, non-alcoholic, not under drug therapy, and with no recent history
51
52 of exposure to mutagens. Informed consent was obtained from all blood donors. The study was
53
54 approved by the Univeristy of Turin ethics committee and was performed in accordance with the
55
56 ethical standards laid down in the 1964 Declaration of Helsinki.
57
58
59
60

2.3 Blood Sample Collection and Lymphocyte cultures

Heparinized blood samples were obtained by venipuncture and collected in heparinised tubes, for genotoxicity testing. All blood samples were coded, cooled (4°C), and processed within 2 h after collection. Heparinized venous blood (0.3 mL) was cultured in 25 cm² flasks in 6 mL of RPMI-1640 medium supplemented with 20% foetal calf serum (FCS), 2% of the mitogenic agent PHA, L-glutamine (2 mM), antibiotics (100 IU/mL penicillin, and 100 µg/mL streptomycin). The cultures were incubated for 72 h at 37°C, under 5% of CO₂ in the air in a humidified atmosphere. After 24 h of incubation, BP-A dissolved in DMSO (for stock solution preparation 200 µg of BP-A were dissolved in 1 mL of DMSO) was added to the cultures to a final concentrations of 0.20 µg/mL, 0.10 µg/mL, 0.05 µg/mL, 0.02 and 0.01 µg/mL. The concentration at 0.05 µg/mL represents the RD established by US EPA²⁸ (0.05 mg/kg BW); 0.20 µg/mL and 0.10 µg/mL are multiples of this RD; 0.020 µg/mL represents the higher concentration of unconjugated BP-A found in literature in human serum,¹⁵ and 0.01 µg/mL is the TDI established by EU¹⁴ (10 µg/kg BW).

Three control cultures were assessed: 1) positive control, by adding only MMC (final concentration 0.1 µg/mL culture); 2) solvent control, by adding only 0.1% of DMSO; 3) negative control, culture without both BP-A and DMSO. Only for MNs assay, after 44 h of incubation, cytochalasin-B was added to the cultures at a concentration of 6 µg/mL to block cytokinesis.

After 48 h (for CAs assay) and 72 h (for MNs assay) of incubation at 37°, the cells were collected by centrifugation and treated for 10 min with a pre-warmed mild hypotonic solution (75 mM KCl). After centrifugation and removal of the supernatant, the cells were fixed with a fresh mixture of methanol/acetic acid (3:1 v/v). The treatment with the fixative was repeated three times. Finally, the supernatant was discarded and the pellet, dissolved in a minimal volume of fixative, was seeded on the slides to detect CAs and MNs by conventional staining with 5% Giemsa (pH 6.8) prepared in Sørensen buffer.

2.4 Cytokinesis-Block Micronucleus Assay

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Microscope analysis was performed at 40X magnification on a light microscope (Dialux 20, Leica, Germany). MNs were scored in 1000 binucleated lymphocytes with well-preserved cytoplasm per subject (total 5000 binucleated cells per concentration), following the established criteria for MNs evaluation.³² A total of 1000 lymphocytes per donor per concentration were scored to evaluate the percentage of cells with 1-4 nuclei. The cytokinesis-block proliferation index (CBPI) was calculated, according to the following formula: $[1 \times N1] + [2 \times N2] + [3 \times (N3 + N4)]/N$, where N1–N4 represents the number of cells with 1-4 nuclei, respectively, and N is the total number of cells scored.

2.5 Chromosomal Aberrations Assay

Microscope analysis was performed at 1000X magnification on a light microscope (Dialux 20, Leica, Germany). Although the Organization for Economic Co-operation and Development (OECD) guidelines call for 300 metaphases to be scored for each dose,³³ in order to obtain results more significant and suitable for statistical analysis, it is the normal practice in our laboratory to score 200 metaphases *per subject per dose*. Thus, in the present work, for each subject and BP-A concentration, 200 well-spread first-division complete metaphases (for a total of 1000 metaphases for each dose) **containing 46±1 chromosomes** were analysed for the following categories of CAs: gaps, chromatid breaks (B'), chromosome breaks (B''), dicentrics (Dic), rings (R), tri- or tetra-radials (TR), acentric fragments (AF), rearrangements (Re) and numerical aberrations. Cells containing **one of more** types of CAs were scored as "aberrant cell" (Ab.C).

With regard to the opportunity to include gaps in the statistical analyses, the discussion is open.

Although some authors considered gaps as the appropriate indicator of genotoxic potential of chemicals,³⁴ the molecular mechanism of BP-A to induce achromatic lesion/gaps is yet to be revealed. Thus, in our statistical analyses we decided to exclude gaps.

2.5 Statistical analysis

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Comparison of mean values of the percentage of cells with MNs, CBPI and CAs between exposition levels and their controls was assessed by the non-parametric Wilcoxon Mann-Whitney *U* test. Statistical calculations were carried out using the SPSS software package program (version 23.0, Inc., Chicago, IL, USA). All *P* values were two tailed, and *P* values of 5% or less were considered statistically significant for all tests carried out.

3. Results

3.1 Effect of BP-A on CAs formation

Table 1 shows values of CAs found in the human peripheral lymphocytes cultured in the presence of different concentrations of BP-A. BP-A was found to induce seven types of structural CAs (gaps, chromatid and chromosome breaks, dicentric chromosomes, rings, tri- or tetradials, acentric fragments and rearrangements). No numerical aberrations were found.

Data obtained indicated that human lymphocytes treated *in vitro* with BP-A at concentrations of 0.20, 0.10 and 0.05 $\mu\text{g}/\text{mL}$ significantly increased the CAs frequency compared with the solvent control, whereas at the lower concentrations of 0.02 and 0.01 $\mu\text{g}/\text{mL}$ this effect was not observed. Among tested concentrations, significant differences were found between 0.02 and 0.01 $\mu\text{g}/\text{mL}$ and the other three tested concentrations (with *p*-value of $P = 0.001$, $P = 0.028$ and $P = 0.023$, respectively, for 0.02 $\mu\text{g}/\text{mL}$ and $P < 0.001$, $P = 0.006$ and $P = 0.006$, respectively, for 0.01 $\mu\text{g}/\text{mL}$), as well as between the higher tested concentration (0.20 $\mu\text{g}/\text{mL}$) and 0.05 $\mu\text{g}/\text{mL}$ ($P = 0.010$). As already observed in our previous work, the DMSO solvent-control cultures did not show any difference compared to the control cultures ($P = 0.058$) confirming that, at this low dose (0.1%), DMSO has no relevant biological effects. Finally, as expected, control cultures treated with the known mutagen MMC showed a significant increase of CAs with respect to all BP-A tested concentrations ($P < 0.001$), as well as to the negative and solvent control cultures ($P < 0.001$).

3.2 Effect of BP-A on MNs formation

1
2 To verify both the aneugenic and clastogenic effects of BP-A, the MN test was assessed in parallel
3
4 with CAs test (Table 2). Our results indicated that BP-A significantly, and in a dose-dependent
5
6 manner, increased the MNs formation compared with the negative and solvent controls, at all
7
8 concentrations tested, including 0.02 µg/mL, that represents the higher concentration of
9
10 unconjugated BP-A found in literature in human serum.² Similarly to what we already observed
11
12 with the CAs assay, at the concentration of 0.01 µg/mL, that represent the TDI established by EU,¹⁴
13
14 no significant cytogenetic effects were observed.
15
16

17
18 The highest tested concentration (0.20 µg/mL) caused a fourfold increase of MNs as compared with
19
20 the solvent culture. Moreover, significant differences in MNs frequency were also found between
21
22 0.20 µg/mL and all other concentrations, as well as between 0.10 µg/mL and the last two lower
23
24 concentrations (0.02 and 0.01 µg/mL). The solvent-control cultures did not show any difference
25
26 with the negative controls ($P = 0.059$), further confirming that at this low concentration DMSO has
27
28 no cytogenetic effects. Similarly to what we observed with CAs assay, cultures treated with the
29
30 mutagen MMC showed a substantially increased MNs formation compared with the negative
31
32 control, solvent controls and all tested concentrations of BP-A. Finally, after 48-h exposure, a
33
34 significant reduction of the CBPI value in cultures treated with BP-A was not observed, indicating
35
36 that BP-A does not seem to produce effects on the proliferation/mitotic index when its
37
38 concentration is equal or less than 0.20 µg/mL.
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

4. Discussion

The genotoxicity of BP-A has been evaluated in several *in vitro* and *in vivo* studies, although with controversial results.^{22, 23, 24, 26} Here, we discuss data obtained from our *in vitro* study conducted in order to establish the clastogenic and aneugenic potential of BP-A on human cultured lymphocytes. The results obtained in the present *in vitro* study with CAs and MNs assays evidenced a possible clastogenic effect of the BP-A on human lymphocytes, also at the concentration of 0.05 µg/mL that represents the reference dose established for humans by the US EPA.²⁸ This result is consistent with that obtained by Tayama *et al.*,³⁵ who demonstrated that BP-A, in cultured CHO-K1 cells, is able to induce significant structural CAs, like gaps, breaks and exchanges. Similarly, Xin *et al.*²² observed a significant increase of MNs frequency in CHO cells after 24 h exposure to doses of 80 and 120 µM of BP-A (corresponding to 0.018 and 0.027 µg/mL, respectively). Finally, Parry *et al.*²⁰ and Johnson and Parry³⁶ showed that BP-A cause MNs induction, also in these cases in a dose-dependent manner, on MCL5 and AHH-1 human lymphoblastoid cell lines and Chinese hamster V79 cells.

The increase of cytogenetic damage observed with both CAs and MNs assays at concentration of 0.05 µg/mL pushes towards the adoption of lower reference limits. Indeed, increased CAs and MNs frequencies in peripheral blood lymphocytes have been positively associated with increased cancer risk and early events in carcinogenesis, respectively.^{29, 31} *Vice versa*, at the concentration of 0.01 µg/mL, the TDI established by the EU (2002)¹⁴, no cytogenetic effects were observed, indicating that this latter represents a more safe concentration for human health with respect to 0.05 µg/mL. The clastogenic properties of the BP-A were attributed by Adamkis *et al.*³⁷ to its capacity to interact with the cell microtubules causing their depolymerization and consequent manifestation of chromosomal abnormalities. Moreover, it is evident from the literature that BP-A is able to generate ROS by decreasing the activities of antioxidant enzymes and increasing lipid peroxidation. A possible consequence of this action could be the increase of the cellular oxidative stress that is considered one of the possible mechanisms underlying DNA damage.³⁸ BP-A was also found to be

1
2 able to up-regulate the expression of ERCC5 encoding a DNA endonuclease involved in
3
4 nucleotide-excision repair. This is an important issue because a reduced repair capacity has been
5
6 reported to be associated with increased risk of genomic damage and cancer development.³⁹
7
8 However, we cannot ascribe the observed cytogenetic damage exclusively to the direct effect of
9
10 BP-A on lymphocytes. Indeed, *in vitro* studies showed that BP-A is converted to DNA-binding
11
12 metabolites. In particular, BP-A is oxidized to bisphenol-*o*-quinone whose chemical reaction with
13
14 DNA produces adducts *in vitro* and *in vivo*.⁴⁰ The irreversible binding of BP-A derived compounds
15
16 to DNA could cause bases loss from the DNA strand and could be responsible for some of the BP-
17
18 A observed toxic effects. It is known that many chemicals, either directly or after conversion to
19
20 reactive metabolites, covalently modify nucleosides of DNA leading to mutational changes and,
21
22 thus, playing a role in cellular toxicity or tumorigenesis induction.⁴⁰
23
24 Moreover, it should be also emphasized that the link between observed toxicological effects and
25
26 BP-A exposure, especially at low doses, is complicated by the fact that BP-A is prevalent in many
27
28 consumer products, including common laboratory plastics, and has the potential to leach from these
29
30 products. Consequently, these background concentrations of BP-A have the potential to affect the
31
32 results of *in vitro* studies that detect low concentration effects.⁴¹ Indeed, some *in vitro* studies
33
34 showed that BP-A induces significant effects also to nominal concentrations. For example, a BP-A
35
36 concentration of 0.23 pg/mL was found to promote proliferation of human seminoma cells (JKT-
37
38 1).⁴²
39
40 Interestingly, BP-A increased MNs frequency also at the concentration of 0.02 µg/mL, where CAs
41
42 failed. These apparently discordant results could be explained by the fact that the MNs assays
43
44 reveal alterations in either chromosome number or chromosome structure, whereas the CAs
45
46 analysis can detect alterations only in the chromosome structure. One of the mechanisms of action
47
48 of aneugenic chemicals is the induction of alterations to centromeric DNA, with consequent failure
49
50 of the damaged chromosomes to attach to the mitotic spindle. As also postulated by other authors
51
52 for other xenobiotics⁴³, it is possible that BP-A may also damage the chromosomes centromeres,
53
54
55
56
57
58
59
60

1
2 determining MNs formation. Moreover, disruption of the mitotic machinery, *i.e.* microtubules and
3
4 microtubule organizing centers, has been shown to be the mechanism of action for BP-A induced
5
6 spindle-damaging effects.⁴⁴ In this scenario it can be hypothesized that BP-A might induce
7
8 aneuploidy through non-disjunction.
9

10 11 12 13 **5. Conclusion**

14
15
16
17 In conclusion, the results reported herein demonstrate cytogenetic effects of BP-A on cultured
18
19 human lymphocytes by increasing MNs and CAs frequencies. A large number of studies showing
20
21 the clastogenic and toxic properties of BP-A were performed using different cell models. For these
22
23 reasons, although simple experimental models, like mammalian and bacterial cells, cannot
24
25 accurately mimic the complex kinetics of xenobiotic compounds *in vivo*, our results suggest the
26
27 adoption of more stringent measures in order to reduce the presence of this compound in the
28
29 environment and to minimize the adverse health effects of BP-A exposure. Finally, being BP-A
30
31 converted *in vivo* and *in vitro* in DNA binding metabolites, it would be useful and interesting to test
32
33 *in vitro* also the genotoxicity of the BP-A metabolites or a combination of both BP-A and its
34
35 metabolites. Indeed, BP-A was found to have a short half-life (47 h in H295R cells)⁴⁵ and therefore
36
37 it is conceivable that in *in vitro* systems about 50% of BP-A could not be metabolized after 48 h
38
39 treatment (typical of MNs assay) and more than 50% after 24 h treatment (typical of CAs assay). In
40
41 this sense, a combined genotoxic action of both BP-A and its metabolites cannot be excluded.
42
43
44
45

46 47 **Acknowledgment**

48
49 We are grateful to all volunteers who participated in this study.
50

51 52 **Conflict of Interest**

53
54 The authors declare no conflicts of interest, with respect to the authorship and/or publication of this
55
56 article.
57

58 59 **Funding**

1
2 This research was supported by grant from the Italian Ministry of University and Scientific
3
4 Research (“ex 60%”).
5
6
7
8

9 References

- 10
11
12 1) Crain DA, Eriksen M, Iguchi T, Jobling S, Laufer H, LeBlanc GA and Guillette LJ Jr.
13
14 Anecological assessment of bisphenol-A: evidence from comparative biology. *Reprod*
15
16 *Toxicol* 2007; 24: 225-239.
17
18
- 19 2) Vandenberg LN, Hauser R, Marcus M, Olea N and Welshons WV. Human exposure to
20
21 bisphenol A (BPA). *Reprod Toxicol* 2007; 24: 139-177.
22
23
- 24 3) Amanti-Kandarakis E, Bourguignon JP, Giudice LC, Hauser R, Prins GS, Soto AM, Zoeller
25
26 RT and Gore AC. Endocrine-disrupting chemicals: an endocrine society scientific statement,
27
28 *Endocr Rev* 2009; 30: 293-342
29
- 30 4) Liao C, Liu F, Moon HB, Yamashita N, Yun S and Kannan K. Bisphenol analogues in
31
32 sediments from industrialized areas in the United States, Japan, and Korea: spatial and
33
34 temporal distributions. *Environ Sci Technol* 2012; 46: 11558-11565
35
36
- 37 5) Tao S, Zhang Y, Yuan C, Gao J, Wu F, and Wang Z. Oxidative stress and immunotoxic
38
39 effects of bisphenol A on the larvae of rare minnow *Gobiocypris rarus*. *Ecotoxicol*
40
41 *Environm Safety* 2016; 124: 377-385
42
- 43 6) Ariemma F, D'Esposito V, Liguoro D, Oriente F, Cabaro S, Liotti A, Cimmino I, Longo M,
44
45 Beguinot F, Formisano P, and Valentino R. Low-Dose Bisphenol-A Impairs Adipogenesis
46
47 and Generates Dysfunctional 3T3-L1 Adipocytes. *PLoS One* 2016; 11(3): e0150762
48
49
- 50 7) Kang JH, Kito K, and Kondo F. Factors influencing the migration of bisphenol A from cans.
51
52 *J Food Prot* 2004; 66: 1444-1447
53
54
- 55 8) Kubwabo C, Kosarac I, Stewart B, Gauthier BR, Lalonde K, and Lalonde PJ. Migration of
56
57 bisphenol A from plastic baby bottles, baby bottle liners and reusable polycarbonate
58
59 drinking bottles. *Food Addit Contam Part A: Chem Anal Control Expo Risk Assess* 2009;
60

- 1
2 26: 928-937
3
4
5 9) Rogers JA, Metz L, and Yong VW. Review: Endocrine disrupting chemicals and immune
6 responses: A focus on bisphenol-A and its potential mechanisms. *Mol Immunol* 2013; 53(4):
7 421-430.
8
9
10
11 10) Calafat AM, Kuklenyik Z, Reidy JA, Caudill SP., Ekong J, and Needham LL. Urinary
12 concentrations of bisphenol A and 4-nonylphenol in a human reference population. *Environ.*
13 *Health Perspect* 2005; 133: 391-395
14
15
16
17
18 11) Ikezuki Y, Tsutsumi O, Takai Y, Kamei Y, and Taketani Y. Determination of bisphenol A
19 concentrations in human biological fluids reveals significant early prenatal exposure. *Hum*
20 *Reprod* 2002; 17: 2839-2841
21
22
23
24
25 12) Schonfelder G, Wittfoht W, Hopp H, Tallness CE, Paul M, and Chahoud I. Parent bisphenol
26 A accumulation in the human maternal-fetal-placental unit. *Environ. Health Perspect* 2002;
27 110: A703-A707.
28
29
30
31 13) Sun Y, Irie M, Kishikawa N, Wada M, Kuroda N and Nakashima K. Determination of
32 bisphenol A in human breast milk by HPLC with column-switching and fluorescence
33 detection. *Biomed Chromatogr* 2004; 18: 501-507.
34
35
36
37
38 14) EU-Report (2002). Draft risk assessment of bisphenol A (carried out in accordance with the
39 Council Regulation (EEC) 793/93 on the evaluation and control of the risks of “existing”
40 substances), Rapporteur, United Kingdom (Contact: Environment Agency, Chemicals
41 Assessment Section, Oxfordshire, U.K.).
42
43
44
45
46
47 15) Völkel W, Colnot T, Csanády GA, Filser JG and Dekant W. Metabolism and Kinetics of
48 Bisphenol A in Humans at Low Doses Following Oral Administration. *Chem Res Toxicol*
49 2002; 15:1281-1287
50
51
52
53
54 16) Toft G, Hagmar L, Giwercman A and Bonde JP. Epidemiological evidence on reproductive
55 effects of persistent organochlorines in humans. *Reprod Toxicol* 2004; 19: 5-26.
56
57
58 17) Pflieger-Bruss S, Schuppe HC and Schill W. The male reproductive system and its
59
60

- 1
2 susceptibility to endocrine disrupting chemicals. 2004; *Andrologia* 36: 337-345.
3
4
5 18) Niclopoulou-Stamati P and Pitsos MA. The impact of endocrine disrupters on the female
6 reproductive system. *Hum Reprod Update* 2001; 7: 323-330.
7
8
9 19) Tsutsui T, Tamura Y, Yagi E, Hasegawa K, Takahashi M, Maizumi N, Yamaguchi F and
10 Barrett JC. Bisphenol-A induces cellular transformation, aneuploidy and DNA adduct
11 formation in cultured Syrian hamster embryo cells. *Int J Cancer* 1998; 75(2): 290-294
12
13
14 20) Parry EM, Parry JM, Corso C, Doherty A, Haddad F, Hermine TF, Johnson G, Kayani M,
15 Quick E, Warr T and Williamson J. Detection and characterization of mechanisms of action
16 of aneugenic chemicals. *Mutagenesis* 2002; 17: 509-521.
17
18
19 21) Bolognesi C, Perrone E, Roggieri P, Pampanin DM and Sciutto A. Assessment of
20 micronuclei induction in peripheral erythrocytes of fish exposed to xenobiotics under
21 controlled conditions. *Aquat Toxicol* 2006; 78(1): S93-S98
22
23
24 22) Xin L, Lin Y, Wang A, Zhu W, Liang Y, Su X, Hong C, Wan J., Wang Y and Tian H.
25 Cytogenetic evaluation for the genotoxicity of bisphenol-A in Chinese hamster ovary cells.
26
27
28 *Environ Toxicol Pharmacol*. 2015; 40(2): 524-529
29
30
31 23) Šutiaková I, Kovalkovičová N and Šutiak V. Micronucleus assay in bovine lymphocytes
32 after exposure to bisphenol A in vitro. *In Vitro Cell Dev Biol Anim* 2014; 50(6): 502-506.
33
34
35 24) Tiwari D, Kamble J, Chilgunde S, Patil P, Maru G, Kawle D, Bhartiya U, Joseph L and
36 Vanage G. Clastogenic and mutagenic effects of bisphenol A: An endocrine disruptor.
37
38
39 *Mutat Res* 2012; 743(1-2): 83-90.
40
41
42 25) Yin R, Gu L, Li M, Jiang C, Cao T and Zhang X. Gene Expression Profiling Analysis of
43 Bisphenol A-Induced Perturbation in Biological Processes in ER-Negative HEK293 Cells.
44
45
46 *Plos-One* 2014; 9(6): e98635.
47
48
49 26) Pacchierotti E, Ranaldi R, Eichenlaub-Titter U, Attia S and Adler ID. Evaluation of
50 aneugenic effects of Bisphenol-A in somatic and germ cells of mouse. *Mutat Res* 2008; 651:
51
52
53 64-70.
54
55
56
57
58
59
60

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
- 27) Tennant RW, Margolin BH, Shelby MD, Zeiger E, Haseman JK and Spalding J. Prediction of chemical carcinogenicity in rodents from *in vitro* genetic toxicity assays. *Science* 1987; 236: 933-941.
- 28) US EPA (United States Environmental Protection Agency) (1988). Bisphenol A Reference Dose for Chronic Oral Exposure (RfD) in Integrated Risk Information System (IRIS), A Toxicology Data File on the National Library of Medicine's (NLM) TOXNET System.
- 29) Bonassi S, Znaor A, Norppa H, Hagmar L. Chromosomal aberrations and risk of cancer in humans: an epidemiologic perspective. *Cytogenet Gen Res* 2004; 104: 376-382.
- 30) Garcia-Sagredo JM. Fifty years of cytogenetics: a parallel view of the evolution of cytogenetics and genotoxicology. *Biochim. Biophys. Acta* 2008; 1779(6-7): 363-375.
- 31) Bonassi S, El-Zein R and Bolognesi C. Micronuclei frequency in peripheral blood lymphocytes and cancer risk: evidence from human studies. *Mutagenesis* 2011; 26: 93-100.
- 32) Fenech M, Chang WP, Kirsch-Volders M, Holland N, Bonassi S and Zeiger E. HUman Micronucleus project (HUMN project): detailed description of the scoring criteria for the cytokinesis-block micronucleus assay using isolated human lymphocyte cultures. *Mutat Res* 2003; 534(1-2): 65-75.
- 33) OECD (Organization for Economic Co-operation and Development). Guidance Documents on Revisions to OECD Genetic Toxicology Test Guidelines.
[https://www.oecd.org/env/ehs/testing/Draft Guidance Document on OECD Genetic Toxicology Test Guidelines.pdf](https://www.oecd.org/env/ehs/testing/Draft%20Guidance%20Document%20on%20OECD%20Genetic%20Toxicology%20Test%20Guidelines.pdf). 2015; 31. Accessed in date 08-22-2016.
- 34) Savage JR. On the nature of visible chromosomal gaps and breaks. *Cytogen Genome Res* 2004; 104(1-4): 46-55.
- 35) Tayama S, Nakagawa Y and Tayama K. Genotoxic effects of environmental estrogen-like compounds in CHO-K1 cells. *Mutat Res* 2008; 649(1-2): 114-125
- 36) Johnson GE and Parry EM. Mechanistic investigations of low dose exposures to the genotoxic compounds bisphenol A and rotenone. *Mutat Res* 2008; 651: 56-63

- 1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
- 37) Adamkis IDS, Panteris E, Cherianidou A and Eleftheriou PE. Effects of bisphenol A on the microtubule arrays in root meristematic cells of *Pisum sativum* L. *Mutat Res* 2013; 750: 111-120.
- 38) Bindhumol V, Chitra KC and Mathur PP. Bisphenol A induces reactive oxygen species generation in the liver of male rats. *Toxicology* 2003; 188: 117-124
- 39) Berwick M and Vineis P. Markers of DNA repair and susceptibility to cancer in humans: an epidemiologic review. *J Natl Cancer Inst* 2000; 92(11): 874-897
- 40) Atkinson A and Roy D. In vitro conversion of environmental estrogenic chemical bisphenol A to DNA binding metabolite(s). *Biochem Biophys Res Commun* 1995a; 210(2): 424-433
- 41) McDonald GR, Hudson AL, Dunn SM, You H, Baker GB, Whittal RM, Martin JW, Jha A, Edmondson DE and Holt A. Bioactive contaminants leach from disposable laboratory plasticware. *Science* 2008; 322:917
- 42) Bouskine A, Nebout M, Brucker-Davis F, Benahmed M and Fenichel P. Low doses of bisphenol A promote human seminoma cell proliferation by activating PKA and PKG via a membrane G-protein-coupled estrogen receptor. *Environ Health Perspect* 2009; 117:1053-1058.
- 43) Yüzbaşıoğlu D, Celik M, Yilmaz S, Unal F and Aksoy H. Clastogenicity of the fungicide afugan in cultured human lymphocytes. *Mutat Res* 2006; 604(1-2): 53-59.
- 44) George O, Bryant BK, Chinnasamy R, Corona C, Arterburn JB and Shuster CB. Bisphenol A directly targets tubulin to disrupt spindle organization in embryonic and somatic cells. *ACS Chem Biol* 2008; 3(3): 167-179
- 45) Zhang X, Chang H, Wiseman S, He Y, Higley E, Jones P, Wong CKC, Al-Khedhairi A, Giesy JP and Hecker M. Bisphenol A Disrupts Steroidogenesis in Human H295R Cells. *Toxicol Sciences* 2011; 121(2): 320-327.

Table 1 – Induction of chromosomal aberrations by Bisphenol-A in human lymphocytes *in vitro*.

Test substance	Treatment Period (h)	Dose (µg/ml)	Total Scored Metaphases	CAs								Total CAs	Total CAs + Gaps	Total Ab.C	Total Ab.C + Gaps	CAs/Cell ± S.E. (%)	Ab.C/Cell ± S.E. (%)	
				Gaps	B'	B''	DC	R	TR	AF	Re							
NC	----	----	1000	2	2	0	1	0	0	0	1	0	4	6	4	6	0.40±0.10 ^s	0.40±0.10 ^s
0.1% DMSO	24	----	1000	8	7	1	1	0	2	1	1	1	13	21	13	21	1.30±0.30	1.30±0.30
MMC	24	0.10	1000	38	35	24	11	6	14	26	10	10	126	164	123	161	12.60±1.30 ^{&}	12.30±1.00 ^{&}
BP-A	24	0.20	1000	17	29	7	10	1	0	18	5	5	70	87	70	87	7.00±0.70*	7.00±0.70*
BP-A	24	0.10	1000	16	22	3	5	2	2	8	7	7	49	65	49	65	4.90±0.60**	4.90±0.60**
BP-A	24	0.05	1000	8	17	3	5	0	0	11	6	6	42	50	41	50	4.20±0.90***	4.10±0.90***
BP-A	24	0.02	1000	8	6	2	2	2	0	7	6	6	25	33	25	33	2.50±0.40	2.50±0.40
BP-A	24	0.01	1000	6	5	2	1	1	0	5	5	5	19	25	19	25	1.90±0.30	1.90±0.30

CAs = chromosomal aberrations; Ab.C = aberrant cells (cells with 1 ore more aberrations); NC = Negative Control; MMC = Mitomycin-C; BP-A = bisphenol-A; B': chromatid break; B'': chromosome break; DC: dicentric; R: ring; TR = tri-tetradials; AF = acentric fragments; Re = rearrangements; S.E. = standard error.

$P < 0.001^*$, $P = 0.001^{**}$ and $P = 0.002^{***}$ with respect to DMSO control; $P < 0.001^{\&}$ with respect to all BP-A concentration and DMSO control. $P < 0.001^s$ with respect to all BP-A concentrations

Table 2 – Induction of micronuclei by Bisphenol-A in human lymphocytes *in vitro*.

Test substance	Treatment Period (h)	Dose (µg/ml)	BNCs scored	Distribution of BNCs according to the number of MN				MN _s	Ab.C	MN/cell ± S.E. (%)	Ab.C/cell ± S.E. (%)	CBPI ± S.E
				1	2	3	4					
NC	----	----	5000	9	1	0	0	11	11	2.20±0.20 ^s	2.20±0.20 ^s	1.712±0.003
0.1% DMSO	48	----	5000	15	1	1	0	20	16	4.00±0.55	3.20±0.37	1.565±0.092
MMC	48	0.10	5000	102	4	3	2	127	121	25.40±1.21*	24.20±1.20*	1.354±0.019
BPA	48	0.20	5000	72	2	2	1	86	78	17.20±2.31*	15.60±1.47*	1.584±0.046
BPA	48	0.10	5000	66	1	0	0	68	65	13.60±1.17*	13.00±1.30***	1.546±0.017
BPA	48	0.05	5000	55	0	3	0	64	56	12.80±1.16**	11.20±0.49***	1.585±0.016
BPA	48	0.02	5000	46	0	0	0	46	46	9.20±1.63**	9.20±1.63**	1.633±0.005
BPA	48	0.01	5000	31	0	0	0	31	25	6.20±0.80	5.00±0.95	1.640±0.009

BNCs = Binucleated cells; MN_s = micronuclei; Ab.C = Aberrant cells (cells with 1 or more MN_s); NC = Negative Control;

MMC = Mitomycin-C; BP-A = Bisphenol-A; S.E. = Standard Error. CBPI = Cytokinesis-Block Proliferation Index.

$P = 0.042^*$, $P = 0.043^{**}$ and $P = 0.039^{***}$ with respect to DMSO control. $P < 0.001^s$ with respect to all BP-A concentrations and MMC