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(Article begins on next page)

1 **Inhibition by phenolic antioxidants of the degradation of aromatic amines**
2 **and sulfadiazine by the carbonate radical ($\text{CO}_3^{\cdot-}$)**

3

4 **Luca Carena,^a Davide Vione,^{a,*} Marco Minella,^a Silvio Canonica^{b,*} and Ursula**
5 **Schönenberger^b**

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7 ^a Università di Torino, Dipartimento di Chimica, Via Pietro Giuria 5, 10125 Torino, Italy.

8 ^b Eawag, Swiss Federal Institute of Aquatic Science and Technology, Überlandstrasse 133,
9 CH-8600 Dübendorf, Switzerland

10

11 * Corresponding Authors:

12 DV: Telephone: +39-011-670-5296. Fax +39-011-670-5242. E-mail: davide.vione@unito.it.

13 SC: Telephone: +41-58-765-5453. Fax +41-58-765-5028. E-mail: silvio.canonica@eawag.ch.

14

15

16 ***Abstract***

17

18 The carbonate radical $\text{CO}_3^{\cdot-}$ and the excited triplet states of chromophoric dissolved organic
19 matter play an important role in the photodegradation of some easily oxidized pollutants in
20 surface waters, such as the aromatic amines. Anilines and sulfadiazine are known to undergo
21 back-reduction processes when their degradation is mediated by the excited triplet states of
22 photosensitizers (triplet sensitization). Back-reduction, which inhibits photodegradation,
23 means that phenols or the antioxidant (mostly phenolic) moieties occurring in the natural
24 dissolved organic matter of surface waters reduce, back to the parent compounds, the radical

25 species derived from the mono-electronic oxidation of anilines and sulfadiazine. Here we
26 show that a similar process takes place as well in the case of substrate oxidation by $\text{CO}_3^{\bullet-}$.
27 The carbonate radical was here produced upon oxidation of $\text{HCO}_3^-/\text{CO}_3^{2-}$ by either HO^\bullet ,
28 generated by nitrate photolysis, or $\text{SO}_4^{\bullet-}$, obtained by photolysis of persulfate. Back-reduction
29 was observed in both cases in the presence of phenols, but at different extents as far as the
30 details of reaction kinetics are concerned, and the occurrence of additional reductants might
31 affect the efficacy by which phenols carry out the reduction process. In particular, when the
32 carbonate radicals were produced by NO_3^- photolysis in the presence of $\text{HCO}_3^-/\text{CO}_3^{2-}$, the
33 numerical values of $[\text{PhOH}]_{1/2}$ (the phenol concentration that halves the photodegradation rate
34 of the substrate) were $2.19 \pm 0.23 \mu\text{M}$ for aniline, $1.15 \pm 0.25 \mu\text{M}$ for 3-chloroaniline, $1.18 \pm$
35 $0.26 \mu\text{M}$ for 4-chloroaniline, and $1.18 \pm 0.22 \mu\text{M}$ for 3,4-dichloroaniline. In contrast, when
36 $\text{CO}_3^{\bullet-}$ was produced by photolysis of persulfate in the presence of $\text{HCO}_3^-/\text{CO}_3^{2-}$, the
37 corresponding values were $0.28 \pm 0.02 \mu\text{M}$ for aniline and $0.79 \pm 0.10 \mu\text{M}$ for sulfadiazine.
38 Back-reduction has the potential to significantly inhibit photodegradation by $\text{CO}_3^{\bullet-}$ and
39 excited triplet states in natural waters, and to comparatively increase the importance of HO^\bullet -
40 mediated degradation that is not affected by the same phenomenon.

41

42 **Keywords:** Back-reduction; aniline; phenol; excited triplet states; photochemistry.

43

44 **1. Introduction**

45
46 Photochemical reactions are important processes for the degradation of organic contaminants,
47 especially biorecalcitrant compounds, in sunlit surface waters. Photoinduced processes consist
48 of the direct photolysis, in which the contaminant absorbs sunlight and gets consequently
49 transformed, and of indirect phototransformation. In the latter, absorption of sunlight by
50 photosensitizers (such as the chromophoric dissolved organic matter – CDOM –, nitrate and
51 nitrite) yields reactive transient species, either directly or upon interaction with the water
52 matrix (e.g., with the inorganic carbon species HCO_3^- and CO_3^{2-}). These photochemically
53 produced reactive intermediates (PPRIs) can react with the contaminants. The main PPRIs are
54 the hydroxyl radical (HO^\bullet), the triplet states of CDOM ($^3\text{CDOM}^*$), the singlet oxygen ($^1\text{O}_2$)
55 and the carbonate radical ($\text{CO}_3^{\bullet-}$) (Boreen et al., 2003; Remucal, 2014; Vialaton and Richard,
56 2002; Yan and Song, 2014).

57 The dissolved organic matter (DOM) and its chromophoric fraction (CDOM) play major roles
58 in the photochemistry of natural surface waters. CDOM is a direct or indirect source of all the
59 PPRIs, but it is also able to inhibit the direct photolysis processes by competing with the
60 contaminants for sunlight irradiance. Furthermore, the DOM is a key scavenger of both HO^\bullet
61 and $\text{CO}_3^{\bullet-}$ (Vione et al., 2014; Yan et al., 2019). Therefore, the natural organic matter that
62 occurs in surface waters is able to both inhibit and enhance the phototransformation reactions.
63 In addition to PPRIs scavenging, DOM can inhibit phototransformation through its
64 antioxidant moieties (AOs, mostly phenolic). Indeed, it has been shown that the reactions
65 between $^3\text{CDOM}^*$ and some contaminants (especially aromatic amines) yield partially
66 oxidized intermediates, which can be back-reduced to the parent contaminants by DOM-AOs
67 (Canonica and Laubscher, 2008; Wenk and Canonica, 2012).

68 This back-reduction process does not only affect reactions induced by $^3\text{CDOM}^*$, but has
69 recently been shown to occur for the reaction of the sulfate radical ($\text{SO}_4^{\bullet-}$) with several
70 anilines and sulfonamide antibiotics (Canonica and Schönenberger, 2019) as well as for the
71 reaction of $\text{CO}_3^{\bullet-}$ with the cyanotoxin cylindrospermopsin (Hao et al., 2020). Since $\text{CO}_3^{\bullet-}$ is a
72 typical one-electron oxidant (Neta et al., 1988) that reacts efficiently with a variety of
73 aromatic amine compounds (Canonica et al., 2005; Huang and Mabury, 2000a, 2000b), it is
74 expected that such compounds are on the one hand significantly oxidized by $\text{CO}_3^{\bullet-}$ in the
75 aquatic environment, and experience on the other hand back-reduction by DOM-AOs.
76 Interestingly, occurrence of back-reduction has not been observed in the case of HO^{\bullet} , which
77 undergoes addition to aromatic rings more easily than one-electron oxidation or H-atom
78 abstraction (Wenk et al., 2011).

79 The present study has the goal of detecting and quantifying back-reduction by two phenols,
80 used as DOM-AOs surrogates, in the reactions of $\text{CO}_3^{\bullet-}$ with several anilines and the
81 sulfonamide antibiotic sulfadiazine. The oxidation of $\text{HCO}_3^-/\text{CO}_3^{2-}$ to $\text{CO}_3^{\bullet-}$ by strongly
82 oxidizing radicals was carried out through two alternative methods (Huang and Mabury,
83 2000a; Neta et al., 1988). The first method is the photolysis of nitrate to yield HO^{\bullet} (Mack and
84 Bolton, 1999), and the second is the photolysis of persulfate ($\text{S}_2\text{O}_8^{2-}$) to form $\text{SO}_4^{\bullet-}$ (Criquet
85 and Leitner, 2009). In addition, the effect of phenol on the transformation of two anilines
86 photosensitized by a model aromatic ketone as a CDOM surrogate was also carried out, with
87 the purpose of comparing the extent of back-reduction among different oxidation reactions.

88

89 2. Materials and Methods

90
91 The list of used chemicals is reported in **Text S1** of the Supplementary Material, hereinafter
92 SM. The main issue with the production of $\text{CO}_3^{\bullet-}$ is that this radical is generated by oxidation
93 of $\text{HCO}_3^-/\text{CO}_3^{2-}$. Therefore, the production of $\text{CO}_3^{\bullet-}$ requires a stronger oxidant than $\text{CO}_3^{\bullet-}$
94 itself which may, however, be able to compete with $\text{CO}_3^{\bullet-}$ for the degradation processes.
95 Great care is thus to be taken to define the operational conditions, so that the chosen oxidant
96 species is almost exclusively involved in the oxidation of inorganic carbon, while the
97 generated $\text{CO}_3^{\bullet-}$ reacts preferentially with the target substrates. In this work, two methods
98 were used and compared for the generation of $\text{CO}_3^{\bullet-}$: (i) oxidation by HO^\bullet , produced by
99 nitrate photolysis, and (ii) oxidation by $\text{SO}_4^{\bullet-}$, produced by photolysis of persulfate ($\text{S}_2\text{O}_8^{2-}$).

100

101 2.1. Formation of the carbonate radical ($\text{CO}_3^{\bullet-}$) by nitrate photolysis

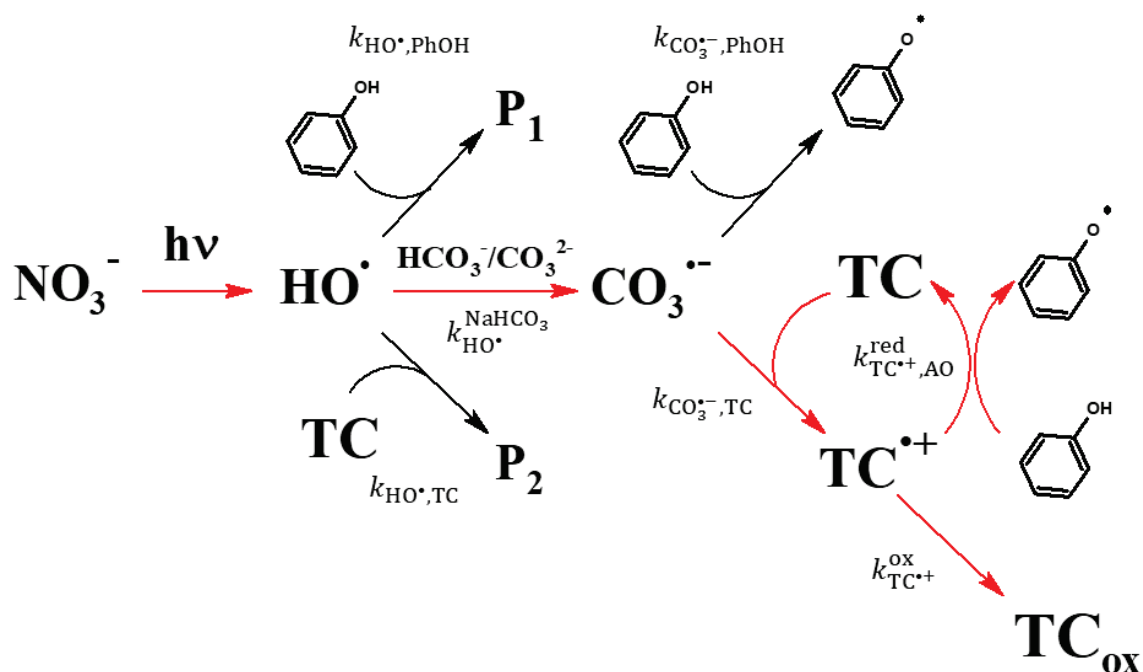
102 The $\text{CO}_3^{\bullet-}$ radicals were produced by irradiation of a mixture of $\text{NaNO}_3 + \text{NaHCO}_3$ (pH 8.3),
103 containing the aniline substrate (target compound TC) to be degraded as well as phenol
104 (PhOH). In particular, TC = aniline (Ani), 3-chloroaniline (3CA), 4-chloroaniline (4CA), and
105 3,4-dichloroaniline (3,4DCA). Under UVB light nitrate photolyzes and gives HO^\bullet , which then
106 oxidizes $\text{HCO}_3^-/\text{CO}_3^{2-}$ to form $\text{CO}_3^{\bullet-}$ and $\text{H}_2\text{O}/\text{HO}^-$ (second-order rate constants for the
107 corresponding reactions are defined in parenthesis) (Buxton et al., 1988).



111 Photogenerated $\cdot\text{NO}_2$ might interfere by reacting with some radical intermediates, but the use
 112 of H_2O_2 as alternative $\text{HO}\cdot$ source was ruled out because H_2O_2 also reacts with $\text{HO}\cdot$ producing
 113 the reductant $\text{O}_2^{\cdot-}$, which in turn has high potential to interfere with back-reduction processes
 114 (Canonica and Schönerberger, 2019).

115 In the system containing nitrate and bicarbonate under irradiation, both $\text{HO}\cdot$ and $\text{CO}_3^{\cdot-}$ could
 116 potentially react with TC and PhOH, causing their degradation. Because our aim was to assess
 117 the importance of back-reduction in the degradation of TC by $\text{CO}_3^{\cdot-}$, the reactions involving
 118 $\text{HO}\cdot + \text{TC}$, $\text{HO}\cdot + \text{PhOH}$ and $\text{CO}_3^{\cdot-} + \text{PhOH}$ had to be minimized (see **Scheme 1**). This goal
 119 can be achieved if photogenerated $\text{HO}\cdot$ is almost exclusively involved in the oxidation of
 120 $\text{HCO}_3^-/\text{CO}_3^{2-}$ to produce $\text{CO}_3^{\cdot-}$.

121



122

123 **Scheme 1.** Simplified scheme showing the main reactions taking place in our irradiated solutions, for
 124 the assessment of the back-reduction process during TC (anilines) degradation by $\text{CO}_3^{\cdot-}$. It is
 125 important to choose conditions so that the system follows the reaction pathway highlighted with red
 126 arrows.

127

128 In other words, the fraction of HO• that reacts with HCO₃⁻/CO₃²⁻ ($\chi_{\text{HO}^\bullet}^{\text{NaHCO}_3}$) should be > 95%

129 (**Eq. 4**). Note that $\chi_{\text{HO}^\bullet}^{\text{NaHCO}_3}$ takes into account the HO• reaction with both bicarbonate and

130 carbonate. At the same time, the majority of CO₃⁻ produced by HO• should react with TC

131 ($\chi_{\text{CO}_3^{\cdot-}}^{\text{TC}} > 95\%$, **Eq. 5**; see also **Text S2** in the SM).

132
$$\chi_{\text{HO}^\bullet}^{\text{NaHCO}_3} = \frac{k_{\text{HO}^\bullet}^{\text{NaHCO}_3} \times c_{\text{NaHCO}_3}}{k_{\text{HO}^\bullet, \text{TC}} \times [\text{TC}] + k_{\text{HO}^\bullet, \text{PhOH}} \times [\text{PhOH}] + k_{\text{HO}^\bullet}^{\text{NaHCO}_3} \times c_{\text{NaHCO}_3}} > 0.95 \quad (4)$$

133
$$\chi_{\text{CO}_3^{\cdot-}}^{\text{TC}} = \frac{k_{\text{CO}_3^{\cdot-}, \text{TC}} \times [\text{TC}]}{k_{\text{CO}_3^{\cdot-}, \text{TC}} \times [\text{TC}] + k_{\text{CO}_3^{\cdot-}, \text{PhOH}} \times [\text{PhOH}]} > 0.95 \quad (5)$$

134 In **Eqs. 4,5**, $k_{i,j}$ is the second-order rate constant of the reaction between the radical i (HO• or

135 CO₃⁻) and the species j (HCO₃⁻/CO₃²⁻, TC or PhOH). Moreover, it is

136 $k_{\text{HO}^\bullet}^{\text{NaHCO}_3} = (k_{\text{HO}^\bullet, \text{HCO}_3^-} \times \alpha_{\text{HCO}_3^-} + k_{\text{HO}^\bullet, \text{CO}_3^{2-}} \times \alpha_{\text{CO}_3^{2-}})$, where $\alpha_{\text{HCO}_3^-}$ and $\alpha_{\text{CO}_3^{2-}}$ are the molar

137 fractions of HCO₃⁻ and CO₃²⁻ at pH 8.3, computed taking into account the pK_a values of

138 carbonic acid (pK_{a1} = 6.3 and pK_{a2} = 10.3; e.g., Millero et al., 2002). Finally, [TC], [PhOH]

139 and c_{NaHCO_3} are the respective molar concentrations of TC (fixed at 5 μM; Vione et al., 2018;

140 Wenk and Canonica, 2012), PhOH and NaHCO₃ (*vide infra*) added to the solution.

141 Being the values of $k_{i,j}$ known (see **Table S1** in the SM) (Neta et al., 1988), **Eqs. 4,5** can be

142 solved for the variables [PhOH] and c_{NaHCO_3} . As a result, one gets the minimum value of

143 c_{NaHCO_3} (0.22-0.38 M, depending on substrate and conditions) and the maximum value of

144 [PhOH] (5-8 μM) that should be used to allow for: (i) the reactions of HO• with HCO₃⁻/CO₃²⁻

145 to prevail over the reactions of HO• with TC and PhOH, and (ii) the reaction CO₃⁻ + TC to

146 prevail over CO₃⁻ + PhOH (**Scheme 1**). The complete chemical composition of the irradiated

147 solutions is listed in **Table S1** (SM).

148 To carry out the above experiments, synthetic aqueous solutions (5 mL) were magnetically
 149 stirred and irradiated in cylindrical Pyrex glass cells, under a Philips narrow band TL 20W/01
 150 lamp. The lamp mainly emits in the UVB wavelength range, with emission maximum at 313
 151 nm. The spectral photon flux density ($p^0(\lambda)$, **Fig. S1**, SM) was assessed by means of chemical
 152 actinometry with 2-nitrobenzaldehyde (Carena et al., 2019; Galbavy et al., 2010; Willett and
 153 Hites, 2000).

154 The direct photolysis can be an additional degradation pathway for TC and, although this
 155 process does not affect the branching ratios of the reactions involving HO^\bullet and $CO_3^{\bullet-}$, it is still
 156 an experimental bias. Chloro-substituted anilines are known to undergo direct photolysis in
 157 the UVB range (**Fig. S2**, SM) (Carena et al., 2018). To minimize their direct photolysis rate,
 158 we used a higher $NaNO_3$ concentration (25 mM), exploiting the inner-filter effect of $NaNO_3$,
 159 and a lower lamp irradiance ($2.9 \pm 0.3 \text{ W m}^{-2}$). In contrast, aniline did not show direct
 160 photolysis, which allowed for the use of lower nitrate (10 mM) and slightly higher lamp
 161 irradiance ($4.2 \pm 0.2 \text{ W m}^{-2}$).

162

163 2.1.1. Modeling TC degradation rate

164 The degradation rate (R_{TC}) of the organic substrate TC (Ani, 3CA, 4CA or 3,4DCA) upon
 165 UVB irradiation of a solution containing $NaNO_3 + NaHCO_3 + TC + PhOH$ is

166 $R_{TC} = R_{CO_3^{\bullet-},TC} + R_{HO^\bullet,TC} + R_{d.p.}$, where $R_{CO_3^{\bullet-},TC}$, $R_{HO^\bullet,TC}$ and $R_{d.p.}$ are the rates of TC

167 degradation by $CO_3^{\bullet-}$, HO^\bullet and the direct photolysis, respectively. Based on the reactions

168 reported in **Scheme 1**, $R_{CO_3^{\bullet-},TC}$ and $R_{HO^\bullet,TC}$ can be expressed as follows:

$$169 \quad R_{CO_3^{\bullet-},TC} = k_{CO_3^{\bullet-},TC} \times [CO_3^{\bullet-}] \times [TC] - k_{TC^{\bullet+},AO}^{red} \times [TC^{\bullet+}] \times [PhOH] \quad (6)$$

$$170 \quad R_{HO^\bullet,TC} = k_{HO^\bullet,TC} \times [HO^\bullet] \times [TC] \quad (7)$$

171 Under the irradiation conditions used in this work, the steady-state approximation can be
 172 applied to HO^\bullet , $CO_3^{\bullet-}$ and $TC^{\bullet+}$. Their steady-state concentrations read as follows:

$$173 \quad [HO^\bullet] = \frac{R_{f,HO^\bullet}}{k_{HO^\bullet,TC} \times [TC] + k_{HO^\bullet,PhOH} \times [PhOH] + k_{HO^\bullet}^{NaHCO_3} \times c_{NaHCO_3}} \quad (8)$$

$$174 \quad [CO_3^{\bullet-}] = \frac{k_{HO^\bullet}^{NaHCO_3} \times c_{NaHCO_3} \times [HO^\bullet]}{k_{CO_3^{\bullet-},TC} \times [TC] + k_{CO_3^{\bullet-},PhOH} \times [PhOH]} \quad (9)$$

$$175 \quad [TC^{\bullet+}] = \frac{k_{CO_3^{\bullet-},TC} \times [CO_3^{\bullet-}] \times [TC]}{k_{TC^{\bullet+},AO}^{red} \times [PhOH] + k_{TC^{\bullet+}}^{ox}} \quad (10)$$

176 By introducing these steady-state concentration values in **Eqs. 6,7** and by considering the
 177 definitions of $\chi_{HO^\bullet}^{NaHCO_3}$, $\chi_{CO_3^{\bullet-}}^{TC}$ (**Eqs. 4,5**) and $\chi_{HO^\bullet}^{TC}$ (see **Text S2**, SM), one gets:

$$178 \quad R_{CO_3^{\bullet-},TC} = R_{f,HO^\bullet} \times \chi_{HO^\bullet}^{NaHCO_3} \times \chi_{CO_3^{\bullet-}}^{TC} \times \left(\frac{k_{TC^{\bullet+}}^{ox}}{k_{TC^{\bullet+},AO}^{red} \times [PhOH] + k_{TC^{\bullet+}}^{ox}} \right) \quad (11)$$

$$179 \quad R_{HO^\bullet,TC} = R_{f,HO^\bullet} \times \chi_{HO^\bullet}^{TC} \quad (12)$$

180 where R_{f,HO^\bullet} is the photoproduction rate of HO^\bullet upon nitrate photolysis (note that HO^\bullet is
 181 mostly involved in the generation of $CO_3^{\bullet-}$). It is $R_{f,HO^\bullet} \times \chi_{HO^\bullet}^{NaHCO_3} \times \chi_{CO_3^{\bullet-}}^{TC} = R_{CO_3^{\bullet-},TC}^0$, which is
 182 the value of $R_{CO_3^{\bullet-},TC}$ in the absence of phenol ($[PhOH] = 0$). When $[PhOH] = [PhOH]_{1/2}$, one
 183 has $R_{CO_3^{\bullet-},TC} = 0.5 \times R_{CO_3^{\bullet-},TC}^0$ and $k_{TC^{\bullet+}}^{ox} \times (k_{TC^{\bullet+},AO}^{red} \times [PhOH]_{1/2} + k_{TC^{\bullet+}}^{ox})^{-1} = 0.5$. As a
 184 consequence, it is $[PhOH]_{1/2} = k_{TC^{\bullet+}}^{ox} \times (k_{TC^{\bullet+},AO}^{red})^{-1}$ and **Eq. 11** is rearranged as follows:

$$185 \quad R_{CO_3^{\bullet-},TC} = R_{f,HO^\bullet} \times \chi_{HO^\bullet}^{NaHCO_3} \times \chi_{CO_3^{\bullet-}}^{TC} \times \left(\frac{1}{1 + \frac{[PhOH]}{[PhOH]_{1/2}}} \right) \quad (13)$$

186 Finally, by substituting **Eqs. 12,13** into the expression for R_{TC} , one gets **Eq. 14**:

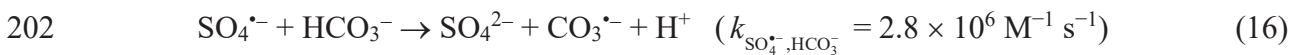
$$R_{TC} = R_{f_{HO^\bullet}} \times \left\{ \chi_{HO^\bullet}^{NaHCO_3} \times \chi_{CO_3^{\bullet-}}^{TC} \times \frac{1}{1 + \frac{[PhOH]}{[PhOH]_{1/2}}} + \chi_{HO^\bullet}^{TC} \right\} + R_{d.p.} \quad (14)$$

187 where $\chi_{HO^\bullet}^{NaHCO_3} = 0.965 - 0.985$ and $\chi_{HO^\bullet}^{TC} = 0.016 - 0.018$ are the fractions of HO^\bullet that react
 188 with, respectively, bicarbonate/carbonate and the anilines (TC) under our experimental
 189 conditions ($c_{NaHCO_3} = 0.32 - 0.38$ M, pH 8.3, $[TC]_0 = 5$ μ M), while $\chi_{CO_3^{\bullet-}}^{TC} = 0.97 - 0.99$ is the
 190 fraction of $CO_3^{\bullet-}$ reacting with TC. Note that the formation rate of $CO_3^{\bullet-}$ is
 191 $R_{f_{CO_3^{\bullet-}}} = R_{f_{HO^\bullet}} \times \chi_{HO^\bullet}^{NaHCO_3} \approx R_{f_{HO^\bullet}}$. The quantity $[PhOH]_{1/2}$ is operationally defined as the phenol
 192 concentration that halves the rate of TC degradation by $CO_3^{\bullet-}$, compared to the rate observed
 193 in the absence of phenol (when $[PhOH] = 0$, one has $R_{TC} \cong R_{f_{HO^\bullet}} + R_{d.p.}$; when $[PhOH] =$
 194 $[PhOH]_{1/2}$, one has $R_{TC} \cong 0.5 R_{f_{HO^\bullet}} + R_{d.p.}$). The lower is $[PhOH]_{1/2}$, the stronger is the
 195 inhibition of TC degradation by PhOH through the back-reduction process.
 196

197

198 2.2. Formation of $CO_3^{\bullet-}$ by persulfate photolysis

199 The carbonate radical was also generated by photolysis of the persulfate anion in the presence
 200 of an excess of bicarbonate anion, according to **Eqs. 15,16** (Huie and Clifton, 1990).



203 Side-reactions of the sulfate radical anion ($SO_4^{\bullet-}$) are discussed in the following. The
 204 carbonate anion (CO_3^{2-}) also reacts with $SO_4^{\bullet-}$ ($k_{SO_4^{\bullet-}, CO_3^{2-}} = 4.1 \times 10^6 \text{ M}^{-1} \text{ s}^{-1}$; Padmaja et al.,
 205 1993), but its contribution to $CO_3^{\bullet-}$ formation is only small (i.e., <1.5%) at pH 8.0, which was
 206 used in this study. Furthermore, no drawback of this reaction is expected, since no further

207 reactive species than $\text{CO}_3^{\bullet-}$ are produced. The reactions of $\text{SO}_4^{\bullet-}$ with organic compounds
 208 used in this study are expected to be fast (second-order rate constants up to $\sim 5 \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$;
 209 Neta et al., 1988). To reduce the transformation of these compounds due to direct reaction
 210 with $\text{SO}_4^{\bullet-}$, the concentrations of target compounds and the competitor were chosen to be a
 211 factor of 10^5 smaller compared to HCO_3^- . $\text{SO}_4^{\bullet-}$ can also react with the substrate used for its
 212 production (i.e., $\text{S}_2\text{O}_8^{2-}$), but this reaction channel is estimated to be ~ 440 times smaller
 213 (Herrmann et al., 1995) compared to $\text{CO}_3^{\bullet-}$ production. Owing to the low concentration of
 214 $\text{S}_2\text{O}_8^{2-}$ (1 mM), possible products of this reaction are not expected to be relevant for the
 215 transformation of the used organic compounds. At the experimental pH, water is a further
 216 reactant to be considered for $\text{SO}_4^{\bullet-}$ and leads to the formation of hydroxyl radical (HO^\bullet) (**Eq.**
 217 **17**) (Herrmann et al., 1995).



219 The contribution of HO^\bullet production from the reaction of $\text{SO}_4^{\bullet-}$ with HO^- at pH 8 ($[\text{HO}^-] = 1$
 220 μM) is about an order of magnitude smaller compared to water ($[\text{H}_2\text{O}] \sim 55 \text{ M}$) (Canonica and
 221 Schönenberger, 2019). At the bicarbonate concentration used in this study (0.10 M), HO^\bullet
 222 formation rates are estimated to be ~ 390 times lower compared to $\text{SO}_4^{\bullet-}$ formation rates. The
 223 formed HO^\bullet reacts at pH 8 with both HCO_3^- and CO_3^{2-} , leading to the formation of $\text{CO}_3^{\bullet-}$
 224 (**Eqs. 2,3**). The reactions of HO^\bullet with the used organic compounds are expected to be fast
 225 (second-order rate constants up to $\sim 1 \times 10^{10} \text{ M}^{-1} \text{ s}^{-1}$; Buxton et al., 1988), but their
 226 importance for the transformation of these compounds is much lower compared to the
 227 aforementioned side-reactions of $\text{SO}_4^{\bullet-}$, owing to the much lower formation rate of HO^\bullet
 228 compared to $\text{SO}_4^{\bullet-}$. Therefore, the role of HO^\bullet would be negligible in the frame of the used
 229 conditions.

230

231 *2.2.1. Competition kinetics experiments*

232 Aqueous solutions containing potassium persulfate (1.0×10^{-3} M), sodium bicarbonate ($1.0 \times$
233 10^{-1} M), sulfuric acid ($\sim 0.8 \times 10^{-3}$ M), a single target compound TC (aniline or sulfadiazine,
234 1.0×10^{-6} M), the phenylurea metoxuron (MET) as a competitor (1.0×10^{-6} M), and a
235 variable concentration of either phenol or 4-methylphenol were prepared in quartz glass tubes
236 (18 mm external diameter, 15 mm internal diameter) using appropriate stock solutions of the
237 various components, with pH of 8.0 ± 0.1 . This pH was achieved by 1:5 dilution of a sodium
238 bicarbonate stock solution (5.0×10^{-1} M) that was supplemented with two drops of a
239 concentrated sulfuric acid solution (4.6 M). The tubes were closed with glass stoppers, shaken
240 by hand, placed in a water bath and kept 15 minutes at a temperature of 25.0 °C. They were
241 then submitted to irradiation in a temperature-controlled (25.0 ± 0.2 °C) DEMA (Hans
242 Mangels GmbH, Bornheim-Roisdorf, Germany) model 125 merry-go-round photoreactor,
243 which was equipped with a Heraeus Noblelight model TQ718 medium-pressure mercury lamp
244 operated at an input power of 500 W. The lamp was placed in a cooling jacket consisting of a
245 quartz glass inner wall, a borosilicate glass outer wall and a UVW-55 glass filter (supplied by
246 DEMA) in between. The combination of these glasses resulted in a band-pass filter for the
247 wavelength range of 308–410 nm. This setup was chosen, in analogy to previous studies
248 (Canonica and Schönenberger, 2019), to limit the direct phototransformation of the target
249 compounds and the competitor while allowing for photolysis of persulfate, mainly induced by
250 the 313 nm emission line of the mercury lamp. A more detailed description of the
251 photoreactor and its operation is given elsewhere (Wegelin et al., 1994). Samples (400 μ L)
252 were taken just before irradiation and during irradiation at regular time intervals, filled into
253 vials and immediately transferred to the refrigerated autosampler (5.0 °C) of the

254 chromatographic equipment. All kinetic experiments were performed at least in duplicate.
 255 Control irradiation experiments were performed with persulfate-free solutions to check for
 256 possible interferences by direct or indirect phototransformation reactions of the target
 257 compounds and metoxuron (MET). Within the time range employed for competition kinetics
 258 experiments, such transformation rates were typically two orders of magnitude lower than the
 259 rates induced by $\text{CO}_3^{\bullet-}$ in the absence of phenolic inhibitors, but became more important for
 260 the higher concentrations of inhibitor. These side-reactions were considered when fitting the
 261 data as explained in the Results and Discussion section.
 262 Apparent second-order rate constants for the reaction of $\text{CO}_3^{\bullet-}$ with a target compound (TC)
 263 leading to the transformation of TC, $k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr,app}}$, were determined based on the following
 264 competition kinetics expression (see **Text S3**, SM for its derivation):

$$265 \quad \ln \frac{[TC]_0}{[TC]} = \frac{k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr,app}}}{k_{\text{CO}_3^{\bullet-}, \text{MET}}^{\text{tr}}} \ln \frac{[MET]_0}{[MET]} \quad (18)$$

266 where $k_{\text{CO}_3^{\bullet-}, \text{MET}}^{\text{tr}}$ is the second-order rate constant for the reaction of $\text{CO}_3^{\bullet-}$ with the competitor
 267 MET, leading to the transformation of MET. The term *apparent* (“app” as a superscript in the
 268 rate constant) was used to allow for possible side-reactions in the transformation of TC. From
 269 the slope α of the regression line of $\ln \frac{[TC]_0}{[TC]}$ vs. $\ln \frac{[MET]_0}{[MET]}$ and the value of $k_{\text{CO}_3^{\bullet-}, \text{MET}}^{\text{tr}} = 8.1$
 270 $\times 10^7 \text{ M}^{-1}\text{s}^{-1}$ (value determined by laser flash photolysis; Canonica et al., 2005), the target
 271 compound rate constant was obtained as: $k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr,app}} = \alpha \times k_{\text{CO}_3^{\bullet-}, \text{MET}}^{\text{tr}}$. In consideration of the
 272 absence of inhibitory effect due to antioxidant species during the excited-triplet induced
 273 oxidation of phenylureas (Cannonica and Laubscher, 2008), we made the assumption that
 274 $k_{\text{CO}_3^{\bullet-}, \text{MET}}^{\text{tr}}$ remains unaffected by the presence of model antioxidants, such as the employed
 275 phenols, in the studied solutions.

276

277 *2.2.2. Inhibitory effect of antioxidants on the second-order rate constant of the carbonate*
278 *radical-induced oxidation of target compounds determined by competition kinetics.*

279 The previously elaborated one-channel model for the oxidation of TC inhibited by an
280 antioxidant (AO) (Canonica and Laubscher, 2008; Wenk et al., 2011) is applied in this section
281 to the derivation of the dependence of $k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr}}$ (see **Text S3**, SM for the definition of this
282 second-order rate constant) on AO concentration, in an analogous manner as performed in the
283 case of the sulfate radical as an oxidant (Canonica and Schönenberger, 2019). We consider
284 the case in which the transformation of TC is exclusively initiated by direct reaction with
285 $\text{CO}_3^{\bullet-}$ (i.e., side-reactions of TC can be neglected). The inhibitory effect of an AO on the
286 oxidation of TC is rationalized in terms of the following reaction equations:



290 where $\text{TC}^{\bullet+}$ is the reactive radical intermediate resulting from one-electron oxidation of TC,
291 $k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{et}}$ is the second-order rate constant for the electron transfer (et) reaction between TC
292 and $\text{CO}_3^{\bullet-}$, $k_{\text{TC}^{\bullet+}}^{\text{ox}}$ is the first-order rate constant for the transformation of $\text{TC}^{\bullet+}$ to a stable
293 oxidation product (TC_{ox}), and $k_{\text{TC}^{\bullet+}, \text{AO}}^{\text{red}}$ is the second-order rate constant for reaction of $\text{TC}^{\bullet+}$
294 with AO, which leads to $\text{TC}^{\bullet+}$ reduction back to TC and yields an oxidized antioxidant
295 (AO_{ox}). The rate equations for TC and $\text{TC}^{\bullet+}$ can be expressed as follows:

296
$$\frac{d[\text{TC}]}{dt} = -k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{et}} [\text{CO}_3^{\bullet-}] [\text{TC}] + k_{\text{TC}^{\bullet+}, \text{AO}}^{\text{red}} [\text{TC}^{\bullet+}] [\text{AO}] \quad (22)$$

297
$$\frac{d[TC^{•+}]}{dt} = k_{CO_3^{\bullet-}, TC}^{et} [CO_3^{\bullet-}] [TC] - k_{TC^{•+}, AO}^{red} [TC^{•+}] [AO] - k_{TC^{•+}}^{ox} [TC^{•+}] \quad (23)$$

298 Applying the steady-state assumption for $TC^{•+}$ leads to the following equation:

299
$$[TC^{•+}] = \frac{k_{CO_3^{\bullet-}, TC}^{et} [CO_3^{\bullet-}] [TC]}{k_{TC^{•+}}^{ox} + k_{TC^{•+}, AO}^{red} [AO]} \quad (24)$$

300 Substituting $[TC^{•+}]$ from **Eq. 24** into **Eq. 22** and rearranging yields **Eq. 25**:

301
$$\frac{d[TC]}{dt} = - \frac{k_{CO_3^{\bullet-}, TC}^{et}}{1 + (k_{TC^{•+}, AO}^{red} / k_{TC^{•+}}^{ox}) [AO]} [CO_3^{\bullet-}] [TC] \quad (25)$$

302 Comparing **Eqs. S5** (see **Text S3, SM**) and **25** leads to the identity:

303
$$k_{CO_3^{\bullet-}, TC}^{tr} = \frac{k_{CO_3^{\bullet-}, TC}^{et}}{1 + (k_{TC^{•+}, AO}^{red} / k_{TC^{•+}}^{ox}) [AO]} \quad (26)$$

304 Since AO is generally also transformed during the kinetic runs, $k_{CO_3^{\bullet-}, TC}^{tr}$ from **Eq. 26** is
 305 properly speaking not a kinetic constant. However, it can be approximated to a constant if
 306 $[AO]$ does not deviate strongly from $[AO]_0$ during irradiation. Note that in the case of $[AO] =$
 307 0 , $k_{CO_3^{\bullet-}, TC}^{tr} = k_{CO_3^{\bullet-}, TC}^{et}$.

308

309 **2.3. Back-reduction processes in triplet-sensitized phototransformation**

310 Additional experiments were carried out to investigate back-reduction by phenol during the
 311 degradation of 3,4-dichloroaniline (3,4DCA) and 3-chloroaniline (3CA) with excited triplet
 312 states. Back-reduction is known to be operational in the degradation of aniline and 4-
 313 chloroaniline by $^3CDOM^*$ (Canonica and Laubscher, 2008; Wenk and Canonica, 2012). The
 314 anionic form of 4-carboxybenzophenone (benzophenone-4-carboxylate, hereafter CBBP) was
 315 here used as CDOM proxy. Indeed, CBBP has already been employed to this purpose and to

316 determine representative second-order rate constants for the reactions between organic
317 substrates and ³CDOM* (Carena et al., 2019; Vione et al., 2018). Solutions containing CBBP
318 were irradiated under a UVA black lamp (Philips TL-D 18 W) with emission maximum at
319 369 nm, which produced a UV irradiance of $42.6 \pm 0.7 \text{ W m}^{-2}$ on top of the irradiated systems
320 (**Fig. S1**, SM). A detailed description of this irradiation setup is reported in Carena et al.
321 (2019). The initial concentrations of TC and CBBP were 5 and 70 μM , respectively, while
322 PhOH was varied between 0 and 10 μM . The pH of the solutions was ~ 7 and did not vary
323 significantly during irradiation. Buffers were not used in these experiments to avoid unwanted
324 side-reactions. The direct photolysis of 3,4DCA and 3CA was negligible under UVA
325 irradiation during the experimental time interval (4 h irradiation).

326

327 **2.4. Analytical Methods.**

328 The concentration of target compounds during kinetic runs was followed by high-performance
329 liquid chromatography (HPLC) using an Agilent 1100 system equipped with a quaternary
330 low-pressure mixing gradient pump, a refrigerated autosampler, a temperature-controlled
331 column compartment, a diode array detector and an Agilent 1200 fluorescence detector.
332 Alternatively, an analogous Thermo Fisher Ultimate 3000 HPLC system or a VWR-Hitachi
333 LaChrom Elite chromatograph (**Text S4**, SM) were used. A detailed list of HPLC analysis
334 methods is given in **Table S2**, SM.

335

336 3. Results and Discussion

337

338 3.1. Comparison of back-reduction in CO₃^{•-}-induced (nitrate photolysis system) and 339 triplet-sensitized transformation

340 The degradation of anilines by CO₃^{•-} (produced upon oxidation of HCO₃⁻/CO₃²⁻ by HO[•],
341 generated by nitrate photolysis) was inhibited when increasing the concentration of added
342 phenol (PhOH, **Fig. S2**, SM). The experimental conditions were chosen to minimize the
343 scavenging of either CO₃^{•-} or HO[•] by phenol, a process that would inhibit aniline degradation
344 in a similar way as back-reduction. By excluding scavenging of CO₃^{•-}/HO[•], it is possible to
345 arguably assume that the observed inhibition was due to the reactions (back-reduction)
346 between phenol and the one-electron oxidized intermediates formed upon transformation of
347 the anilines (**Scheme 1**).

348 Assuming $R_{TC} = R_{CO_3^{\bullet-}, TC} + R_{HO^{\bullet}, TC} + R_{d.p.}$ as per *Section 2.1.1* and taking into account the
349 possible back-reduction process, R_{TC} can be expressed as **Eq. 14** that was written in close
350 analogy to the kinetic equations used in previous works, to describe back-reduction in the
351 triplet-sensitized phototransformation of organic compounds (Vione et al., 2018; Wenk et al.,
352 2011; Wenk and Canonica, 2012).

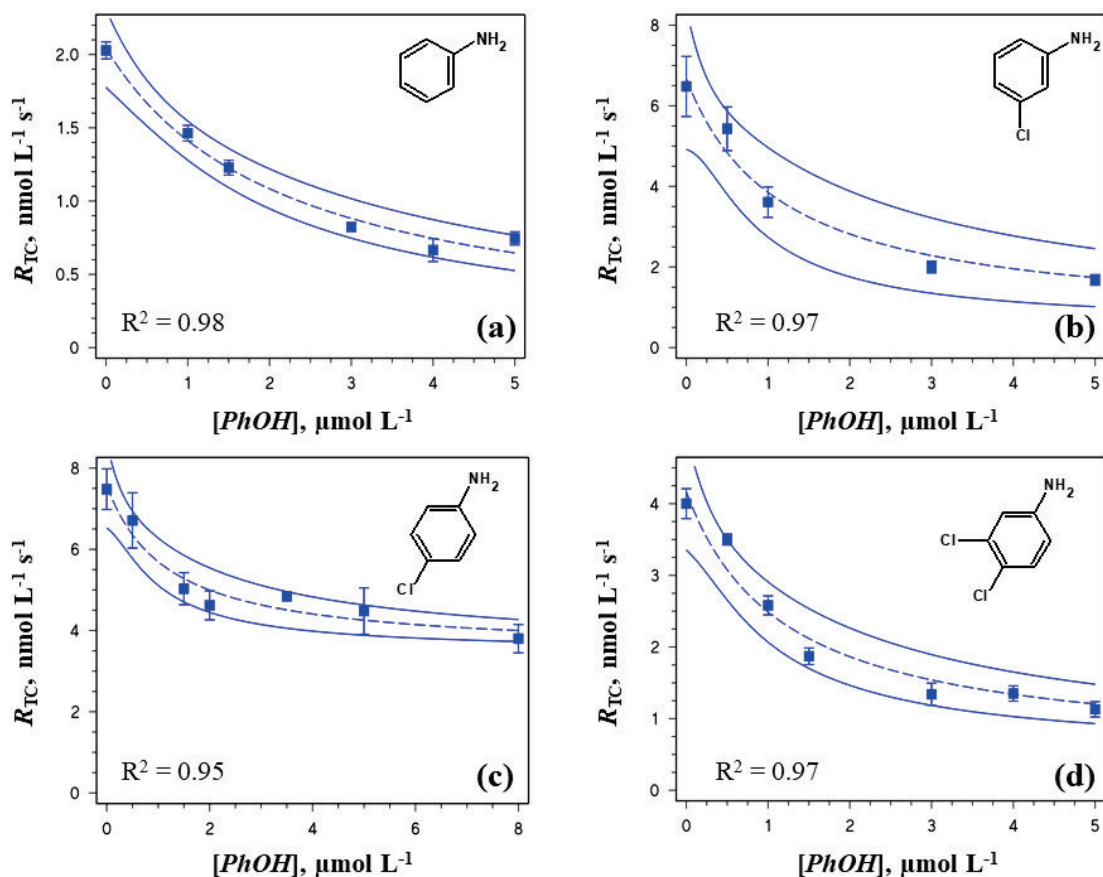
353 The rate of TC direct photolysis ($R_{d.p.}$) was assessed by correcting the direct photolysis rate
354 measured in ultrapure water ($R_{d.p.,exp}$; see **Fig. S2**, SM for the experimental data) for the inner-
355 filter effect of nitrate (the direct photolysis of TC is slower in the presence of nitrate, **Eq. 27**;
356 Carena et al., 2018).

$$357 R_{d.p.} = \frac{\int_{\lambda} p^{\circ}(\lambda) \times \frac{\varepsilon_{TC}(\lambda) \times [TC]}{\varepsilon_{NO_3^-}(\lambda) \times [NO_3^-] + \varepsilon_{TC}(\lambda) \times [TC]} \times \left[1 - 10^{-b \times (\varepsilon_{NO_3^-}(\lambda) \times [NO_3^-] + \varepsilon_{TC}(\lambda) \times [TC])} \right] d\lambda}{\int_{\lambda} p^{\circ}(\lambda) \times \left[1 - 10^{-\varepsilon_{TC}(\lambda) \times b \times [TC]} \right] d\lambda} \times R_{d.p.,exp} \quad (27)$$

358 In **Eq. 27**, $p^\circ(\lambda)$ is the UVB lamp photon flux density incident over the solution, $\varepsilon(\lambda)$ is a
359 molar absorption coefficient (of TC or NO_3^-), and b is the solution optical path length (0.4
360 cm). The correction factor was ~ 0.94 ($R_{d,p.} \approx 0.94 R_{d,p.,\text{exp}}$). Note that $R_{d,p.}$ was here
361 considered not to vary with $[\text{PhOH}]$, because the direct photolysis of chloroanilines should
362 occur *via* loss of chlorine atoms rather than through photoionization of the amino group
363 (Carena et al., 2018). Consequently, back reduction by PhOH should not affect the direct
364 photolysis process.

365 **Fig. 1** shows the experimental profiles of R_{TC} vs. $[\text{PhOH}]$ for the studied anilines. A very
366 good fit to the experimental data was obtained with **Eq. 14**, which suggests that the kinetic
367 model proposed for back-reduction matches the experimental findings. The calculated $R_{d,p.}$
368 values (**Eq. 27**) were $0.49 \pm 0.03 \text{ nM s}^{-1}$ for 3CA, $3.4 \pm 0.5 \text{ nM s}^{-1}$ for 4CA, and 0.44 ± 0.02
369 nM s^{-1} for 3,4DCA. Direct photolysis was negligible in the case of Ani. The values of R_{tHO}
370 (**Eq. 14**) obtained from data fit were $2.1 \pm 0.1 \text{ nM s}^{-1}$ for Ani, $6.3 \pm 0.4 \text{ nM s}^{-1}$ for 3CA, $4.3 \pm$
371 0.5 nM s^{-1} for 4CA, and $3.9 \pm 0.3 \text{ nM s}^{-1}$ for 3,4DCA. Data fit also yielded $[\text{PhOH}]_{1/2} = 2.19 \pm$
372 $0.23 \text{ }\mu\text{M}$ for Ani, $1.15 \pm 0.25 \text{ }\mu\text{M}$ for 3CA, $1.18 \pm 0.26 \text{ }\mu\text{M}$ for 4CA, and $1.18 \pm 0.22 \text{ }\mu\text{M}$ for
373 3,4DCA.

374 Interestingly, **Fig. 2** shows that the values of $[\text{PhOH}]_{1/2}$ obtained for the degradation of the
375 anilines by $\text{CO}_3^{\cdot-}$ do not significantly differ from those observed in triplet-sensitized
376 oxidation induced by the triplet states of benzophenone-4-carboxylate ($^3\text{CBBP}^*$) and 2-
377 acetophenone ($^3\text{2AN}^*$). This finding suggests that $\text{CO}_3^{\cdot-}$, $^3\text{CBBP}^*$ and $^3\text{2AN}^*$ produce the
378 same intermediates upon aniline oxidation, which all undergo a similar back-reduction
379 process in the presence of PhOH.

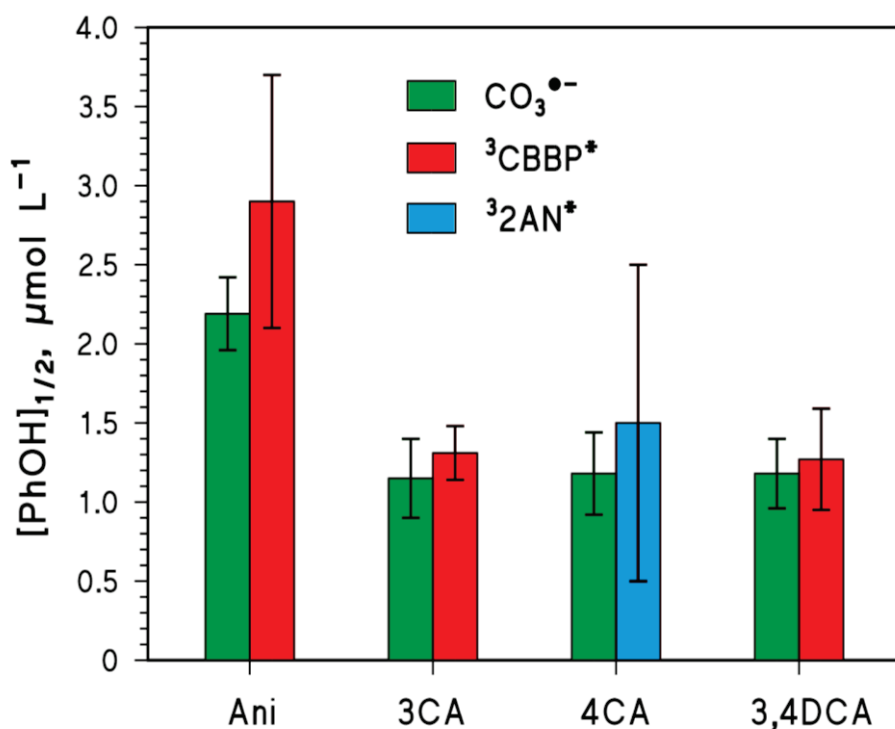


380

381

382 **Figure 1.** Experimental degradation rates of (a) Ani, (b) 3CA, (c) 4CA and (d) 3,4DCA (initial
 383 concentration 5 μM), as a function of the concentration of added phenol. The dashed curves represent
 384 the fit results with **Eq. 14**, while the solid curves are the 95% confidence bands of the fit. The R^2
 385 parameter shows the goodness of the fit. The error bounds to the R_{TC} data represent the associated
 386 sigma-level uncertainty. The irradiated solutions (pH 8.3) contained $\text{NaNO}_3 + \text{NaHCO}_3$ and were
 387 exposed to UVB radiation. See **Table S1**, SM for the detailed experimental conditions.

388



389

390

391 **Figure 2.** Comparison between the $[PhOH]_{1/2}$ values measured for anilines degradation by $\text{CO}_3^{\bullet-}$, and
 392 those observed for $^3\text{CBBP}^*$ - and $^3\text{2AN}^*$ -induced oxidation. Data concerning the reaction $4\text{CA} +$
 393 $^3\text{2AN}^*$ were taken from Wenk and Canonica (2012), those for $\text{Ani} + ^3\text{CBBP}^*$ were taken from
 394 Canonica and Laubscher (2008), while those for $3\text{CA}/3,4\text{DCA} + ^3\text{CBBP}^*$ are from this work (**Text**
 395 **S5**, SM). Error bars represent standard errors (in the case of data from this work, they represent the
 396 goodness of the fit of the experimental data with **Eq. S10** (SM).

397

398

399 **3.2. Back-reduction processes upon generation of $\text{CO}_3^{\bullet-}$ by $\text{SO}_4^{\bullet-}$ as reactive** 400 **intermediate**

401 Initially, competition kinetics experiments were performed for individual target compounds
 402 (TC), i.e., aniline and sulfadiazine, in the absence of phenolic inhibitors. Analogous
 403 experiments were also executed for phenol and 4-methylphenol as antioxidant compounds.

404 These experiments yielded second-order rate constants for the reaction of TC with $\text{CO}_3^{\bullet-}$ in
405 the absence of additives, which are termed as $k_{\text{CO}_3^{\bullet-},\text{TC}}^{\text{tr}}$ and provided in **Table 1**. The
406 determined $k_{\text{CO}_3^{\bullet-},\text{TC}}^{\text{tr}}$ values are within a factor of two compared to those known from the
407 literature.

408 Interestingly, the determined $\text{CO}_3^{\bullet-}$ rate constants for phenol and 4-methylphenol are at least
409 11 and 3 times lower, respectively, compared to the rate constants for aniline and sulfadiazine.
410 This represents an obvious experimental advantage, since these phenols are more resistant to
411 direct $\text{CO}_3^{\bullet-}$ -induced transformation, and consequently expected to persist and exert their
412 inhibitory effect during the whole course of TC transformation. Examples of kinetics runs
413 showing the transformation of aniline as TC and metoxuron (MET) as competitor in the
414 presence and absence of 0.5 μM phenol as inhibitor are provided in **Figure 3a**, which also
415 shows the transformation kinetics of phenol. All compounds exhibit zero-order kinetics, i.e., a
416 linear decrease in their residual concentration with irradiation time. For the system with
417 aniline and metoxuron without added phenol, this is an indication that aniline, which reacts
418 much faster than metoxuron, is the major species responsible for the scavenging of $\text{CO}_3^{\bullet-}$
419 during the whole kinetic run, and that no reaction products significantly reacting with $\text{CO}_3^{\bullet-}$
420 are formed. Indeed, if $R_{\text{CO}_3^{\bullet-}}$ is the formation rate of $\text{CO}_3^{\bullet-}$, if $R_{\text{TC}} = k_{\text{CO}_3^{\bullet-},\text{TC}}^{\text{tr}} [\text{TC}] [\text{CO}_3^{\bullet-}]$ and
421 if $\text{CO}_3^{\bullet-}$ mainly reacts with TC, it is $[\text{CO}_3^{\bullet-}] \cong R_{\text{CO}_3^{\bullet-}} (k_{\text{CO}_3^{\bullet-},\text{TC}}^{\text{tr}} [\text{TC}])^{-1}$ and $R_{\text{TC}} \cong R_{\text{CO}_3^{\bullet-}}$, which
422 means that R_{TC} is independent of $[\text{TC}]$ (zero-order kinetics).

423

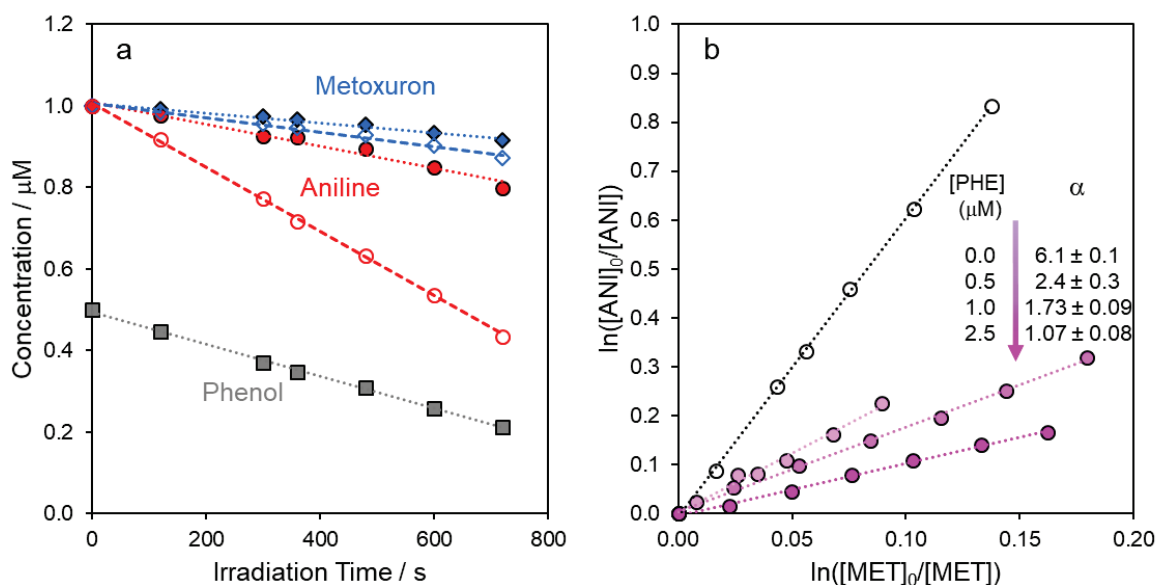
424 **Table 1.** Second-order rate constants for the reaction of the carbonate radical with target
 425 compounds, determined in this study by competition kinetics at pH 8.0±0.1 ^a

Target compound	α^b	$k_{\text{CO}_3^{\cdot-}, \text{TC}}^{\text{tr}} / 10^7 \text{ M}^{-1} \text{ s}^{-1}$		$\text{p}K_{\text{a}}^c$ (literature values)
		this study ^d	literature values ^e	
Aniline	6.12 ± 0.13	50 ± 1	46 – 67	4.60
Sulfadiazine	6.80 ± 0.10	55 ± 1	27.8, 28	6.5, ^f 6.4 ^g
Phenol	0.435 ± 0.008	3.5 ± 0.2	0.49 – 2.2 ^h	9.99
4-Methylphenol	1.58 ± 0.03	12.8 ± 0.2	15 ⁱ	10.26

426

427 ^a Composition of the solutions used for competition kinetics experiments: initial target compound and
 428 metoxuron concentrations ($1.0 \times 10^{-6} \text{ M}$); potassium persulfate ($1.0 \times 10^{-3} \text{ M}$); sodium bicarbonate
 429 ($1.0 \times 10^{-1} \text{ M}$); sulfuric acid ($\sim 0.8 \times 10^{-3} \text{ M}$). ^b Slope α of the competition kinetics plots according to
 430 **Eq. 18**, obtained by linear regression, and corresponding to $k_{\text{CO}_3^{\cdot-}, \text{TC}}^{\text{tr}} / k_{\text{CO}_3^{\cdot-}, \text{MET}}^{\text{tr}}$. The average values
 431 from at least two independent experiments are given. The errors represent 95% confidence intervals
 432 obtained from linear regression and by applying Gaussian error propagation. ^c $\text{p}K_{\text{a}}$ values (from Dean,
 433 1999, except where noted) refer to the following deprotonation equilibria: a) for aniline (An):
 434 AnH^+/An ; b) for sulfadiazine (SD): $\text{SD}/\text{SD}(-\text{H})^-$ (i.e., second $\text{p}K_{\text{a}}$ or $\text{p}K_{\text{a}2}$); for the phenols (R-PhOH):
 435 $\text{R-PhOH}/\text{R-PhO}^-$. ^d Calculated as $\alpha \times k_{\text{CO}_3^{\cdot-}, \text{MET}}^{\text{tr}}$, where the second-order rate constant for the reaction
 436 of the carbonate radical with metoxuron was set as $k_{\text{CO}_3^{\cdot-}, \text{MET}}^{\text{tr}} = 8.1 \times 10^7 \text{ M}^{-1} \text{ s}^{-1}$. The error in
 437 $k_{\text{CO}_3^{\cdot-}, \text{MET}}^{\text{tr}}$ ($0.6 \times 10^7 \text{ M}^{-1} \text{ s}^{-1}$; Canonica et al., 2005) was not considered in the calculation of the errors,
 438 which represent 95% confidence intervals. ^e Range of values from Wojnárovits et al., 2020. ^f Boreen et
 439 al., 2005. ^g Ricci and Cross, 1993. ^h At pH 7.0 or 8.0. ⁱ At pH 8.3.

440



441
 442 **Figure 3.** (a) Kinetic runs for aniline (red circles) and metoxuron (blue diamonds) in the presence
 443 (filled symbols) and absence (open symbols) of 0.5 μM phenol (grey filled squares). The lines
 444 represent linear regressions. (b) Competition kinetics plots for aniline (1.0 μM initial concentration) as
 445 target compound and metoxuron (1.0 μM initial concentration) as competitor, for several different
 446 phenol concentrations indicated numerically. The intensity of the filling of the markers increases with
 447 increasing phenol concentration. The lines represent linear regressions of the various experimental
 448 runs, and their slope α is shown numerically as a function of phenol concentration.

449
 450
 451 In the absence of phenol, the depletion of both aniline and MET is faster than in its presence,
 452 but the reduction in depletion rate is more important for aniline. **Figure 3b** displays
 453 competition kinetics examples for the same compounds and different concentrations of added
 454 phenol. The slope of the regression lines (defined above as α and given numerically in the
 455 graph as a function of the concentration of added phenol) decreases with increasing phenol
 456 concentration, indicating a concomitantly decreasing $k_{\text{CO}_3^{\cdot-}, \text{aniline}}^{\text{tr,app}}$. The trend in α shown in
 457 **Figure 3b** clearly demonstrates the inhibitory effect of phenol on the transformation of aniline

458 induced by $\text{CO}_3^{\bullet-}$. Based on **Eq. 26**, we define here the inhibition factor, IF , in analogy to
 459 previous studies as the ratio between $k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr,app}}$ and $k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr}}$, where the latter is the second-order
 460 rate constant for the transformation of TC induced by $\text{CO}_3^{\bullet-}$ determined without antioxidant
 461 addition.

$$462 \quad IF([AO]) = k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr,app}} / k_{\text{CO}_3^{\bullet-}, \text{TC}}^{\text{tr}} \quad (28)$$

463 Substituting the right-hand term of **Eq. 26** into **Eq. 28** yields:

$$464 \quad IF([AO]) = \frac{1}{1 + (k_{\text{TC}^{\bullet+}, \text{AO}}^{\text{red}} / k_{\text{TC}^{\bullet+}}^{\text{ox}}) \times [AO]} \quad (29)$$

465 With $[AO]_{1/2} = k_{\text{TC}^{\bullet+}}^{\text{ox}} / k_{\text{TC}^{\bullet+}, \text{AO}}^{\text{red}}$, **Eq. 29** transforms to:

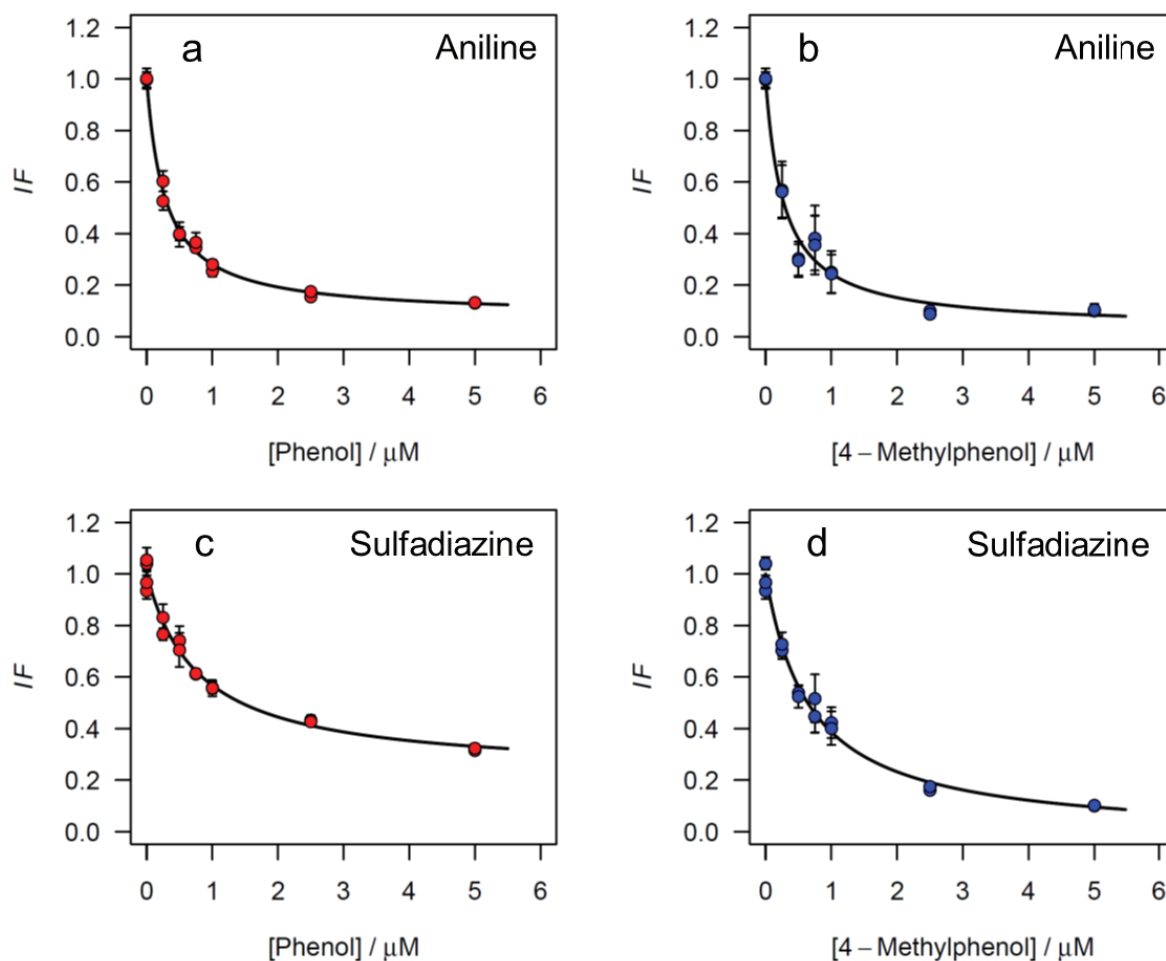
$$466 \quad IF([AO]) = \frac{1}{1 + [AO]/[AO]_{1/2}} \quad (30)$$

467 Series of competition kinetics experiments were performed for the two target compounds,
 468 namely aniline and sulfadiazine, using phenol and, alternatively, 4-methylphenol as
 469 antioxidants. The results of these experiments are shown in **Figure 4** in terms of inhibition
 470 factor as a function of AO concentration. For all four target compound/antioxidant pairs, a
 471 decrease in IF with increasing $[AO]$ is observed. The four IF data series were fitted to the
 472 following **Eq. 31**, which is a slightly modified form of **Eq. 30** and is used to account for
 473 possible side-reactions of the target compounds in analogy to previous studies (Wenk et al.,
 474 2011; Wenk and Canonica, 2012).

$$475 \quad IF([AO]) = \frac{f}{1 + [AO]/[AO]_{1/2}} + (1 - f) \quad (31)$$

476 Here f is the fraction of TC reacting through the main pathway that is inhibited by
 477 antioxidants. For aniline, fits yielded identical $[AO]_{1/2}$ values for phenol ($0.28 \pm 0.02 \mu\text{M}$) and
 478 4-methylphenol ($0.28 \pm 0.04 \mu\text{M}$). For sulfadiazine, fitted $[AO]_{1/2}$ values were $0.79 \pm 0.10 \mu\text{M}$

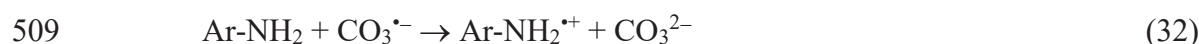
479 for phenol and $0.67 \pm 0.06 \mu\text{M}$ for 4-methylphenol, and therefore higher compared to aniline
480 by a factor of 2.4 and 2.8, respectively. Obtained f values were high, namely 0.78 for the pair
481 sulfadiazine/phenol and > 0.92 for all other TC/AO pairs, indicating the minor importance of
482 side-reactions not subject to inhibition by AO.
483



484
485
486 **Figure 4.** Inhibition factors (IF) for aniline (**a, b**) and sulfadiazine (**c, d**) determined at pH 8.0 in
487 duplicate at various concentrations of added phenol (left panels) and 4-methylphenol (right panels),
488 respectively, in the range of 0.0 – 5.0 μM . Error bars correspond to 95% confidence intervals obtained
489 from linear regression by applying **Eq. 18** and Gaussian error propagation rules. Black lines are best
490 fitting functions according to **Eq. 31**.
491

492 In the case of aniline, it is interesting to observe that the value of $[PhOH]_{1/2}$ (AO = phenol) is
 493 lower by an order of magnitude compared to the value obtained when $CO_3^{\bullet-}$ was produced by
 494 nitrate photolysis (see **Section 3.1**). It is also an order of magnitude lower compared to the
 495 values determined for triplet-sensitized aniline transformation, conducted using CBBP and 2-
 496 AN as photosensitizers (Canonica and Laubscher, 2008; Wenk and Canonica, 2012). All these
 497 $[PhOH]_{1/2}$ values were obtained from experiments carried out at very similar pH values (8.0 –
 498 8.3). Therefore, the slight changes in speciation of the various species involved in the back-
 499 reduction process are probably not the cause of the differences in $[PhOH]_{1/2}$.

500 Anilines (Ar-NH₂) undergo mono-electronic oxidation by $CO_3^{\bullet-}$ (Elango et al., 1984; Huang
 501 and Mabury, 2000a; Jonsson et al., 1994; Liu et al., 2018; Wojnárovits et al., 2020) and by the
 502 triplet states of 2-acetonaphthone (³2AN*) and benzophenone-4-carboxylate (³CBBP*)
 503 (Canonica and Laubscher, 2008; Vione et al., 2018; Wenk and Canonica, 2012), forming
 504 radical cations (Ar-NH₂^{•+}) that easily deprotonate to produce Ar-NH[•]. Previous works have
 505 shown that phenolic compounds including phenol (PhOH) can act as antioxidants for the
 506 species Ar-NH[•] (Canonica and Laubscher, 2008; Vione et al., 2018; Wenk and Canonica,
 507 2012). Therefore, the following reactions should take place in the studied systems (where T*
 508 represents an excited triplet state, such as that of ³2AN*):



514 A possibility to account for differences in the $[PhOH]_{1/2}$ values observed in different systems
 515 is to assume that antioxidant compounds are produced even in the absence of added PhOH,

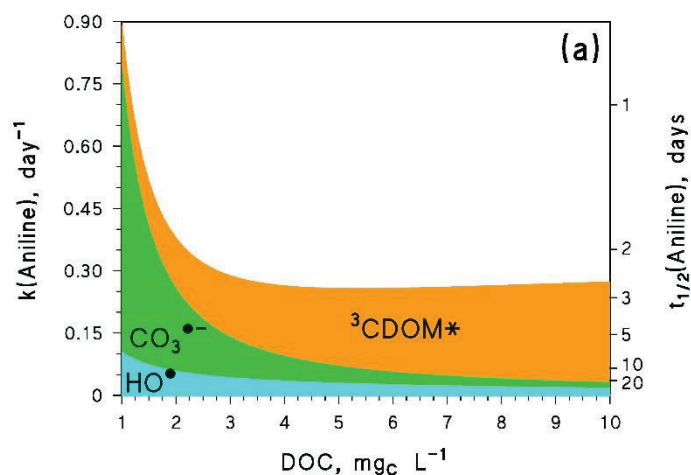
516 thereby providing a baseline level of back-reduction (Canonica and Schönenberger, 2019).
517 Such a process would decrease the steady-state $[Ar-NH^*]$ with the consequence that, the
518 higher the baseline reductants, the higher is the concentration of PhOH (quantified as
519 $[PhOH]_{1/2}$) required to obtain the same IF value. Therefore, different systems could show
520 varying levels of baseline back-reduction and provide different values of $[PhOH]_{1/2}$. A
521 potential photogenerated reductant is superoxide, $O_2^{\bullet-}$, which is for instance produced by
522 reaction between $T^{\bullet-}$ and dissolved oxygen (Canonica and Schönenberger, 2019; Huber et al.,
523 2003). Moreover, the one-electron oxidation of anilines is well known to produce phenolic
524 intermediates (Bossmann et al., 1998) that might act as baseline antioxidants.

525 As far as back-reduction in sulfadiazine transformation is concerned, the $[PhOH]_{1/2}$ value
526 found in this study ($0.79 \pm 0.10 \mu\text{M}$) is well matched with $[PhOH]_{1/2} = 0.86 \pm 0.10 \mu\text{M}$ that
527 was obtained for the transformation of sulfadiazine photosensitized by CBBP at pH 8.5
528 (Vione et al., 2018), but four times lower than a previously determined value ($3.2 \pm 0.8 \mu\text{M}$)
529 for triplet-sensitized transformation at pH 8.0 (Wenk and Canonica, 2012). Therefore, for
530 sulfadiazine the presence of possible baseline antioxidants appears to play a less important
531 role than in the case of aniline.

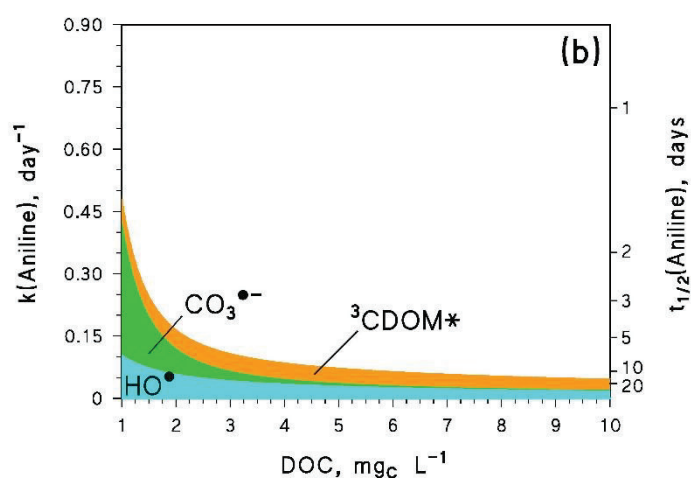
532

533 **3.3. Environmental significance**

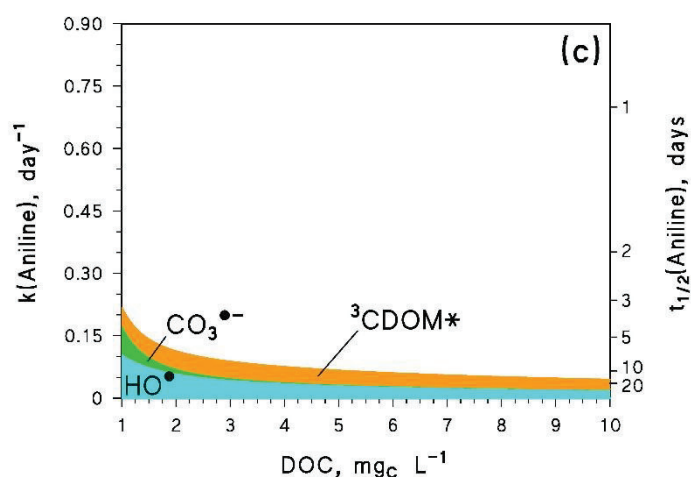
534 To assess the role of back-reduction on the phototransformation of water pollutants, the
535 photochemical fate of aniline was modeled in a lake-water scenario by means of the APEX
536 software (**Figure 5**) (Vione, 2020).



537



538



539

540 **Figure 5.** Modeled photochemical pseudo-first-order degradation rate constant (left Y-axis) and half-
 541 life time (right Y-axis) of aniline in lake water, as a function of the DOC (dissolved organic carbon).
 542 (a) Scenario 1 (no back reduction); (b) Scenario 2 (back reduction for ${}^3\text{CDOM}^*$, moderate back
 543 reduction for $\text{CO}_3^{\bullet-}$); (c) Scenario 3 (back reduction for ${}^3\text{CDOM}^*$, strong back reduction for $\text{CO}_3^{\bullet-}$).
 544 Other conditions: $100 \mu\text{mol L}^{-1} \text{NO}_3^-$, $1 \mu\text{mol L}^{-1} \text{NO}_2^-$, 2mmol L^{-1} alkalinity, pH 8.0, 1 m depth,
 545 sunlight irradiance as per mid-July, mid latitude, northern hemisphere.

546 Aniline was chosen here because two alternative values of $[PhOH]_{1/2}$ were obtained in this
547 work for this compound, which could be observed under different conditions as far as
548 transformation by $CO_3^{\bullet-}$ is concerned. Aniline does not undergo direct photolysis to a
549 significant extent, but it reacts with HO^{\bullet} (second-order rate constant $k_{HO^{\bullet},Ani}^{tr} = 1.4 \times 10^{10} M^{-1}$
550 s^{-1} ; Buxton et al., 1988), $CO_3^{\bullet-}$ ($k_{CO_3^{\bullet-},Ani}^{tr} = 5 \times 10^8 M^{-1} s^{-1}$, this work) and ${}^3CDOM^*$
551 ($k_{{}^3CDOM^*,Ani}^{tr} = 4 \times 10^9 M^{-1} s^{-1}$; estimated from values published by Erickson et al., 2015, for
552 several anilines using methylene blue as model photosensitizer). In the case of natural waters,
553 the natural organic matter is the provider of phenolic antioxidant moieties. The experimental
554 $[PhOH]_{1/2}$ values can be translated into $DOC_{1/2}$ equivalents as $DOC_{1/2} (mg_C L^{-1}) = 0.4 \times [PhOH]_{1/2}$
555 (μM), as reported by Vione et al. (2018) (see also Leresche et al., 2016). In the case of aniline
556 + HO^{\bullet} , no back-reduction is expected (Wenk et al., 2011). In the case of aniline + ${}^3CDOM^*$,
557 one has $[PhOH]_{1/2} = 2.8 \mu M \Rightarrow DOC_{1/2} = 1.1 mg_C L^{-1}$ (Wenk and Canonica, 2012). In the case
558 of aniline + $CO_3^{\bullet-}$ (this work), one has either $[PhOH]_{1/2} = 0.28 \mu M \Rightarrow DOC_{1/2} = 0.11 mg_C L^{-1}$,
559 or $[PhOH]_{1/2} = 2.2 \mu M \Rightarrow DOC_{1/2} = 0.88 mg_C L^{-1}$.

560 Here, three different scenarios are considered. In the first (Scenario 1), no back-reduction is
561 supposed to be operational with either ${}^3CDOM^*$ or $CO_3^{\bullet-}$. In Scenario 2, we used $DOC_{1/2} = 1.1$
562 $mg_C L^{-1}$ for the ${}^3CDOM^*$ process and $DOC_{1/2} = 0.88 mg_C L^{-1}$ for $CO_3^{\bullet-}$ (moderate back-
563 reduction for $CO_3^{\bullet-}$). In Scenario 3, we used again $DOC_{1/2} = 1.1 mg_C L^{-1}$ for ${}^3CDOM^*$ but
564 $DOC_{1/2} = 0.11 mg_C L^{-1}$ for $CO_3^{\bullet-}$ (strong back-reduction for $CO_3^{\bullet-}$).

565 **Figure 5** shows that, in the absence of back-reduction (**5a**), $CO_3^{\bullet-}$ is the main aniline
566 photodegradation process at low DOC while ${}^3CDOM^*$ prevails at high DOC (indeed, $CO_3^{\bullet-}$ is
567 very efficiently scavenged by DOM; Yan et al., 2019). Reaction with HO^{\bullet} always plays a
568 secondary role. The back-reduction processes increase the relative role of HO^{\bullet} , and the HO^{\bullet}

569 reaction becomes a very important if not the main pathway of aniline phototransformation. In
570 the most extreme back-reduction scenario (Scenario 3, **Figure 5c**), the reaction with HO[•]
571 would prevail at low DOC and would play a comparable role as ³CDOM* at high DOC. The
572 back-reduction processes also slow down significantly the photodegradation kinetics of
573 aniline, with an important increase in the relevant lifetimes.

574

575

576 **4. Conclusions**

577

- 578 • Back-reduction by phenolic compounds is operational in the case of the CO₃^{•-}-mediated
579 photodegradation of anilines and sulfadiazine.
- 580 • The observed similarities concerning back-reduction for CO₃^{•-}-induced and triplet-
581 sensitized degradation suggest that the same reaction intermediates are involved in both
582 cases.
- 583 • The operational conditions by which CO₃^{•-} is photogenerated (photolysis of nitrate or
584 persulfate) affect the experimental values of [AO]_{1/2} for aniline, presumably as a
585 consequence of the different formation of additional reducing species in different systems.
- 586 • In sunlit surface waters scenarios, back-reduction decreases the contributions of both
587 CO₃^{•-} and ³CDOM* to aniline photodegradation. In the absence of back-reduction, HO[•]
588 would always be a secondary process and aniline photodegradation would be dominated
589 by CO₃^{•-} at low DOC and by ³CDOM* at high DOC. In contrast, HO[•] and ³CDOM* play
590 comparable roles at high DOC with back-reduction. Moreover, in the strongest back-
591 reduction scenario for the CO₃^{•-} process, aniline photodegradation at low DOC would be
592 dominated by HO[•].

593

594 **References**

595

596 Boreen, A.L., Arnold, W.A., McNeill, K., 2005. Triplet-Sensitized Photodegradation of Sulfa
597 Drugs Containing Six-Membered Heterocyclic Groups: Identification of an SO₂
598 Extrusion Photoproduct. *Environ. Sci. Technol.* 39, 3630–3638.
599 <https://doi.org/10.1021/es048331p>

600 Boreen, A.L., Arnold, W.A., McNeill, K., 2003. Photodegradation of pharmaceuticals in the
601 aquatic environment: A review. *Aquat. Sci.* 65, 320–341.
602 <https://doi.org/10.1007/s00027-003-0672-7>

603 Bossmann, S.H., Oliveros, E., Göb, S., Siegwart, S., Dahlen, E.P., Payawan, L., Straub, M.,
604 Wörner, M., Braun, A.M., 1998. New Evidence against Hydroxyl Radicals as Reactive
605 Intermediates in the Thermal and Photochemically Enhanced Fenton Reactions. *J. Phys.*
606 *Chem. A* 102, 5542–5550. <https://doi.org/10.1021/jp980129j>

607 Buxton, G. V, Greenstock, C.L., Helman, P.W., Ross, A.B., 1988. Critical Review of rate
608 constants for reactions of hydrated electrons, hydrogen atoms and hydroxyl radicals
609 ($\cdot\text{OH}/\cdot\text{O}^-$) in Aqueous Solution. *J. Phys. Chem. Ref. Data* 17, 513–886.
610 <https://doi.org/https://doi.org/10.1063/1.555805>

611 Canonica, S., Kohn, T., Mac, M., Real, F.J., Wirz, J., Von Gunten, U., 2005. Photosensitizer
612 method to determine rate constants for the reaction of carbonate radical with organic
613 compounds. *Environ. Sci. Technol.* 39, 9182–9188. <https://doi.org/10.1021/es051236b>

614 Canonica, S., Laubscher, H.U., 2008. Inhibitory effect of dissolved organic matter on triplet-
615 induced oxidation of aquatic contaminants. *Photochem. Photobiol. Sci.* 7, 547–551.
616 <https://doi.org/10.1039/b719982a>

617 Canonica, S., Schönenberger, U., 2019. Inhibitory Effect of Dissolved Organic Matter on the
618 Transformation of Selected Anilines and Sulfonamide Antibiotics Induced by the Sulfate

619 Radical. Environ. Sci. Technol. 53, 11783–11791.
620 <https://doi.org/10.1021/acs.est.9b04105>

621 Carena, L., Proto, M., Minella, M., Ghigo, G., Giovannoli, C., Brigante, M., Mailhot, G.,
622 Maurino, V., Minero, C., Vione, D., 2018. Evidence of an Important Role of
623 Photochemistry in the Attenuation of the Secondary Contaminant 3,4-Dichloroaniline in
624 Paddy Water. Environ. Sci. Technol. 52, 6334–6342.
625 <https://doi.org/10.1021/acs.est.8b00710>

626 Carena, L., Puscasu, C.G., Comis, S., Sarakha, M., Vione, D., 2019. Environmental
627 photodegradation of emerging contaminants: A re-examination of the importance of
628 triplet-sensitised processes, based on the use of 4-carboxybenzophenone as proxy for the
629 chromophoric dissolved organic matter. Chemosphere 237, 124476.
630 <https://doi.org/10.1016/j.chemosphere.2019.124476>

631 Criquet, J., Leitner, N.K.V., 2009. Degradation of acetic acid with sulfate radical generated by
632 persulfate ions photolysis. Chemosphere 77, 194–200.
633 <https://doi.org/https://doi.org/10.1016/j.chemosphere.2009.07.040>

634 Dean, J.A., 1999. Lange's Handbook of Chemistry. 15th ed, 15th ed. ed. McGraw-Hill: New
635 York, Etc.

636 Elango, T.P., Ramakrishnan, V., Vancheesan, S., Kuriacose, J.C., 1984. Reaction of the
637 carbonate radical with substituted anilines. Proc. Indian Acad. Sci. - Chem. Sci. 93, 47–
638 52. <https://doi.org/10.1007/BF02841982>

639 Erickson, P.R., Walpen, N., Guerard, J.J., Eustis, S.N., Arey, J.S., McNeill, K., 2015.
640 Controlling Factors in the Rates of Oxidation of Anilines and Phenols by Triplet
641 Methylene Blue in Aqueous Solution. J. Phys. Chem. A 119, 3233–3243.
642 <https://doi.org/10.1021/jp511408f>

643 Galbavy, E.S., Ram, K., Anastasio, C., 2010. Chemistry 2-Nitrobenzaldehyde as a chemical

644 actinometer for solution and ice photochemistry. *J. Photochem. Photobiol. A Chem.* 209,
645 186–192. <https://doi.org/10.1016/j.jphotochem.2009.11.013>

646 Hao, Z., Ma, J., Miao, C., Song, Y., Lian, L., Yan, S., Song, W., 2020. Carbonate Radical
647 Oxidation of Cylindrospermopsin (Cyanotoxin): Kinetic Studies and Mechanistic
648 Consideration. *Environ. Sci. Technol.* 54, 10118–10127.
649 <https://doi.org/10.1021/acs.est.0c03404>

650 Herrmann, H., Reese, A., Zellner, R., 1995. Time-resolved UV/VIS diode array absorption
651 spectroscopy of SO_x^- ($x=3, 4, 5$) radical anions in aqueous solution. *J. Mol. Struct.* 348,
652 183–186. [https://doi.org/https://doi.org/10.1016/0022-2860\(95\)08619-7](https://doi.org/https://doi.org/10.1016/0022-2860(95)08619-7)

653 Huang, J., Mabury, S.A., 2000a. A New Method for Measuring Carbonate Radical Reactivity
654 toward Pesticides. *Environ. Toxicol. Chem.* 19, 1501–1507.
655 <https://doi.org/https://doi.org/10.1002/etc.5620190605>

656 Huang, J., Mabury, S.A., 2000b. Steady-state concentrations of carbonate radicals in field
657 waters. *Environ. Toxicol. Chem.* 19, 2181–2188.
658 <https://doi.org/https://doi.org/10.1002/etc.5620190906>

659 Huie, R.E., Clifton, C.L., 1990. Temperature dependence of the rate constants for reactions of
660 the sulfate radical, SO_4^- , with anions. *J. Phys. Chem.* 94, 8561–8567.
661 <https://doi.org/10.1021/j100386a015>

662 Jonsson, M., Lind, J., Eriksen, T.E., Merényi, G., 1994. Redox and Acidity Properties of 4-
663 Substituted Aniline Radical Cations in Water. *J. Am. Chem. Soc.* 116, 1423–1427.
664 <https://doi.org/10.1021/ja00083a030>

665 Liu, T., Yin, K., Liu, C., Luo, J., Crittenden, J., Zhang, W., Luo, S., He, Q., Deng, Y., Liu, H.,
666 Zhang, D., 2018. The role of reactive oxygen species and carbonate radical in
667 oxcarbazepine degradation via UV , UV/H₂O₂ : Kinetics , mechanisms and toxicity
668 evaluation. *Water Res.* 147, 204–213. <https://doi.org/10.1016/j.watres.2018.10.007>

669 Mack, J., Bolton, J.R., 1999. Photochemistry of nitrite and nitrate in aqueous solution: a
670 review. *J. Photochem. Photobiol. A Chem.* 128, 1–13.
671 [https://doi.org/https://doi.org/10.1016/S1010-6030\(99\)00155-0](https://doi.org/https://doi.org/10.1016/S1010-6030(99)00155-0)

672 Millero, F.J., Pierrot, D., Lee, K., Wanninkhof, R., Feely, R., Sabine, C.L., Key, R.M.,
673 Takahashi, T., 2002. Dissociation constants for carbonic acid determined from field
674 measurements. *Deep. Res. Part I Oceanogr. Res. Pap.* 49, 1705–1723.
675 [https://doi.org/https://doi.org/10.1016/S0967-0637\(02\)00093-6](https://doi.org/https://doi.org/10.1016/S0967-0637(02)00093-6)

676 Neta, P., Huie, R.E., Ross, A.B., 1988. Rate Constants for Reactions of Inorganic Radicals in
677 Aqueous Solution. *J. Phys. Chem. Ref. Data* 17, 1027.
678 <https://doi.org/https://doi.org/10.1063/1.555808>

679 NIST, 2004. Critically Selected Stability Constants of Metal Complexes Database. Stand. Ref.
680 Data Program, Vol.46. Natl. Inst. Stand. Technol. U.S. Dep. Commer.

681 Padmaja, S., Neta, P., Huie, R.E., 1993. Rate constants for some reactions of inorganic
682 radicals with inorganic ions. Temperature and solvent dependence. *Int. J. Chem. Kinet.*
683 25, 445–455. <https://doi.org/https://doi.org/10.1002/kin.550250604>

684 Remucal, C.K., 2014. The role of indirect photochemical degradation in the environmental
685 fate of pesticides: a review. *Environ. Sci. Process. Impacts* 16, 628–653.
686 <https://doi.org/10.1039/c3em00549f>

687 Ricci, M.C., Cross, R.F., 1993. Capillary electrophoresis separation of sulphonamides and
688 dihydrofolate reductase inhibitors. *J. Microcolumn Sep.* 5, 207–215.
689 <https://doi.org/https://doi.org/10.1002/mcs.1220050305>

690 Vialaton, D., Richard, C., 2002. Phototransformation of aromatic pollutants in solar light:
691 Photolysis versus photosensitized reactions under natural water conditions. *Aquat. Sci.*
692 64, 207–215. <https://doi.org/10.1007/s00027-002-8068-7>

693 Vione, D., 2020. A Critical View of the Application of the APEX Software (Aqueous

694 Photochemistry of Environmentally-Occurring Xenobiotics) to Predict Photoreaction
695 Kinetics in Surface Freshwaters. *Molecules* 25, 9.
696 <https://doi.org/https://doi.org/10.3390/molecules25010009>

697 Vione, D., Fabbri, D., Minella, M., Canonica, S., 2018. Effects of the antioxidant moieties of
698 dissolved organic matter on triplet-sensitized phototransformation processes:
699 Implications for the photochemical modeling of sulfadiazine. *Water Res.* 128, 38–48.
700 <https://doi.org/10.1016/j.watres.2017.10.020>

701 Vione, D., Minella, M., Maurino, V., Minero, C., 2014. Indirect Photochemistry in Sunlit
702 Surface Waters: Photoinduced Production of Reactive Transient Species. *Chem. Eur. J.*
703 20, 10590–10606. <https://doi.org/10.1002/chem.201400413>

704 Wegelin, M., Canonica, S., Mechsner, K., Fleischmann, T., Pesaro, F., Metzler, A., 1994.
705 Solar water disinfection: Scope of the process and analysis of radiation experiments.
706 *Aqua J. Water Supply Res. Technol.* 43, 154–169.

707 Wenk, J., Canonica, S., 2012. Phenolic antioxidants inhibit the triplet-induced transformation
708 of anilines and sulfonamide antibiotics in aqueous solution. *Environ. Sci. Technol.* 46,
709 5455–5462. <https://doi.org/10.1021/es300485u>

710 Wenk, J., Von Gunten, U., Canonica, S., 2011. Effect of Dissolved Organic Matter on the
711 Transformation of Contaminants Induced by Excited Triplet States and the Hydroxyl
712 Radical. *Environ. Sci. Technol.* 45, 1334–1340. <https://doi.org/10.1021/es202028w>

713 Willett, K.L., Hites, R.A., 2000. Chemical Actinometry : Using o-Nitrobenzaldehyde to
714 Measure Light Intensity in Photochemical Experiments. *J. Chem. Educ.* 77, 900–902.
715 <https://doi.org/https://doi.org/10.1021/ed077p900>

716 Wojnárovits, L., Tóth, T., Takács, E., 2020. Rate constants of carbonate radical anion
717 reactions with molecules of environmental interest in aqueous solution : A review. *Sci.*
718 *Total Environ.* 717, 137219. <https://doi.org/10.1016/j.scitotenv.2020.137219>

719 Yan, S., Liu, Y., Lian, L., Li, R., Ma, J., Zhou, H., Song, W., 2019. Photochemical formation
720 of carbonate radical and its reaction with dissolved organic matters. *Water Res.* 161,
721 288–296. <https://doi.org/10.1016/j.watres.2019.06.002>

722 Yan, S., Song, W., 2014. Photo-transformation of pharmaceutically active compounds in the
723 aqueous environment: a review. *Environ. Sci. Process. Impacts* 16, 697–720.
724 <https://doi.org/10.1039/C3EM00502J>